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**Willows for environmental projects: A literature review
of results on evapotranspiration rate and its driving
factors across the genus *Salix***

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4 **Abstract**
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7 2 Willows are increasingly used for a wide range of environmental projects, including
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9 3 biomass production, leachate treatment, riparian buffers and treatment wetlands.
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11 4 Evapotranspiration (ET), assumed to be high for most willow species used in
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13 5 environmental projects, affects hydrological cycles and is of key interest for project
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15 6 managers working with willows. Here, we present a comprehensive review of ET rates
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17 7 provided in the literature for the genus *Salix*. We aim to summarize current knowledge of
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19 8 willow ET and analyze its variability depending on context. We compiled and analyzed
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21 9 data from 57 studies, covering 16 countries, 19 willow species and dozens of cultivars.
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24 10 We found a mean reported ET rate of 4.6 ± 4.2 mm/d, with minimum and maximum
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26 11 values of 0.7 and 22.7 mm/d respectively. Although results reported here varied
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28 12 significantly between some species, overall interspecific standard deviation (± 3.6 mm/d)
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30 13 was similar to intraspecific variation (± 3.3 mm/d) calculated for *S. viminalis*, suggesting
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32 14 a greater influence of the growing context on ET than species identity. In terms of
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34 15 environmental and management variables, water supply, fertilization and contamination
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36 16 were identified as driving factors of ET across willow species. Effects of root age,
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38 17 experimental context, planting density and soil type were more nuanced. Our findings
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40 18 provide synthetic data regarding willow ET. We encourage practitioners who use ET data
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42 19 from the literature to be aware of the main drivers of ET and to consider the influence of
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44 20 the experimental aspects of a study in order to interpret data accurately and improve
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Keywords: evapotranspiration variability, water use, irrigation planning, wetland design, water loss, willow coppicing

1. Introduction

Willows (genus *Salix*) are comprised of hundreds of species, distributed throughout the world, but mostly in the northern hemisphere (Argus, 1986). They can take various growth forms, from small shrubs to large trees. Although some species are adapted to harsh or arid conditions, they more often colonize humid or wet habitats (Dickmann and Kuzovkina, 2014). Aside from traditional pharmaceutical and artisanal uses, willows also have many environmental and energy applications. For some uses, they are produced in short rotation coppice plantations (Zsuffa *et al.*, 1984; Gullberg, 1993; Volk *et al.*, 2006; Guidi *et al.* 2013), sometimes irrigated with wastewater (Lachapelle-T. *et al.*, 2019), sewage sludge (Dimitriou and Rosenqvist, 2011) or leachate (Duggan, 2005). They are thus suitable for use in prevention of leaching of hazardous wastes in evapotranspirative plantations (ET covers; R  th *et al.*, 2007; Mirck and Volk, 2009), phytoremediation of contaminated soils (Witters *et al.*, 2009; Grenier *et al.* 2015), treatment wetlands (Gregersen and Brix, 2001; Curneen and Gill, 2014), and urban and agricultural catchment runoff systems (H  nault  thier *et al.* 2017) or even to prevent erosion (Yoder, 1993). Over time, *Salix* species performance has been enhanced by selection and genetic improvement programs (Lindegaard and Barker, 1997; Kopp *et al.*, 2001; Smart and Cameron, 2008), and most environmental projects involving willows have used selected or improved cultivars rather than natural species.

Along with high biomass production, willows are known for their high water consumption. Little information is available to enable comparison of willow transpiration

(T) with that of other woody species, but it is generally accepted that willow species used for biomass production and other wetland or riparian occurring species in a temperate climate transpire much more than other herbaceous crops (Persson, 1997). Although a high evapotranspiration (ET) rate is essential for some of the uses cited above, such as ET covers, it may be undesirable in other cases. In Europe, for instance, rapid expansion of willow plantations for biomass production has raised concerns about potential disturbance of natural hydrological systems (Dimitriou *et al.*, 2009). An example of such disturbance has been documented in Australia, where willow introduction is thought to have increased water shortage problems, and caused other environmental damage (Doody and Benyon, 2011); willows are now even considered an invasive and prohibited species in some parts of the world (Doody *et al.*, 2014; Marttila *et al.*, 2018; Tang *et al.*, 2018). ET is also an important factor to consider for the design and performance evaluation of treatment wetlands (Beebe *et al.*, 2014; Białowiec *et al.*, 2014), which are sometimes planted with willows. ET rate thus represents an essential design and operational tool for practitioners working with willows, as well as an important factor to consider before extensive introduction of willows in a given area.

ET measurement is complex and requires substantial time, as well as human, technical and financial resources (Allen *et al.*, 2011). In most cases, it is far more practical to use values provided by the scientific literature to plan a project involving willows. However, ET rate is highly context-specific, meaning that results obtained in a given set of conditions might not be relevant to practitioners working in a different environment. Indeed, ET is driven by meteorological conditions, plant related factors and environmental parameters (Allen *et al.*, 1998), all of which can vary greatly from one

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4 69 site/study to another. Meteorological factors can be partially controlled when plants are
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7 70 grown in greenhouses, but are otherwise mainly governed by geographic location. For
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9 71 environmental projects, willows tend to be treated as a single species, but the numerous
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11 72 cultivars derived from many individual species and their respective morphology and
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13 73 physiology are obviously important plant factors that can influence ET variation across
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15 74 the *Salix* genus. Some environmental conditions can be at least partially controlled, such
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17 75 as irrigation, fertilization and coppicing cycle. These factors are most likely to vary
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19 76 depending on the purpose of the study and management decisions, and thus represent a
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21 77 wide range of possible growing conditions. Although not related to the ET process itself,
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23 78 the method used for measurement or estimation of ET is also known to greatly influence
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25 79 results, as most methodological approaches require a high level of expertise and rigor to
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27 80 provide reliable results (see Allen *et al.*, 2011, for a detailed review on that matter).
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29 81 Presentation of methodology and results is also highly heterogeneous, which makes
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31 82 comparing studies difficult. In the end, it can prove rather challenging to find suitable ET
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33 83 information regarding a willow cultivar for a given environmental purpose.
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35 84 The first objective of this paper was to gather the available ET rate data published for
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37 85 willow species and synthesize this information in a standardized and comparable way.
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39 86 The second objective was to assess the variation of ET across the genus and identify the
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41 87 main drivers of this variability. This review aims to improve our global knowledge of ET
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43 88 potential in rapid growing woody species like willows, and point out opportunities for
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45 89 further research on this topic. Finally, this review should serve as guide for practitioners
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47 90 working with willows for environmental projects to improve irrigation planning,
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49 91 treatment wetland sizing and other decision-making that requires willow ET information.
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93 2. Methods

94 2.1 Literature review

95 Evapotranspiration is, in fact, the combination of both plant T and soil evaporation (E_s).
96 Willows are woody plants that are often fast growing, and thus develop a considerable
97 leaf area. According to Shuttle and Wallace's energy partitioning model (1985), high leaf
98 area index (LAI) implies a reduced E_s proportion in ET. This is illustrated in numerous
99 studies presented in this review, as we see the E_s to ET ratio decline in the growing
100 season as the willow leaf cover becomes established (Grip et al., 1989; Iritz et al., 2001;
101 Lindroth et al., 1994; Persson, 1997). For the purpose of this review, T results have been
102 considered along with ET results, under the premise that willow T is a fair estimate of
103 total ET. We are, however, aware that T might represent an under-estimation of the true
104 ET value.

105 2.1.1 Articles selection

106 A literature review was performed using the keywords "*willow* OR *Salix*" AND
107 "*evapotranspiration* OR *transpiration* OR *water use*", in the Web of Science, Scopus and
108 Google Scholar databases. We selected peer-reviewed articles presenting original results
109 of ET (or T) rates, or data allowing easy calculation of ET rate (*e.g.* irrigation and
110 drainage volumes). We excluded studies presenting data related to ET but not detailed
111 enough to calculate a daily rate (*e.g.* instantaneous rate of T, water-use efficiency), ET
112 results from plant communities including other species than willows and studies
113 measuring willow T at laboratory or growth chamber scale. For instance, for an ET rate
114 provided as an amount of water transpired by a leaf area per unit of time, the leaf area

index as well as the typical daily transpiration period (*e.g.* hours of sunlight per day) would have been necessary to convert the results to a mm/d unit. For studies presenting only stemflow results, scaling-up calculations based on sap wood area and various mathematical equations would have been necessary to convert stemflow into transpiration results. ET rates had to be convertible to mm/d units (see section 2.2), and obtained under experimental conditions that could be described by at least 3 of 8 experimental variables selected for results analysis and interpretation, as detailed in section 2.3 (willow species, age of plantation/root system, experimental conditions, water supply, planting density, dominant soil type, fertilization and contamination).

2.1.2 ET data transformation

As expected, the ET rates gathered from the literature review varied in absolute value, but also in unit of expression. For comparison purposes, we converted each result to a millimeter per day basis (mm/d), the most common unit for ET rate. For studies that presented total ET values for a given period, we divided these values by the number of days of the experiment. As some authors reported ET rates only graphically, some results were extracted from these graphs. For studies that reported ET rates in terms of volume per plant, the conversion in mm/d was calculated based on the soil area of the plant container (*e.g.* lysimeter surface area) or soil area covered by the plant (inferred from canopy area or planting density).

2.2 Comparative analysis based on experimental variables

To interpret the variability of ET rates across studies testing various factors, we used an approach based on a semi-quantitative classification of the experimental and environmental conditions under which the studies were performed. These "conditions",

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4 138 also referred to as "variables" or "factors", include both independent variables and
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6 139 conditions imposed by the authors. We decided to exclude typical meteorological and
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9 140 climatic ET limiting factors such as temperature, solar radiation, wind and water vapor
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11 141 pressure deficit (VPD) of our analysis, since the effect of those factors on potential ET
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13 142 (pET) are already well described in scientific literature related to ET and should mainly
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15 143 be driven by geographic location. We then considered plant related variables and
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17 144 environmental and management variables; each variable was divided into several
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19 145 qualitative or semi-quantitative levels (Table 1).
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23 146 2.2.1. Plant variables

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26 147 Different plant species have a different T rate according to their intrinsic
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28 148 ecophysiological properties and environment (Bohnert *et al.*, 1995). Including the plant
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30 149 species in a variance analysis would potentially reveal a difference in ET rate between
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32 150 species of the willow genus. T rate should also vary for a given species according to plant
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34 151 growing conditions. To estimate if differences between species were more likely due to
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36 152 taxonomical differences or to growing conditions, we evaluated inter and intraspecific ET
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38 153 rate variation (α_{inter} and α_{intra} respectively). An interspecific variation greater than
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40 154 intraspecific variation would suggest an influence of the species itself on ET rate. ET rate
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42 155 is closely linked to growth rate, which itself is thought to decrease with age (Willebrand
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44 156 and Verwijst, 1993). Consequently, we also considered the age of the plantation as a
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46 157 potential explanatory factor for ET variation. We divided this variable into 3 categories:
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48 158 the establishment year (*first year*), for willows grown from cuttings that have to develop
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50 159 their root system, *young* and *mature* willows (Table 1). Willows with a root system of 5
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years of age or more were considered as *mature* because we supposed that, at this point, the root system should be well established.

2.2.2. *Environmental and management variables*

In every study, willows are grown under various conditions determined by the experimenter (management variables) or naturally present on the study site (environmental variables). Some variables like planting density or soil type can be either managed or naturally determined depending on the experimental context. Other factors like water supply can be both determined and random, when plants are provided with rainfall and controlled irrigation at the same time, for instance. Fertilization and contamination are normally deliberately provided to the plants.

The *experimental context* variable was chosen to represent the spatial scale of the willow stand, the *plantation* level being the largest scale and the *mesocosm* the smallest. The levels of this variable also indicate if the experimental unit is an open (*floodplain* and *plantation*) or closed (*treatment wetland* and *mesocosm*) system in terms of hydrological and soil processes.

Water supply is typically considered a limiting factor for ET (Payero *et al.*, 2008; Novák, 2012). Not all references provided sufficient methodological information to calculate the actual volume of water provided to the plants. Thus, we classified this variable with semi-quantitative levels (Table 1) according to the global volume of water available or provided to the plants. When water supplies were quantified, we calculated the mean daily volume provided to plants and classified it as follows: < 5 mm/d was considered *low*, 5 to 10 mm/d *medium* and > 10 mm/d *high*. When insufficient quantitative information was provided, water supply was considered *low* when the only water input

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4 183 was rain (in semi-arid to arid climate) or when water stress was imposed or reported by
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6 184 the authors; *medium* when input was rain in humid to very humid climate, when a small
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9 185 amount of artificial irrigation was added to rainfall or when the water table was
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11 186 controlled to a high but non-saturating level; and *high* when high levels of irrigation were
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14 187 provided or when the water level saturated the media (*e.g.* in a treatment wetland or a
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16 188 floodplain).

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19 189 Planting density can affect willows negatively, by increasing competition between
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21 190 individuals for soil resources, or positively, by maximizing light interception (Willebrand
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23 191 and Verwijst, 1993). We categorized a density of 1 plant per m² or less as *low*. The
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25 192 *medium* level included a density from 1 to 4, based on common values used for willow
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27 193 plantation (Willebrand *et al.*, 1993; Volk *et al.*, 2006, Walle *et al.*, 2007). A density
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29 194 higher than 4 plants per m² was considered *high*.

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33 195 We also selected soil type as a variable because of its influence on soil water potential
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35 196 and water availability (Novák, 2012). The relation between water and soil depends on the
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37 197 type of soil particles and can act on two levels. The first level, which is referred to in
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39 198 agriculture as field capacity, determine the soil water content after gravitational drainage
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41 199 has occurred. The more sand is contained in the soil, the less water will remain in the soil
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43 200 at field capacity because of the low attraction between sand particles and water
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45 201 molecules, while an increase in clay proportion, and furthermore in organic content,
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47 202 increases soil water retention capacity (Waller and Yitayew, 2015). However on a second
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49 203 level, at the same water content, water will be more easily available to plants in a sandy
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51 204 soil, where water potential is higher (due lower water molecules attraction) than in a
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53 205 clayey or organic soil water that have lower water potential due to the matrix attraction
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(Waller and Yitayew, 2015). Because the substrates used in the studies reviewed were never composed of one type of particles alone, we classified this variable according to the dominant type of particles in the media (Table 1). We also treated gravel media separately and excluded articles with a very specific soil type (to avoid having a level of the category with only one observation) or that did not provide information on the media. The effect of fertilization and contamination were treated for their direct effect on plant T (Feldhake *et al.*, 1983; Trapp *et al.*, 2000). They were treated as a binomial variable (presence or absence; Table 1) because of the disparities between the type of nutrient sources and contaminants and their method of addition. Landfill leachate was a particular case, and was considered here as both a source of nutrients and contamination. Indeed, willow can use ammonia (typically present in leachate) as a source of a nutrient which can become a toxicant when its concentration is too high. Other leachate constituents such as chlorinated compounds can have a similar toxic effect.

2.3 Statistical analysis

When a study tested more than one level of at least one variable, it was considered to have more than one result (n) in the variance analysis. For example, a study measuring ET of two species with two different fertilization levels accounted for four individual results (n=4) in the analysis. When results were reported for many replicates of the same treatment, only the mean value was considered. Using this approach, we built a data base by associating each individual ET rate result to the appropriate level of each variable from Table 1. We then proceeded to the comparative analysis, which consisted of a variance analysis (ANOVA) using R statistical software (version 3.5.1). The model tested in the analysis included all variables, in order to consider their simultaneous effect on ET

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4 229 rate. The ET results followed a Fisher distribution, and a log transformation was used to
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6 230 normalize the data prior to statistical analysis. Missing information for some variables
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9 231 (no observation for one or more variables for a given ET result) yielded an unbalanced
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11 232 statistical plan. However, the most commonly used type of ANOVA (type I) has the
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13 233 effect of giving significantly different results depending on how the variables are ordered
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15 234 in the model when provided with an unbalanced data set. Therefore, we decided to
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17 235 perform a type II ANOVA, which typically gives higher P values (less significant results)
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19 236 but is not influenced by the order of the variables in the model. Type II ANOVAs are
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21 237 generally suggested as the best substitute for a type I analysis for unbalanced data
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23 238 (Langsrud, 2003). We also used a correlogram to illustrate possible interactions between
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25 239 the variables of the comparative analysis, except for the variable *plant species*, which is
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27 240 composed of more than fifteen levels. Following the comparative analysis, we also
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29 241 performed linear regression analysis between ET results and both planting density
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31 242 (plants/m^2) and water input (mm/d) for the articles where quantitative information was
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33 243 provided for those two variables. For all analyses, a P value lower than 0.05 was
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35 244 considered significant. Finally, α_{intra} was calculated as the standard deviation of the
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37 245 results associated with the most frequently studied species (*S. viminalis*, $n=53$), while
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39 246 α_{inter} was calculated as the standard deviation between the average ET rate reported for
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41 247 each specie ($n=18$).
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53 249 **3. Results**

54 250 ***3.1 Article selection and data transformation***

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4 251 Out of the 800+ articles analyzed, 57 met our selection criteria. The studies covered the
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6 252 period from 1986 to 2019 and were from 16 countries, although half (27) originated from
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9 253 Northern Europe. Results were obtained for natural willow species (21 articles) and
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11 254 cultivars (36 articles), each articles testing one to four species and up to 6 different
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14 255 cultivars, for a total of 19 species studied (Table 2). Plants growing conditions ranged
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16 256 from wild to cultivated/controlled, stressed to non-stressed. Overall, 20 studies reported
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19 257 results in mm/d, 26 studies were in mm for a given period (most of the time, per season),
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21 258 and the remaining 9 studies required additional calculations to express results in mm/d.
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24 259 Sixteen articles presented plant T results only.
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26 260 At least 4 of the 8 variables considered for categorization of the results were provided in
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28 261 each article (Table 2). Information regarding planting density was missing in 6 articles,
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31 262 and root system age in six other articles, while both types of information were missing in
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33 263 13 studies. However, this information was mainly missing from studies conducted on
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36 264 natural willow stands, where age and density are heterogeneous and more difficult to
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38 265 document. The soil type turned out to be very difficult to categorize due to the wide range
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41 266 of substrates used and the ambiguous nature of the dividing line between clayish and
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43 267 sandy soil (*e.g.* a soil with 50% sand particles and 40% clay particles was considered as
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45 268 *sand* even if it varies greatly from pure sand). After extracting information from all the
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48 269 studies according to the different levels of the categorical variables (see Section 2.2 and
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51 270 Table 1), 110 ET rate results could be treated individually ($n = 110$, Table 2). Thirty-five
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53 271 articles presented results obtained with homogenous experimental variables (1 study =
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55 272 1 result), and the studies that tested the most factors resulted in nine individual results
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58 273 (Table 2; Martin and Stephens, 2006). Some studies tested different treatments but were
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4 274 still considered as one result in our analysis because variation between the treatments
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6 275 could not be captured with our variable categorization (*e.g.* 3 irrigation rates tested, but
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9 276 all below 5mm/d, which is considered *low* for the variable *water supply*)
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11 277 **3.2 Comparative analysis**

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14 278 According to the 110 observations, ET rates ranged from 0.7 up to more than 20 mm/d.
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16 279 The lowest rate was reported for T (rather than ET), expressed on an annual basis, of *S.*
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18 280 *fragilis* grown in a gravelly/sandy soil on the banks of a stream (Marttila *et al.*, 2017),
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21 281 while the highest average rate of 22.7 mm/d measured over one growing season by water
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23 282 balance for the species *S. miyabeana* ‘SX67’ with a mature root system and grown in a
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25 283 treatment wetland with high water supply, medium planting density, organic soil and low
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27 284 contamination and fertilization (Frédette *et al.*, 2019). Mean reported ET rate across all
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29 285 studies was 4.6 mm/d (\pm 4.5), with about 80% of reported ET rates ranging from 0 to 10
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31 286 mm/d. We observed some trends regarding factors interactions (Figure 1). For example,
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33 287 we observe that willows growing in *floodplain* are almost systematically associated with
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35 288 *mature* trees, *medium* to *high* water supply, *high* planting density and natural conditions
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37 289 (no fertilization or contamination), that *first year* cuttings and *young* willows are mainly
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39 290 used in *mesocosms* studies while most *mature* trees studied are in *plantation*, or that
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41 291 fertilization was more frequently associated with *treatment wetlands* and *mesocosms*
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43 292 rather than *floodplains* or *plantations*.
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50 293 **3.2.1 Plant variables**

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53 294 While 30 and 40 results were reported for *first year* and *young* willows respectively, only
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55 295 13 pertained to willows with a *mature* root system (Figure 2). The age of the root system
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58 296 did not significantly affect the results, even though fresh stems newly developed from
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cuttings tended to be associated with slightly lower ET than *young* or *mature* willow plants (4.2 mm/d compared to 5.3 and 5.0 mm/d respectively; Figure 2). Sixteen of the 19 species were associated to 5 results or less, compared to the most studied species, *S. viminalis*, which was associated to 53 results. Three articles did not provide the exact taxonomic identity of the willow studied (*Salix* sp.). There was a significant difference of the results according to species (Figure 2). However, α_{intra} for *S. viminalis* (3.3 mm/d) was very similar to variation between species mean ET rate ($\alpha_{\text{inter}} = 3.2$ mm/d). *Salix amygdalina*, *S. exigua* and *S. psammophila* were the three species with the lowest mean ET rate (< 2 mm/d), while *S. babylonica*, *S. cinerea*, *S. goodgingii*, *S. miyabeana* and *S. nigra* (all cultivars combined) had the highest (> 7 mm/d; Figure 2).

3.2.2 Environmental and management variables

The majority of the articles reviewed studied willows growing either in mesocosms or in plantations (Figure 3). The effect of experimental context on ET rates was not significant (Figure 3). Nonetheless, *treatment wetlands* were generally associated with higher results (7.9 mm/d on average), followed by *mesocosms* (5.7 mm/d), *floodplain* (3.6 mm/d) and finally *plantation* results (2.9 mm/d; Figure 3). Water supply was found to be a significant experimental variable (Figure 3), with *low* water supplies associated to the lower results (2.4 mm/d on average), compared to *medium* and *high* water supply (5.0 and 7.0 mm/d, respectively; Figure 3). Almost half of the results were measured or calculated for willows that were poorly supplied with water (n=47; Figure 3). Furthermore, we found a significant linear correlation between daily water input and daily ET rate for open systems ($r^2 = 0.7$, Figure 4). The planting density did not significantly explain ET rate variations in our factorial analysis (Figure 3). However,

average ET rates were the same for *medium* and *high* planting density (5.4 mm/d), but slightly lower at *low* density (3.2 mm/d; Figure 3). Linear regression of ET rate over planting density did not show a clear trend either (Figure 5), but the few results reported at very high planting density suggest the existence of a threshold, after which ET is limited (here estimated to be approximately 5 plants/m²; Figure 5). Regarding the type of soil in which willows were grown, most results were reported for sandy soils, followed by clayey soils. No significant effect of soil type was found (Figure 3), but the following average ET rate gradient could be observed: in organic soil (6.1 mm/d) > in clayey soil (5.3 mm/d) > in sandy soil (4.9 mm/d) > in gravel (1.6 mm/d). We should mention that only 3 results were reported for gravel substrate. Finally, fertilization and contamination both had a significant effect in the comparative analysis (Figure 3). Studies that used some kind of fertilization treatment reported ET rates 40% higher on average compared to unfertilized willows (6.1 mm/d vs. 3.5 mm/d). On the contrary, ET rates were generally lower in the presence of contaminants, although average rates were very similar (4.6 mm/d in the presence of contamination compared to 4.7 mm/d in non-contaminated conditions; Figure 3).

4. Discussion

Our review shows that mean ET rates in willows are generally below 10 mm/d, but may rise well over that value, reaching up to 23 mm/d. According to a factorial analysis performed on 110 ET rate results from 57 articles, we found that water supply, fertilization and contamination significantly affected ET rates. We identified a strong correlation between daily water input and ET rate in open systems. The effects of plant

age, experimental context, and planting density were not statistically significant, although some trends could be observed. Soil type in fact was less important than the other variables, when their simultaneous effect on ET was tested. Willow species seemed to significantly affect ET rates, but α_{inter} and α_{intra} variation of ET were equivalent. Variation of T rate between species is to be expected, because its regulation mechanisms are not the same for every taxa (Sperry, 2000). These mechanisms are generally adapted to the plant environment (Bohnert *et al.*, 1995), a good example being xerophytic species, which display various ways of preventing water loss through T (Fahn and Cutler, 1992). This could explain why *S. psammophila*, a willow species adapted to dry environments (Xiao *et al.*, 2005), had one of the lowest ET rates, while *S. nigra*, a water dependent species (Pezeshki *et al.*, 2007), had the highest. Overall, different willow species had different ET rate ranges, but in the end there were so few studies on each species and so many other factors that varied between studies that we cannot conclude that taxonomical identity dictates mean ET rate in the willow genus. Furthermore, the fact that ET variation between willows of the same species (*S. viminalis*) was the same as that between different species suggests that species identity is not the most important factor in ET variation across the willow genus, particularly for species adapted to similar environments (*e.g.* wet habitat). However, willow cultivars developed in breeding programs can promote high T rates for environmental applications like phytoremediation (Smart *et al.*, 2005) or promote increased water use efficiency (WUE) and tolerance to water limitation for biomass production (Karp *et al.*, 2011). This could explain the high variability of ET in the *S. viminalis* species, which in this review is comprised of more than 20 genetically different cultivars.

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4 366 Regarding the age of the willow root system, our hypothesis was that plants in their *first*
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6 367 *year* – the establishment year, as well as *mature* shrubs, which should have a lower
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9 368 growth rate, would be associated with lower ET rates compared to young, fast growing
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11 369 plants. Indeed, we observed lower ET for plants newly developed from cuttings, but not
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14 370 for *mature* shrubs. However, it appears that the mean average ET rate for mature trees
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16 371 was driven up mainly by the results of one study (Frédette *et al.* 2019); when those
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18 372 results are set aside, mean ET rate for mature trees drops from 5.9 mm/d to 2.4 mm/d.
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21 373 This difference could be explained by the fact that ET results in Frédette *et al.* (2019)
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23 374 were obtained from a treatment wetland with a high water supply, while all the other
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25 375 results from mature shrubs came from plantations with a low water supply. Furthermore,
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27 376 willows in the Frédette *et al.* study were recently coppiced, while most of the other
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29 377 studies were conducted on willows with much older stems. Coppicing of willows is
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31 378 known to help keep the plants in a juvenile, and thus more productive, state and it could
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33 379 then be responsible of those high ET rates. A decrease in biomass production with time
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35 380 has been documented for willows in the past, even in a coppicing system (Willebrand *et*
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37 381 *al.*, 1993), but our analysis did not allow us to demonstrate this pattern. Further studies
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39 382 should be conducted on this specific issue to provide clearer answers.
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46 383 Our findings suggest that ET rate is greater in closed and relatively small-scale systems
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48 384 (treatment wetlands and mesocosms) than in open and full-size systems (floodplain and
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50 385 plantations). In open systems, ET is higher in floodplains, where the water table (and thus
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52 386 water availability) is generally high and some flooded conditions can even occur, than in
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54 387 plantations, where water may be limited and will drain to lower soil horizons. In
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56 388 comparison, in closed systems like treatment wetlands or some mesocosms, water supply
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4 389 is often equal to or greater than plants' water demand, meaning that water is not a
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6 390 limiting factor and ET occurs at a rate closer to maximal pET. Furthermore, pET can be
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9 391 exceeded in small scale willow stands by processes like an "oasis" or "clothesline" effect
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11 392 (Allen *et al.*, 1998; Frédette *et al.*, 2019; Dotro *et al.*, 2017). An oasis effect is the result
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14 393 of a difference in temperature between willows and their surroundings, due to the cooling
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16 394 effect of ET, which increases available energy to willows by a heat advection effect (Hao
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19 395 *et al.*, 2016; Dotro *et al.*, 2017). The clothesline effect increases ET on the edges of the
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21 396 willow stand because of enhanced wind influence, as a result of the height difference
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24 397 between willows and the surrounding vegetation (Brix and Arias, 2011; Dotro *et al.*,
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26 398 2017). Both those effects could partially explain higher ET rates reported in mesocosms
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29 399 and treatment wetlands. Another aspect of the experimental context variable is that it
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31 400 shared many associations with other variable levels (Figure 1). Thus, mesocosms were
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33 401 mainly associated with younger willows and medium to high planting density; treatment
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36 402 wetlands generally had a high water supply, medium to low planting density and organic
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38 403 soil; floodplains had a medium to high water supply, high planting density, sandy or
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41 404 clayish soil, unfertilized and uncontaminated environment; and finally, plantations were
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43 405 associated with low to medium water supply, medium planting density, various soil
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46 406 types, but mainly uncontaminated conditions. When considered as the only explanatory
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48 407 variable, experimental context significantly explains ET variation ($p < 0.001$). On the one
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51 408 hand, the experimental context might provide a global indicator of ET rate combining
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53 409 many environmental and management variables, but on the other hand, it might be
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55 410 interesting to replace it by finer variables (*e.g.* experimental unit area and permeability)
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58 411 to add precision to a global analysis.
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4 412 Of all the chosen variables, water supply was one of the most significant driving factors
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6 413 of ET rate variation. Along with meteorological conditions, water is a direct limiting
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8 414 factor for ET, and the impact of water stress on ET rates is generally well described in the
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10 415 ET literature (Sperry, 2000; Bohnert *et al.*, 1995). This review highlights a strong
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12 416 correlation between water supply and ET rate across the willow genus. For open systems
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14 417 where water supplies could be quantified, this factor alone could explain most of the ET
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16 418 rate variation. However, according to the same correlation analysis, the difference
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18 419 between water supply and ET rate increased with increasing water supply, illustrating
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20 420 that the less water is limiting, the more other factors become limiting. This relation may
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22 421 not hold in a closed system, as a lesser effect of water availability on ET has been
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24 422 demonstrated in closed versus open systems (Rana and Katerji, 2000). For example,
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26 423 Guidi and Labrecque (2010) found no increase in ET rate for *S. viminalis* ‘5027’ with
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28 424 very high irrigation rates, compared to “normal” irrigation, in a pot experiment. As
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30 425 previously discussed, water use strategy may also vary from one species to another,
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32 426 depending on its natural environment but also on its breeding strategy. Most of the
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34 427 species studied here are naturally associated with humid habitats, and therefore do not
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36 428 require a very efficient water regulation mechanism, which has given willows their
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38 429 “water-wasting” plant reputation.

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40 430 Generally, increasing planting density of a crop will also increase biomass yield, until an
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42 431 optimal threshold density is reached; beyond that threshold, a higher density will not
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44 432 produce more biomass due to competition for resources such as for water or light (Assefa
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46 433 *et al.*, 2018; Ngouajio, 2001; Willebrand and Verwijst, 1993). As willow biomass is
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48 434 thought to be closely linked to ET (Martin and Stephens, 2006; Marmioli *et al.*, 2012;
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4 435 Białowiec *et al.*, 2007), the same threshold hypothesis could apply to ET rate. Our results
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6 436 strongly suggest that the planting density at which willow ET is maximal is higher than 1
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9 437 plant/m² studies using this density systematically reported lower ET rates. No significant
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11 438 differences were found between *medium* and *high* planting density, but plotting ET rates
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14 439 with the corresponding density suggests a threshold around 5 trees/m². However, only 12
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16 440 of the 57 articles reviewed reported results for densities higher than this potential
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19 441 threshold. Furthermore, yield increases for willow have been documented at a density as
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21 442 high as 11 plants/m² (Bullard *et al.*, 2002).

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24 443 In addition to water supply, water availability (often expressed as soil water potential)
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26 444 can affect ET, and the type of soil impacts water potential for a given water supply
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29 445 (Rawls *et al.*, 1982). However, the soil effect, through attraction force between soil
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32 446 particles and water, can act on two levels, as described in section 2.3.2 of the present
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34 447 manuscript. This dual effect may explain why we did not observe significantly different
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36 448 ET rates according to soil type in this review. Presence of organic matter in the soil even
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39 449 adds another level of interaction by providing additional nutrients to plants, which can
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42 450 increase growth and, consequently, ET rate, which is supported by the slightly higher ET
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44 451 rates reported here for *organic* soils. For the three studies in which *gravel* was used as a
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46 452 substrate, a high ET rate would have been expected, because the substrate was constantly
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49 453 kept saturated with water that should be highly available because of gravel's physical
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51 454 properties. However, low ET rates were measured, probably due to late season
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54 455 measurements in one case (Jing *et al.* 2010), water contamination in another (Białowiec
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56 456 *et al.*, 2003) and ET rates reported on an annual basis (including low ET rates in winter)
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59 457 in the last (Marttila *et al.*, 2017). This and the previous explanations highlight the
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4 458 simultaneous effect of multiple factors and suggest that soil type alone is not a strong
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6 459 explanatory variable for ET variation.
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10 460 As expected, fertilization increased willow ET, probably by increasing growth rate. Only
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12 461 one study used fertilization as the main treatment variation, and it reported a 96%
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14 462 increase in ET due to fertilization (Guidi *et al.*, 2008). Pistocchi *et al.* (2009) also
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16 463 reported a 51% increase of willow ET when switching from low to high fertilization. For
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18 464 some studies, the variation in the fertilization treatment was due to amendments to the
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20 465 substrate in various forms, such as compost, mechanical-biological pretreated waste
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22 466 material, sewage sludge or other forms of organic matter addition (Rüth *et al.*, 2007;
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24 467 Białowiec *et al.*, 2007; Martin and Stephens, 2006). Despite the presence of other
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26 468 interacting factors, the *fertilized* treatment in these studies was always associated with
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28 469 slightly higher ET rates. Interestingly, most of the articles that were associated with
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30 470 fertilization were, in fact, exposing willows to various types of wastewater, mainly
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32 471 landfill leachate or from domestic and agricultural source. These types of water did
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34 472 contain nutrients such as nitrogen and phosphorus, but also contained harmful
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36 473 compounds such as chloride and sulfate, high ammonium and salt concentrations, and
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38 474 metalloids, particularly when leachates were the source of fertilization. A good
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40 475 illustration of the dual effect of this type of effluent is provided by Białowiec *et al.*
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42 476 (2003), describing how a low concentration of landfill leachate had a positive effect on
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44 477 willow ET but increasing concentrations became deleterious to the plants. Conversely,
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46 478 Curneen and Gill (2014) reported an increase in ET when using primary (more
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48 479 concentrated) instead of secondary (less concentrated) effluent from domestic
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50 480 wastewater, probably because the beneficial effect of the high levels of nitrogen and
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4 481 phosphorus in this type of wastewater exceeded other potentially negative water
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6 482 characteristics. This may also explain why average ET rate was similar for contaminated
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9 483 and uncontaminated results; 9 of the 14 studies that measured ET rates in contaminated
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11 484 conditions provided fertilized conditions at the same time. When testing chloride
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14 485 contamination only, Stephens (2000) clearly demonstrated the negative impact of
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16 486 increasing chloride concentration on ET. Furthermore, ET rate is frequently used as a
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19 487 toxicity indicator in lab tests, due to its sensitivity to increasing pollutant concentration
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21 488 (Trapp *et al.*, 2000, Clausen *et al.*, 2018). Therefore, contamination and fertilization
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24 489 should be considered together to accurately judge their influence on ET in view of their
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26 490 compensatory effect on each other.

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29 491 ET is a complex process, and despite the numerous factors evaluated here, there are
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32 492 additional variables that were not analyzed numerically but that could provide a better
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34 493 understanding of ET results. As previously mentioned, biogeographical variation along
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37 494 with meteorological conditions are important factors, and a synthetic and theoretical
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39 495 explanation of those variables can be found in ET literature (see for example Holdridge,
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41 496 1947 and Allen *et al.*, 1998). For example, higher temperatures and smaller seasonal
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44 497 variations correlate with high ET rates reported in regions as such as Arizona (Nagler *et*
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46 498 *al.*, 2003) and Louisiana (Conger and Portier, 2001). In this review, we also found that
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49 499 some results reflected coupling and decoupling of willow T with atmosphere and its
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51 500 associated water vapor pressure deficit, which is variable along with plant development
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54 501 (Mirck and Volk, 2009). Otherwise, ET rates show obvious seasonal variation that is
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56 502 accentuated in northern countries, which have shorter growing periods and little to no ET
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59 503 during winter. ET also varies according to phenology and leaf development during the

growing period. Although this concept might seem obvious, we consider it pertinent for practitioners planning a project based only on published ET values. According to most of the articles reviewed here, maximum leaf area of willows is generally reached in late summer months, and ET rate is maximal from July to September in the northern hemisphere. This phenological pattern is quite different from that in typical grass species, which develop their total aerial biomass earlier in the season (Persson, 1997). Therefore, the willow crop coefficient (K_c ; *i.e.* ratio between willow ET and a reference well-watered grass surface ET) has proven to be very high late in the season (Curneen and Gill, 2016; Persson, 1995; Irmak *et al.*, 2013; Guidi *et al.*, 2008). The crop coefficient is a thus a very useful tool for irrigation planning or project design, and being aware of the temporal variation of willow K_c is an asset.

Finally, although the methodological approach adopted by researchers to measure ET has no direct influence on ET processes, it can contribute to greater ET measurements and calculations. Allen *et al.* (2011) suggested an error range from 5 to 200% in ET measurement, depending on the method used, experimenter experience and training, as well as equipment reliability. Water balance, when performed in a closed system where water fluxes are controlled (*e.g.* lysimeter, treatment wetlands) should yield the most reliable results; this type of method was the most commonly used among the articles reviewed here. When used alone, open water balance can be imprecise due to a high degree of uncertainty regarding leakage and runoff processes. Sap flow approaches are a subset of methods that estimate plant T based on water transport in stems. The method itself presents a number of potential sources of error (Allen *et al.*, 2011), and requires extensive calculations and precautions to scale up the ET values from stems to a whole

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4 527 tree stand (Green *et al.*, 2003; Grime and Sinclair, 1999). It can therefore be considered a
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6 528 difficult method that requires great expertise and experimental rigor (Allen *et al.*, 2011).
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9 529 Still, the general homogeneity of sapwood in fast-growing willow shrubs developed for
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11 530 coppice plantations makes scaling up results for them easier and more reliable than for
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14 531 other shrubs or trees with more complex arborescence patterns. Modelling methods
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16 532 comprise several distinct approaches, including micrometeorological methods such as
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19 533 energy balance or Penman methods, and models based on different variables like leaf or
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21 534 soil parameters, or a combination of modelling approaches. In this review, we found that
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24 535 studies based on modelling approaches tended to provide low ET rates and less variation
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26 536 across studies than the two previous approaches. This could be due to the fact that most
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29 537 of these modelling studies were conducted in plantations (associated here with lower ET
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31 538 rates) or to over parameterization of models that tend to limit ET in additive or even
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34 539 multiplicative ways. Still, modelling studies are often based on field measurements and
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36 540 serve as practical and sometimes more realistic tools for irrigation planning.
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42 542 **5. Conclusions**

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44 543 Overall, willow ET rates reported in scientific literature varied mainly according to plant
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46 544 species, water supply, fertilization and contamination, although species influence remains
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49 545 unclear. It can be hypothesized that environmental/experimental factors have more
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51 546 influence on ET of willows that share similar plant life-forms (*e.g.* fast-growing shrubs
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54 547 naturally found in wet habitats) than taxonomical identity. Water supply seems to be the
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57 548 most limiting factor among those investigated here. In open systems and until pET is
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59 549 reached, there is a positive linear relation between water supply and ET rate. The
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4 550 projected use of the willows (*e.g.* ET cover, treatment wetland, biomass production)
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7 551 informs us on many aspects of the growing conditions, such as the relative water
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9 552 availability and the scale of the willow stand. This variable alone could thus be used to
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11 553 estimate whether ET should be expected to be high or low, although it does not allow
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13 554 precise estimation of ET. A planting density of two to five trees per square meter should
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15 555 be favored to maximize ET and avoid excessive competition. Based on the present
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17 556 review, the effect of soil type on ET remains unclear but may not be one of the most
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19 557 important driving factors. Fertilization and contamination levels provided to plants
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21 558 should be compared to estimate their global effect on plant growth and ET, particularly in
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23 559 cases where willows are irrigated with wastewater or leachate. Finally, biogeographic
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25 560 location will always influence potential ET rate and should be considered by project
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27 561 planners, in addition to the plants, environmental and experimental issues pointed out in
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29 562 this review. Future research on willow ET should focus on 1) specifying the root or stem
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31 563 age effect on ET, 2) confirming the optimal density for ET processes, as well as 3) testing
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33 564 whether, under a given set of growing conditions, species or cultivar identity has a
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35 565 significant effect on ET or not.
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52 570 **References**

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62
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64
65

571 Allen, R.G., Pereira, L.S., Howell, T.A., Jensen, M.E. (2011). Evapotranspiration
 572 information reporting: I. Factors governing measurement accuracy. *Agr Water*
 573 *Manage*, 98(6), 899-920. <https://doi.org/10.1016/j.agwat.2010.12.015>
 574 Allen, R.G., Pereira, L.S., Raes, D., Smith, M. (1998). Crop evapotranspiration-
 575 Guidelines for computing crop water requirements-FAO Irrigation and drainage paper
 576 56. FAO, Rome, 300(9), D05109.
 577 Argus, G.W. (1986). The genus *Salix* (Salicaceae) in the southeastern United States. *Syst*
 578 *Bot*, 1-170. DOI: 10.2307/25027618
 579 Assefa, Y., Carter, P., Hinds, M., Bhalla, G., Schon, R., Jeschke, M., Paszkiewicz, S.,
 580 Smith, S., Ciampitti, I.A. (2018). Analysis of long term study indicates both
 581 agronomic optimal plant density and increase maize yield per plant contributed to
 582 yield gain. *Sci Rep*, 8(1), 4937. <https://doi.org/10.1038/s41598-018-23362-x>
 583 Beebe, D.A., Castle, J.W., Molz, F.J., Rodgers, J.H.. (2014). Effects of
 584 evapotranspiration on treatment performance in constructed wetlands: experimental
 585 studies and modeling. *Ecol Eng* 71, 394-400.
 586 <https://doi.org/10.1016/j.ecoleng.2014.07.052>
 587 Benettin, P., Quelo, P., Bensimon, M., McDonnell, J. J., & Rinaldo, A. (2019).
 588 Velocities, Residence Times, Tracer Breakthroughs in a Vegetated Lysimeter: A
 589 Multitracer Experiment. *Water Resources Research*, 55(1), 21-33.
 590 <https://doi.org/10.1029/2018WR023894>
 591 Białowiec, A., Albuquerque, A., Randerson, P.F. (2014). The influence of
 592 evapotranspiration on vertical flow subsurface constructed wetland performance. *Ecol*
 593 *Eng*, 67, 89-94. <https://doi.org/10.1016/j.ecoleng.2014.03.032>

- 594 Białowiec, A., Wojnowska-Baryła, I., Agopsowicz, M. (2007). The efficiency of
595 evapotranspiration of landfill leachate in the soil–plant system with willow *Salix*
596 *amygdalina* L. Ecol Eng, 30(4), 356-361.
597 <https://doi.org/10.1016/j.ecoleng.2007.04.006>
- 598 Białowiec, A., Wojnowska-Baryła, I., Hasso-Agopsowicz, M. (2003). Effectiveness of
599 leachate disposal by the young willow sprouts *Salix amygdalina*. Waste Manage
600 Res, 21(6), 557-566. <https://doi.org/10.1177/0734242X0302100608>
- 601 Bohnert, H.J., Nelson, D.E., Jensen, R.G. (1995). Adaptations to environmental stresses.
602 Plant Cell, 7(7), 1099. <https://dx.doi.org/10.1105%2Ftpc.7.7.1099>
- 603 Borek, R., Faber, A., Kozyra, J. (2010). Water implications of selected energy crops
604 cultivated on a field scale. J Food Agr Environ, 8, 1345-1351.
- 605 Brix, H., Arias, C.A. (2011). Use of willows in evapotranspirative systems for onsite
606 wastewater management–theory and experiences from Denmark. Proceedings of the
607 STREPOW International Workshop, Andrevlje-Novi Sad, Serbia, 15-29.
- 608 Budny, M.L., Benscoter, B.W. (2016). Shrub encroachment increases transpiration water
609 loss from a subtropical wetland. Wetlands, 36(4), 631-638.
610 <https://doi.org/10.1007/s13157-016-0772-5>
- 611 Bullard, M.J., Mustill, S.J., McMillan, S.D., Nixon, P.M., Carver, P., Britt, C.P. (2002).
612 Yield improvements through modification of planting density and harvest frequency in
613 short rotation coppice *Salix* spp.—1. Yield response in two morphologically diverse
614 varieties. Biomass Bioenerg, 22(1), 15-25. [https://doi.org/10.1016/S0961-](https://doi.org/10.1016/S0961-9534(01)00054-X)
615 [9534\(01\)00054-X](https://doi.org/10.1016/S0961-9534(01)00054-X)

- 1
2
3
4 616 Čermák, J., Kučera, J., Nadezhdina, N. (2004). Sap flow measurements with some
5
6 617 thermodynamic methods, flow integration within trees and scaling up from sample
7
8 618 trees to entire forest stands. *Trees*, 18, 529–546. DOI 10.1007/s00468-004-0339-6
9
10
11 619 Clausen, L.P.W., Jensen, C.K., Trapp, S. (2018). Toxicity of 2, 3, 5, 6-tetrachlorophenol
12
13 620 to willow trees (*Salix viminalis*). *Hum Ecol Risk Ass*, 24(4), 941-948.
14
15 621 <https://doi.org/10.1080/10807039.2017.1403280>
16
17
18 622 Conger, R.M., Portier, R.J. (2001). Transpiration in black willow phytoremediation plots
19
20 623 as determined by the treet trunk heat balance method. *Remediation Journal: The*
21
22 624 *Journal of Environmental Cleanup Costs, Technologies & Techniques*, 11(4), 79-88.
23
24 625 <https://doi.org/10.1002/rem.1016>
25
26
27 626 Cureton, P.M., Groenevelt, P.H., McBride, R.A. (1991). Landfill leachate recirculation:
28
29 627 effects on vegetation vigor and clay surface cover infiltration. *J Environ Qual*, 20(1),
30
31 628 17-24. doi:10.2134/jeq1991.00472425002000010005x
32
33
34 629 Curneen, S.J., Gill, L.W. (2014). A comparison of the suitability of different willow
35
36 630 varieties to treat on-site wastewater effluent in an Irish climate. *J Environ*
37
38 631 *Manage*, 133, 153-161. <https://doi.org/10.1016/j.jenvman.2013.12.004>
39
40
41 632 Curneen, S., Gill, L.W. (2016). Willow-based evapotranspiration systems for on-site
42
43 633 wastewater effluent in areas of low permeability subsoils. *Ecol Eng*, 92, 199-209.
44
45 634 <https://doi.org/10.1016/j.ecoleng.2016.03.032>
46
47
48 635 Dickmann, D.I., Kuzovkina, J. (2014). Poplars and willows of the world, with emphasis
49
50 636 on silviculturally important species. *Poplars and willows: trees for society and the*
51
52 637 *environment*, 22(8).
53
54
55
56
57
58
59
60
61
62
63
64
65

- 1
2
3
4 638 Dimitriou, I., Aronsson, P. (2004). Nitrogen leaching from short-rotation willow coppice
5
6 639 after intensive irrigation with wastewater. *Biomass Bioenerg*, 26(5), 433-441.
7
8 640 <https://doi.org/10.1016/j.biombioe.2003.08.009>
9
10
11 641 Dimitriou, I., Aronsson, P. (2010). Landfill leachate treatment with willows and poplars—
12
13 642 Efficiency and plant response. *Waste manage*, 30(11), 2137-2145.
14
15 643 <https://doi.org/10.1016/j.wasman.2010.06.013>
16
17
18 644 Dimitriou, I., Aronsson, P. (2011). Wastewater and sewage sludge application to willows
19
20 645 and poplars grown in lysimeters—plant response and treatment efficiency. *Biomass*
21
22 646 *Bioenerg*, 35(1), 161-170. <https://doi.org/10.1016/j.biombioe.2010.08.019>
23
24
25 647 Dimitriou, I., Busch, G., Jacobs, S., Schmidt-Walter, P., Lamersdorf, N. (2009). A review
26
27 648 of the impacts of short rotation coppice cultivation on water issues. *Landbauforsch*
28
29 649 *Volk*, 59(3), 197-206.
30
31
32 650 Dimitriou, I., Rosenqvist, H. (2011). Sewage sludge and wastewater fertilisation of Short
33
34 651 Rotation Coppice (SRC) for increased bioenergy production—biological and
35
36 652 economic potential. *Biomass Bioenerg*, 35(2), 835-842.
37
38 653 <https://doi.org/10.1016/j.biombioe.2010.11.010>
39
40
41 654 Doody, T., Benyon, R. (2011). Quantifying water savings from willow removal in
42
43 655 Australian streams. *J Environ Manage*, 92(3), 926-935.
44
45 656 <https://doi.org/10.1016/j.jenvman.2010.10.061>
46
47
48 657 Doody, T. M., Nagler, P. L., Glenn, E. P., Moore, G. W., Morino, K., Hultine, K. R., &
49
50 658 Benyon, R. G. (2011). Potential for water salvage by removal of non- native woody
51
52 659 vegetation from dryland river systems. *Hydrological Processes*, 25(26), 4117-4131.
53
54 660 <https://doi.org/10.1002/hyp.8395>
55
56
57
58
59
60
61
62
63
64
65

- 661 Dotro, G., Langergraber, G., Molle, P., Nivala, J., Puigagut, J., Stein, O., von Sperling,
662 M. (2017). Biological Wastewater Treatment Series – Volume 7: Treatment Wetlands,
663 IWA Publishing, United-Kingdom.
- 664 Duan, L., Lv, Y., Yan, X., Liu, T., & Wang, X. (2017). Upscaling stem to community-
665 level transpiration for two sand-fixing plants: *Salix gordejvii* and *Caragana*
666 *microphylla*. *Water*, 9(5), 361. <https://doi.org/10.3390/w9050361>
- 667 Duggan, J. (2005). The potential for landfill leachate treatment using willows in the
668 UK—a critical review. *Resour Conserv Recy*, 45(2), 97-113.
669 <https://doi.org/10.1016/j.resconrec.2005.02.004>
- 670 Fahn, A., Cutler, D.F. (1992). *Xerophytes*. Gebrüder Borntraeger, Germany.
- 671 Feldhake, C.M., Danielson, R.E., Butler, J.D. (1983). Turfgrass Evapotranspiration. I.
672 Factors Influencing Rate in Urban Environments 1. *Agron J*, 75(5), 824-830.
673 [doi:10.2134/agronj1983.00021962007500050022x](https://doi.org/10.2134/agronj1983.00021962007500050022x)
- 674 Frédette, C., Grebenshchykova, Z., Comeau, Y., & Brisson, J. (2019). Evapotranspiration
675 of a willow cultivar (*Salix miyabeana* SX67) grown in a full-scale treatment wetland.
676 *Ecological Engineering*, 127, 254-262.
- 677 Glenn, E., Tanner, R., Mendez, S., Kehret, T., Moore, D., Garcia, J., Valdes, C. (1998).
678 Growth rates, salt tolerance and water use characteristics of native and invasive
679 riparian plants from the delta of the Colorado River, Mexico. *J Arid Environ*, 40(3),
680 281-294. <https://doi.org/10.1006/jare.1998.0443>
- 681 Green, S., Clothier, B., Jardine, B. (2003). Theory and practical application of heat pulse
682 to measure sap flow. *Agron J*, 95(6), 1371-1379. [doi:10.2134/agronj2003.1371](https://doi.org/10.2134/agronj2003.1371)

683 Gregersen, P., Brix, H. (2001). Zero-discharge of nutrients and water in a willow
 684 dominated constructed wetland. *Water Sci Technol*, 44(11-12), 407-412.
 685 <https://doi.org/10.2166/wst.2001.0859>
 686 Grenier, V., Pitre, F. E., Nissim, W. G., & Labrecque, M. (2015). Genotypic differences
 687 explain most of the response of willow cultivars to petroleum-contaminated soil.
 688 *Trees*, 29(3), 871-881. <https://doi.org/10.1007/s00468-015-1168->
 689 Grime, V.L., Sinclair, F.L. (1999). Sources of error in stem heat balance sap flow
 690 measurements. *Agr Forest Meteorol*, 94(2), 103-121. <https://doi.org/10.1016/S0168->
 691 1923(99)00011-8
 692 Grip, H., Halldin, S., Lindroth, A. (1989). Water use by intensively cultivated willow
 693 using estimated stomatal parameter values. *Hydrol Process*, 3(1), 51-63.
 694 <https://doi.org/10.1002/hyp.3360030106>
 695 Guidi, W., Labrecque, M. (2010). Effects of high water supply on growth, water use, and
 696 nutrient allocation in willow and poplar grown in a 1-year pot trial. *Water Air Soil*
 697 *Poll*, 207(1-4), 85-101. <https://doi.org/10.1007/s11270-009-0121-x>
 698 Guidi, W., Piccioni, E., Bonari, E. (2008). Evapotranspiration and crop coefficient of
 699 poplar and willow short-rotation coppice used as vegetation filter. *Bioresour*
 700 *Technol*, 99(11), 4832-4840. <https://doi.org/10.1016/j.biortech.2007.09.055>
 701 Guidi, W., Pitre, F.E., Labrecque, M. (2013). Short-rotation coppice of willows for the
 702 production of biomass in eastern Canada. In *Biomass now-sustainable growth and use*.
 703 Intech.
 704 Gullberg, U. (1993). Towards making willows pilot species for coppicing
 705 production. *Forest Chron*, 69(6), 721-726. <https://doi.org/10.5558/tfc69721-6>

- 706 Hall, R.L., Allen, S.J., Rosier, P.T., Hopkins, R. (1998). Transpiration from coppiced
707 poplar and willow measured using sap-flow methods. *Agr Forest Meteorol*, 90(4),
708 275-290. [https://doi.org/10.1016/S0168-1923\(98\)00059-8](https://doi.org/10.1016/S0168-1923(98)00059-8)
- 709 Halldin, S., Lindroth, A. (1989). Water use by willow in southern Sweden. *Proceedings*
710 *of estimation of areal evapotranspiration*, Vancouver, Canada, 177, 257-262.
- 711 Hao, X., Li, W., Deng, H. (2016). The oasis effect and summer temperature rise in arid
712 regions-case study in Tarim Basin. *Scientific reports*, 6, 35418.
713 <https://doi.org/10.1038/srep35418>
- 714 Hartwich, J., Schmidt, M., Bölscher, J., Reinhardt-Imjela, C., Murach, D., Schulte, A.
715 (2016). Hydrological modelling of changes in the water balance due to the impact of
716 woody biomass production in the North German plain. *Environ Earth Sci*, 75(14),
717 1071. <https://doi.org/10.1007/s12665-016-5870-4>
- 718 Hénault-Ethier, L., Gomes, M.P., Lucotte, M., Smedbol, É., Maccario, S., Lepage, L.,
719 Juneau, P., Labrecque, M. (2017). High yields of riparian buffer strips planted with
720 *Salix miyabena* ‘SX64’ along field crops in Québec, Canada. *Biomass and Bioenergy*,
721 105, 219-229. <https://doi.org/10.1016/j.biombioe.2017.06.017>
- 722 Holdridge, L.R. (1947). Determination of world plant formations from simple climatic
723 data. *Science*, 105(2727), 367-368. DOI: 10.1126/science.105.2727.367
- 724 Huang, J., Zhou, Y., Hou, R., & Wenninger, J. (2015). Simulation of water use dynamics
725 by *Salix* bush in a semiarid shallow groundwater area of the Chinese Erdos Plateau.
726 *Water*, 7(12), 6999-7021. <https://doi.org/10.3390/w7126671>
- 727 Huang, J., Zhou, Y., Yin, L., Wenninger, J., Zhang, J., Hou, G., Zhang, E., Uhlenbrook,
728 S. (2015). Climatic controls on sap flow dynamics and used water sources of *Salix*

729 *psammophila* in a semi-arid environment in northwest China. Environ Earth
 730 Sci, 73(1), 289-301. <https://doi.org/10.1007/s12665-014-3505-1>
 731 Iritz, Z., Tourula, T., Lindroth, A., Heikinheimo, M. (2001). Simulation of willow
 732 short- rotation forest evaporation using a modified Shuttleworth–Wallace
 733 approach. Hydrol Process, 15(1), 97-113. <https://doi.org/10.1002/hyp.118>
 734 Irmak, S., Kabenge, I., Rudnick, D., Knezevic, S., Woodward, D., Moravek, M. (2013).
 735 Evapotranspiration crop coefficients for mixed riparian plant community and
 736 transpiration crop coefficients for Common reed, Cottonwood and Peach-leaf willow
 737 in the Platte River Basin, Nebraska-USA. J Hydrol, 481, 177-190.
 738 <https://doi.org/10.1016/j.jhydrol.2012.12.032>
 739 Jing, D.B., Hu, H.Y. (2010). Chemical oxygen demand, nitrogen, and phosphorus
 740 removal by vegetation of different species in pilot-scale subsurface wetlands. Environ
 741 Eng Sci, 27(3), 247-253. <http://doi.org/10.1089/ees.2009.0440>
 742 Kabenge, I., Irmak, S. (2012). Evaporative losses from a common reed- dominated
 743 peachleaf willow and cottonwood riparian plant community. Water Resour Res, 48(9).
 744 <https://doi.org/10.1029/2012WR011902>
 745 Kadlec, R.H., Wallace, S. (2008). Treatment Wetlands (2nd Edition). CRC Press, United-
 746 States.
 747 Karp, A., Hanley, S. J., Trybush, S. O., Macalpine, W., Pei, M., & Shield, I. (2011).
 748 Genetic improvement of willow for bioenergy and biofuels free access. Journal of
 749 integrative plant biology, 53(2), 151-165.

- 1
2
3
4 750 Kopp, R.F., Smart, L.B., Maynard, C.A., Isebrands, J.G., Tuskan, G.A., Abrahamson,
5
6 751 L.P. (2001). The development of improved willow clones for eastern North
7
8
9 752 America. *Forest Chron*, 77(2), 287-292. <https://doi.org/10.5558/tfc77287-2>
10
11 753 Kučerová, A., Pokorný, J., Radoux, M., Nemcova, M., Cadelli, D., Dušek, J. (2001). In:
12
13
14 754 J. Vyzymal (Ed), *Transformations of nutrients in natural and constructed wetlands*,
15
16 755 Backhuys Publishers, Netherlands, p. 413-427.
17
18 756 Lachapelle-T, X., Labrecque, M., & Comeau, Y. (2019). Treatment and valorization of a
19
20
21 757 primary municipal wastewater by a short rotation willow coppice vegetation filter.
22
23 758 *Ecological Engineering*, 130, 32-44. <https://doi.org/10.1016/j.ecoleng.2019.02.003>
24
25 759 Langsrud, Ø. (2003). ANOVA for unbalanced data: Use Type II instead of Type III sums
26
27 760 of squares. *Stat Comput*, 13(2), 163-167. <https://doi.org/10.1023/A:1023260610025>
28
29 761 Lindegaard, K.N., Barker, J.H.A. (1997). Breeding willows for biomass. *Aspects of*
30
31 762 *Applied Biology*, 49, 155-162.
32
33 763 Linderson, M.L., Iritz, Z., Lindroth, A. (2007). The effect of water availability on stand-
34
35 764 level productivity, transpiration, water use efficiency and radiation use efficiency of
36
37 765 field-grown willow clones. *Biomass Bioenerg*, 31(7), 460-468.
38
39 766 <https://doi.org/10.1016/j.biombioe.2007.01.014>
40
41 767 Lindroth, A., Verwijst, T., Halldin, S. (1994). Water-use efficiency of willow: variation
42
43 768 with season, humidity and biomass allocation. *J Hydrol*, 156(1-4), 1-19.
44
45 769 [https://doi.org/10.1016/0022-1694\(94\)90068-X](https://doi.org/10.1016/0022-1694(94)90068-X)
46
47 770 Marmioli, M., Robinson, B.H., Clothier, B.E., Bolan, N.S., Marmioli, N., Schulin, R.
48
49 771 (2012). Effect of dairy effluent on the biomass, transpiration, and elemental
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

composition of *Salix kinuyanagi* Kimura. Biomass Bioenerg, 37, 282-288.
<https://doi.org/10.1016/j.biombioe.2011.12.001>

Marttila, H., Dudley, B. D., Graham, S., & Srinivasan, M. S. (2018). Does transpiration from invasive stream side willows dominate low- flow conditions? An investigation using hydrometric and isotopic methods in a headwater catchment. Ecohydrology, 11(2), e1930. <https://doi.org/10.1002/eco.1930>

Mata-González, R., Evans, T.L., Martin, D.W., McLendon, T., Noller, J.S., Wan, C., Sosebee, R.E. (2014). Patterns of water use by Great Basin plant species under summer watering. Arid Land Res Manag, 28(4), 428-446.
<https://doi.org/10.1080/15324982.2014.886088>

Martin, P.J., Stephens, W. (2006). Willow growth in response to nutrients and moisture on a clay landfill cap soil. I. Growth and biomass production. Bioresource Technol, 97(3), 437-448. <https://doi.org/10.1016/j.biortech.2005.03.003>

Mirck, J., & Volk, T. A. (2010). Response of three shrub willow varieties (*Salix* spp.) to storm water treatments with different concentrations of salts. Bioresource technology, 101(10), 3484-3492. <https://doi.org/10.1016/j.biombioe.2004.08.012>

Mirck, J., Volk, T.A. (2009). Seasonal sap flow of four *Salix* varieties growing on the Solvay wastebeds in Syracuse, NY, USA. Int J Phytoremediat, 12(1), 1-23.
<https://doi.org/10.1080/15226510902767098>

Nagler, P.L., Glenn, E.P., Thompson, T.L. (2003). Comparison of transpiration rates among saltcedar, cottonwood and willow trees by sap flow and canopy temperature methods. Agr Forest Meteorol, 116(1-2), 73-89. [https://doi.org/10.1016/S0168-1923\(02\)00251-4](https://doi.org/10.1016/S0168-1923(02)00251-4)

- 1
2
3
4 795 Ngouaijo, M. (2001). Using the right planting density is critical for optimum yield and
5
6 796 revenue for vegetable crops. Michigan State University, United-States.
7
8
9 797 <https://www.canr.msu.edu/outreach>
10
11 798 Nissim, W.G., Voicu, A., & Labrecque, M. (2014). Willow short-rotation coppice for
12
13 799 treatment of polluted groundwater. *Ecol Eng*, 62, 102-114.
14
15 800 <https://doi.org/10.1016/j.ecoleng.2013.10.005>
16
17 801 Novák, V. (2012). *Evapotranspiration in the soil-plant-atmosphere System*. Springer
18
19 802 Science & Business Media, Germany.
20
21 803 Pauliukonis, N., Schneider, R. (2001). Temporal patterns in evapotranspiration from
22
23 804 lysimeters with three common wetland plant species in the eastern United
24
25 805 States. *Aquat Bot*, 71(1), 35-46. [https://doi.org/10.1016/S0304-3770\(01\)00168-1](https://doi.org/10.1016/S0304-3770(01)00168-1)
26
27 806 Payero, J.O., Tarkalson, D.D., Irmak, S., Davison, D., Petersen, J.L. (2008). Effect of
28
29 807 irrigation amounts applied with subsurface drip irrigation on corn evapotranspiration,
30
31 808 yield, water use efficiency, and dry matter production in a semiarid climate. *Agr*
32
33 809 *Water Manage*, 95(8), 895-908. <https://doi.org/10.1016/j.agwat.2008.02.015>
34
35 810 Peng, X., Fan, J., Wang, Q., Warrington, D. (2015). Discrepancy of sap flow in *Salix*
36
37 811 *matsudana* grown under different soil textures in the water-wind erosion crisscross
38
39 812 region on the Loess Plateau. *Plant Soil*, 390(1-2), 383-399.
40
41 813 <https://doi.org/10.1007/s11104-014-2333-0>
42
43 814 Persson, G. (1995). Willow stand evapotranspiration simulated for Swedish soils. *Agr*
44
45 815 *Water Manage*, 28(4), 271-293. [https://doi.org/10.1016/0378-3774\(95\)01182-X](https://doi.org/10.1016/0378-3774(95)01182-X)
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

- 1
2
3
4 816 Pezeshki, S.R., Li, S., Shields Jr, F.D., Martin, L.T. (2007). Factors governing survival of
5
6 817 black willow (*Salix nigra*) cuttings in a streambank restoration project. Ecol Eng,
7
8
9 818 29(1), 56-65. <https://doi.org/10.1016/j.ecoleng.2006.07.014>
10
11 819 Pistocchi, C., Guidi, W., Piccioni, E., Bonari, E. (2009). Water requirements of poplar
12
13 820 and willow vegetation filters grown in lysimeter under Mediterranean conditions:
14
15 821 results of the second rotation. Desalination, 246(1-3), 137-146.
16
17 822 <https://doi.org/10.1016/j.desal.2008.03.047>
18
19 823 Přibáň, K., Ondok, J.P. (1986). Evapotranspiration of a willow carr in summer. Aquat
20
21 824 Bot, 25, 203-216. [https://doi.org/10.1016/0304-3770\(86\)90055-0](https://doi.org/10.1016/0304-3770(86)90055-0)
22
23 825 Rana, G., Katerji, N. (2000). Measurement and estimation of actual evapotranspiration in
24
25 826 the field under Mediterranean climate: a review. Eur J Agron, 13(2-3), 125-153.
26
27 827 [https://doi.org/10.1016/S1161-0301\(00\)00070-8](https://doi.org/10.1016/S1161-0301(00)00070-8)
28
29 828 Rawls, W.J., Brakensiek, D.L., Saxton, K.E. (1982). Estimation of soil water properties.
30
31 829 T ASAE, 25(5), 1316-1320. doi: 10.13031/2013.33720
32
33 830 Royygar, J.K.F., Bolan, N.S., Clothier, B.E., Green, S.R., Sims, R.E.H. (1999). Short
34
35 831 rotation forestry for land treatment of effluent: a lysimeter study. Soil Res, 37(5), 983-
36
37 832 992. <https://doi.org/10.1071/SR98067>
38
39 833 R  th, B., Lennartz, B., Kahle, P. (2007). Water regime of mechanical—biological
40
41 834 pretreated waste materials under fast-growing trees. Waste Manage Res, 25(5), 408-
42
43 835 416. <https://doi.org/10.1177/0734242X07076940>
44
45 836 Saxton, K.E., Rawls, W.J., Romberger, J.S., Papendick, R.I. (1986). Estimating
46
47 837 generalized soil-water characteristics from texture1. Soil Sci Soc America J, 50, 1031.
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

838 Scheirlink, H., Lust, N., Nachtergale, L. (1996). Transpiration of two willow species
839 (*Salix viminalis* and *Salix triandra*) growing on a landfill of dredged sludge. Silva
840 Gandavensis, 61, 33-45.

841 Schmidt-Walter, P., Richter, F., Herbst, M., Schuldt, B., Lamersdorf, N.P. (2014).
842 Transpiration and water use strategies of a young and a full-grown short rotation
843 coppice differing in canopy cover and leaf area. Agr Forest Meteorol, 195, 165-178.
844 <https://doi.org/10.1016/j.agrformet.2014.05.006>

845 Smart, L.B., Cameron, K.D. (2008). In:W. Vermerris (ed.), Genetic improvement of
846 bioenergy crops, Springer Science+Business Media, Germany, p. 377-396.

847 Smart, L.B., Volk, T.A., Lin, J., Kopp, R.F., Phillips, I.S., Cameron, K.D., White, E.H.,
848 Abrahamson, L.P. (2005). Genetic improvement of shrub willow (*Salix* spp.) crops for
849 bioenergy and environmental applications in the United States. Unasylva, 56(2), 51-
850 55.

851 Sperry, J.S. (2000). Hydraulic constraints on plant gas exchange. Agr Forest Meteorol,
852 104(1), 13-23. [https://doi.org/10.1016/S0168-1923\(00\)00144-1](https://doi.org/10.1016/S0168-1923(00)00144-1)

853 Stephens, W., Tyrrel, S.F., Tiberghien, J.E. (2000). Irrigating short rotation coppice with
854 landfill leachate: constraints to productivity due to chloride. Bioresource
855 Technol, 75(3), 227-229. [https://doi.org/10.1016/S0960-8524\(00\)00065-1](https://doi.org/10.1016/S0960-8524(00)00065-1)

856 Tallis, M.J., Casella, E., Henshall, P.A., Aylott, M.J., Randle, T.J., Morison, J.I., Taylor,
857 G. (2013). Development and evaluation of forest growth- SRC a process- based
858 model for short rotation coppice yield and spatial supply reveals poplar uses water
859 more efficiently than willow. Gcb Bioenerg, 5(1), 53-66.
860 <https://doi.org/10.1111/j.1757-1707.2012.01191.x>

- 1
2
3
4 861 Trapp, S., Zambrano, K.C., Kusk, K.O., Karlson, U. (2000). A phytotoxicity test using
5
6 862 transpiration of willows. Arch Environ Con Tox, 39(2), 154-160.
7
8 863 <https://doi.org/10.1007/s002440010091>
9
10
11 864 Volk, T.A., Abrahamson, L.P., Nowak, C.A., Smart, L.B., Tharakan, P.J., White, E.H.
12
13
14 865 (2006). The development of short-rotation willow in the northeastern United States for
15
16 866 bioenergy and bioproducts, agroforestry and phytoremediation. Biomass
17
18 867 Bioenerg, 30(8-9), 715-727. <https://doi.org/10.1016/j.biombioe.2006.03.001>
19
20
21 868 Walle, I.V., Van Camp, N., Van de Castele, L., Verheyen, K., Lemeur, R. (2007). Short-
22
23 869 rotation forestry of birch, maple, poplar and willow in Flanders (Belgium) I—Biomass
24
25 870 production after 4 years of tree growth. Biomass Bioenerg, 31(5), 267-275.
26
27 871 <https://doi.org/10.1016/j.biombioe.2007.01.019>
28
29
30
31 872 Waller, P., Yitayew, M. (2015). Irrigation and drainage engineering. Springer
32
33 873 Science+Business Media, Germany.
34
35
36 874 Wang, S., Fan, J., Ge, J., Wang, Q., & Fu, W. (2019). Discrepancy in tree transpiration of
37
38 875 *Salix matsudana*, *Populus simonii* under distinct soil, topography conditions in an
39
40 876 ecological rehabilitation area on the Northern Loess Plateau. Forest Ecology and
41
42 877 Management, 432, 675-685. <https://doi.org/10.1016/j.foreco.2018.10.011>
43
44
45 878 Wang, S., Fan, J., Wang, Q. (2015). Determining evapotranspiration of a Chinese willow
46
47 879 stand with three-needle heat-pulse probes. Soil Sci Soc Am J, 79(6), 1545-1555.
48
49 880 [doi:10.2136/sssaj2015.05.0180](https://doi.org/10.2136/sssaj2015.05.0180)
50
51
52
53 881 Willebrand, E., Ledin, S., Verwijst, T. (1993). Willow coppice systems in short rotation
54
55 882 forestry: effects of plant spacing, rotation length and clonal composition on biomass
56
57
58
59
60
61
62
63
64
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production. Biomass Bioenerg, 4(5), 323-331. [https://doi.org/10.1016/0961-9534\(93\)90048-9](https://doi.org/10.1016/0961-9534(93)90048-9)

Willebrand, E., Verwijst, T. (1993). Population dynamics of willow coppice systems and their implications for management of short-rotation forests. Forest Chron, 69(6), 699-704. <https://doi.org/10.5558/tfc69699-6>

Witters, N., Van Slycken, S., Ruttens, A., Adriaensen, K., Meers, E., Meiresonne, L., Filip, M.G., Tack, T.T., Laes, E., Vangronsveld, J. (2009). Short-rotation coppice of willow for phytoremediation of a metal-contaminated agricultural area: a sustainability assessment. BioEnergy Res, 2(3), 144-152. <https://doi.org/10.1007/s12155-009-9042-1>

Xiao, C.W., Zhou, G.S., Zhang, X.S., Zhao, J.Z., Wu, G. (2005). Responses of dominant desert species *Artemisia ordosica* and *Salix psammophila* to water stress. Photosynthetica, 43(3), 467-471. <https://doi.org/10.1007/s11099-005-0075-1>

Yin, L., Zhou, Y., Huang, J., Wenninger, J., Hou, G., Zhang, E., Wang, X., Dong, J., Zhang, J., Uhlenbrook, S. (2014). Dynamics of willow tree (*Salix matsudana*) water use and its response to environmental factors in the semi-arid Hailiutu River catchment, Northwest China. Environ Earth Sci, 71(12), 4997-5006. <https://doi.org/10.1007/s12665-013-2891-0>

Yoder, K.S., Moser, B.C. (1993). 696 (PS 5) pussy willow branches – A new crop for sustainable agriculture. Hort Sci, 28(5), 551-551.

Zsuffa, L., Mosseler, A., Raj, Y. (1984). In: K.L. Perttu (Eds.), Ecology and management of forest biomass production systems, Swedish University of Agricultural Sciences, Sweden, p. 261–281.

Figure 1
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WILLOWS FOR ENVIRONMENTAL PROJECTS: A LITERATURE
REVIEW OF RESULTS ON EVAPOTRANSPIRATION RATE AND
ITS DRIVING FACTORS ACROSS THE GENUS SALIX.

Frédette *et al.*

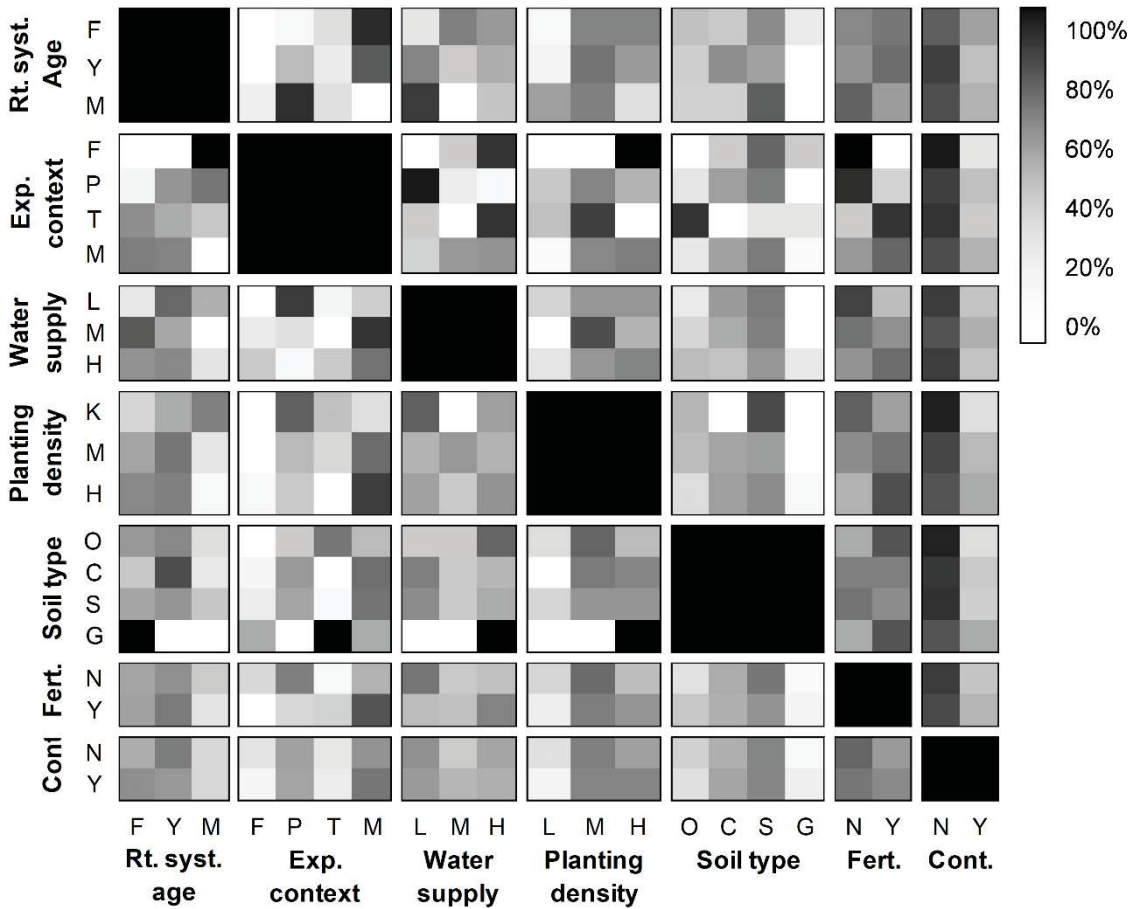


Figure 1. Correlogram illustrating the frequency (%) of association between the levels of nine variables selected to explain the variation of evapotranspiration rate across the willow genus (*Salix* sp.). Darker colors indicate a frequent association between levels of two variables (black = 100%, *i.e.* levels always associated), while pale colors indicate that the levels of the two variables were not likely to be combined (white = 0%, *i.e.* levels never associated). The codes used for variables levels are detailed in table 1 of the present article.

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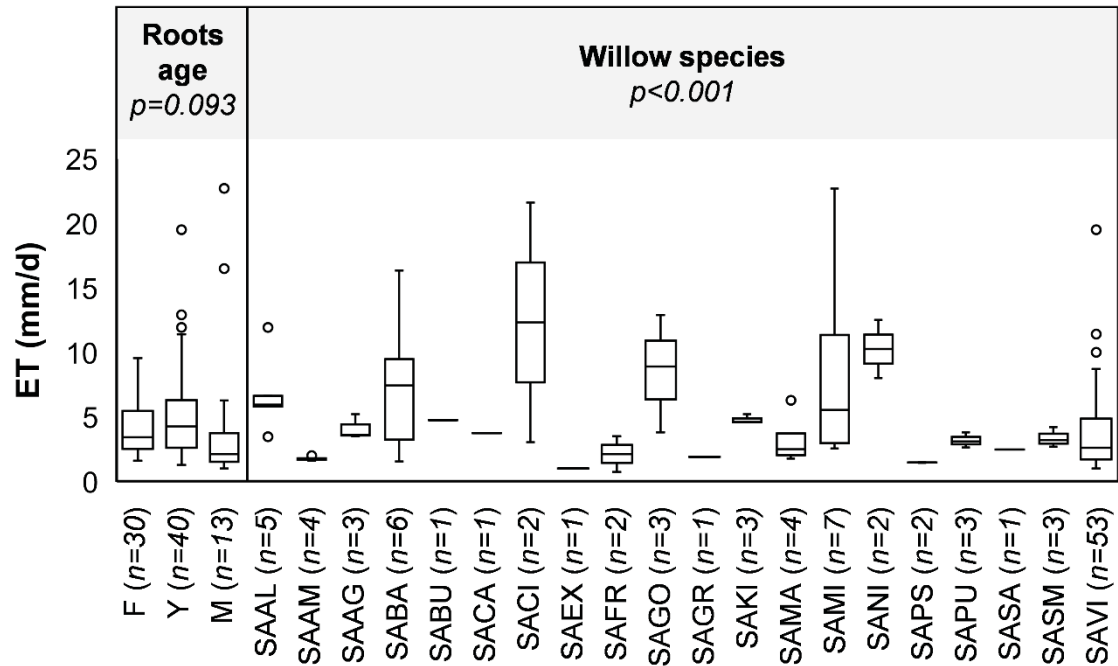


Figure 2. Mean evapotranspiration (ET) rates reported in 57 articles in 16 countries, according to plant related variables (root system age and species). Numbers in parenthesis (n) represent the number of average results considered for each variable level. The codes used for variables levels are detailed in table 1 of the present article. P values indicate if the variables affect significantly ($\alpha=0.05$) ET results according to a Type II ANOVA analysis testing the simultaneous effect of 10 variables.

Size: 2 columns

Figure 3
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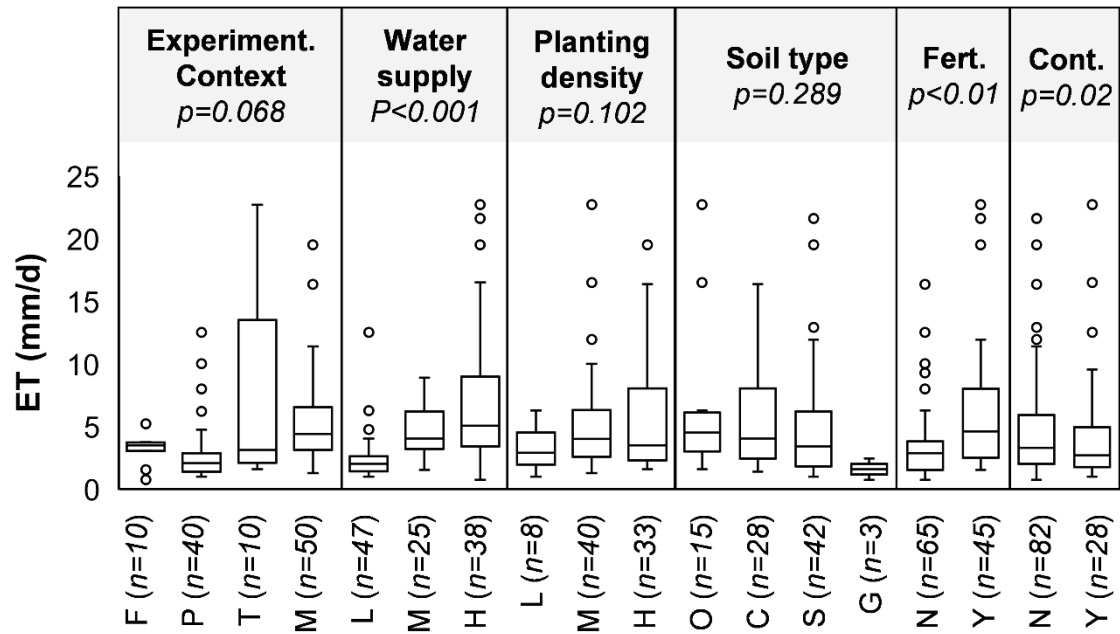


Figure 3. Mean evapotranspiration (ET) rates reported in 57 articles in 16 countries, according to experimental/management variables (experimental context, water supply, planting density, dominant soil type, fertilization and contamination). Numbers in parenthesis (n) represent the number of average results considered for each variable level. The codes used for variables levels are detailed in table 1 of the present article. P values indicate if the variables affect significantly ($\alpha=0.05$) ET results according to a Type II ANOVA analysis testing the simultaneous effect of 10 variables.

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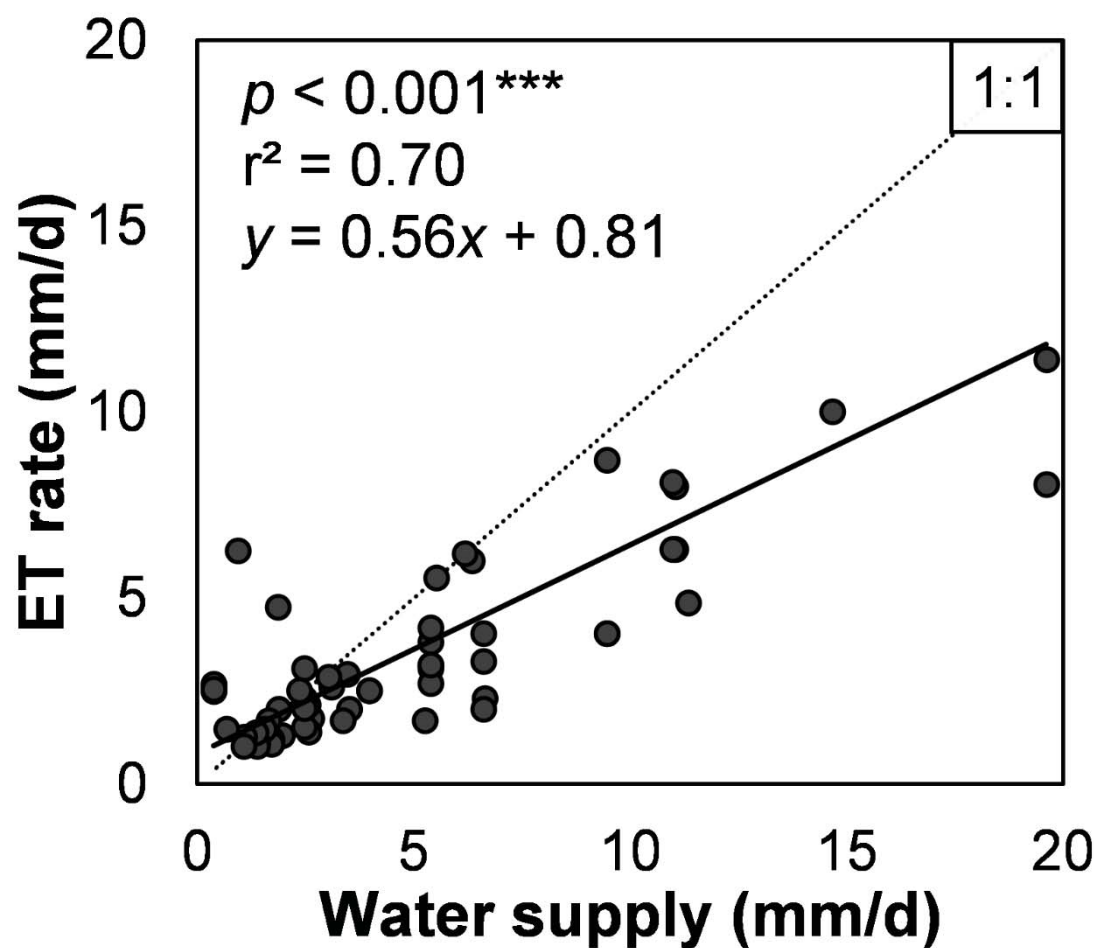


Figure 4. Summary of the linear regression between mean daily evapotranspiration rate of willows reported in scientific literature and the amount of water supplied daily, either by precipitation or irrigation ($n = 63$). Reference articles included in this analysis are detailed in Table 2 of the present article, and are comprised of studies of open systems with water table low enough to allow drainage.

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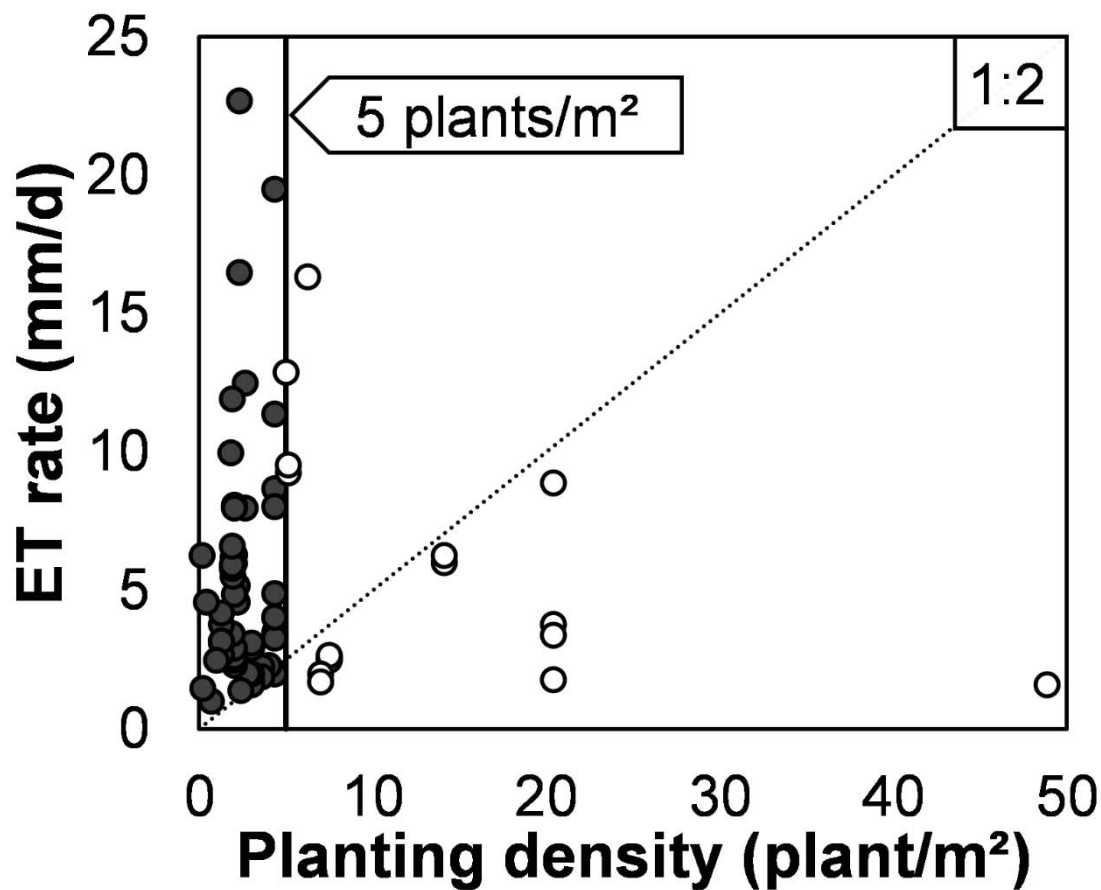


Figure 5. Mean daily evapotranspiration rate of willows reported in scientific literature in relation to planting density ($n = 75$). Reference articles included in this analysis are detailed in Table 2 of the present article. An arbitrary threshold (dashed line) for ET was drawn at a planting density of 5 trees per m^2 .

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Table 1
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WILLOWS FOR ENVIRONMENTAL PROJECTS: A LITERATURE REVIEW OF RESULTS ON EVAPOTRANSPIRATION RATE AND ITS DRIVING FACTORS ACROSS THE GENUS SALIX. Frédette *et al.*

Table 1. Summary of ten variables selected to categorize, compare and identify driving factors of willow (*Salix* sp.) evapotranspiration rates results found in the scientific literature.

Type	Variable	Levels	Description	Code
Plant variables	Willow species	19 species (see Table 2 for species listing and codes)		
	Age of	First year	Establishment year	F
	plantation	Young	2 to 5 years old roots	Y
		Mature	> 5 years old roots	M
Environmental/management variables	Experimental	Flood plain	Natural stands in wet habitat	F
	context	Plantation	Mand made plantation or natural stand	P
			in mesic to dry habitat	
		Treatment wetland	Pilot and full-scale	T
		Mesocosm	Lysimeters and pots	M
	Water supply	Low	> 10 mm/d or saturated root zone	L
		Medium	5 to 10 mm/d or field capacity	M
		High	< 5mm/d or water deficit	H
	Planting	Low	≤ 1 plants/m ²	L
	density	Medium	1 to 4 plants/m ²	M
		High	> 4 plants/m ²	H
	Dominant soil	Organic	Significant organic matter content	O
	type	Clay	> 50% clay particles	C
		Sand	> 50% sand particles	S
		Gravel	> 50% gravel content	G
	Fertilization	Yes	Fertilizer, soil amendment or nutrient	Y
			rich wastewaters	
		No		N
	Contamination	Yes	Soil or water contamination	Y
		No		N

Table 2
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Table 2. Range of evapotranspiration rates (mm/d) reported in 57 articles for 19 different willow species (and various cultivars) in 16 countries, along with the corresponding information about plants, experimental and methodological variables. Results of transpiration only are indicated in parenthesis (T). Information missing about some variables is due either to non-reported information or to values that did not fitted the selected levels of a variable. Numerical value of water supply and planting density are detailed in parenthesis when available. The codes used for variables levels are detailed in table 1 of the present article. Each article tested one to nine experimental treatments (n), for a total of 110 mean results considered for comparative analysis.

Species 'cultivar'	Code	ET range (mm/d)	Age	Context Water	(mm/d)	Density (plant/m ²)	Soil	Fert.	Cont.	n	Country	Ref.
<i>S. alba</i> 'SI62-059'	SAAL	3.4-11.9	F, Y	M	M	M (1.9)	S	Y, N	N	4	Italy	1
<i>S. alba</i> 'SI62-059'	SAAL	4.6-7.0	Y	M	M	M (1.9)	S	Y	N	1	Italy	2
<i>S. amygdalina</i>	SAAM	0.6-2.3	F	M	H	H (48.8)	G	Y	Y	1	Poland	3
<i>S. amygdalina</i>	SAAM	1.0-3.0	F, Y	M	L, M (3.4-5.3)	H (7)	S	Y	Y	3	Poland	4
<i>S. amygdaloïdes</i>	SAAG	3.6-5.2	-	F	H, M	-	S	N	N	2	U.S.	5
<i>S. amygdaloïdes</i>	SAAG	3.5 (T)	-	F	H	-	S	N	N	1	U.S.	6
<i>S. babylonica</i>	SABA	1.5-6.6	-	F	H, M	-	-	N	N	2	Australia	7
<i>S. babylonica</i>	SABA	2.4	F	T	H	-	G	Y	N	1	China	8
<i>S. babylonica</i>	SABA	9.3-9.6	F	M	H	H (5.1)	C	Y, N	Y,N	2	Canada	9
<i>S. babylonica</i>	SABA	16.4	-	M	H	H (6.25)	C	N	N	1	U.S.	10
<i>S. bujartica</i> 'Germany'	SABU	4.8 (T)	Y	P	L (1.9)	-	C	N	N	1	Sweden	11

<i>S. caroliniana</i>	SACA	3.8	M	F	H	-	-	N	Y	1	U.S.	12
<i>S. cinerea</i>	SACI	21.6	-	T	H	-	S	Y	N	1	Belgium	13
<i>S. cinerea</i>	SACI	3.0	-	F	H	H	C	N	N	1	Czechoslovakia	14
<i>S. exigua</i>	SAEX	0.7-1.6	M	P	L (1.1)	L (0.7)	S	N	N	1	U.S.	15
<i>S. fragilis</i>	SAFR	3.5	-	F	H	-	-	N	N	1	Australia	16
<i>S. fragilis</i>	SAFR	0.7	-	F	H	-	G	N	N	1	New-Zeland	17
<i>S. gooddingii</i>	SAGO	2.5-8.9 (T)	F	M	M	H (20.4)	S	Y	Y, N	2	U.S.	18
<i>S. gooddingii</i>	SAGO	12.9 (T)	Y	M	H	H (5.0)	S	N	N	1	U.S.	19
<i>S. gordejvii</i>	SAGR	1.9 (T)	-	P	L	H (3.6)	S	N	N	1	China	20
<i>S. kinuyanagi</i> 'Kimura'	SAKI	4.6-5.4	F	M	H	M (2.2)	S	Y, N	Y, N	2	New-Zeland	21
<i>S. kinuyanagi</i> 'Kimura'	SAKI	4.6	Y	M	H	L (0.4)	S	Y	Y	1	New-Zeland	22
<i>S. matsudana</i>	SAMA	2.1	M	P	L (2.6)	L	S	N	N	1	China	23
<i>S. matsudana</i>	SAMA	1.8	M	P	L (2.7)	L	S	N	N	1	China	24
<i>S. matsudana</i>	SAMA	6.3	M	P	L (0.9)	L (0.2)	S	N	N	1	China	25
<i>S. matsudana</i>	SAMA	1.2-5.3 (T)	M	P	L (3.0)	-	S	N	N	1	China	26
<i>S. miyabeana</i> 'SX67'	SAMI	16.5-22.7	M	T	H	M (2.3)	O	Y	Y	2	Canada	27
<i>S. miyabeana</i> 'SX67'	SAMI	5.5-6.2	F, Y	P	M (5.5-6.2)	M (2.0)	O	N	N	2	Canada	28
<i>S. miyabeana</i> 'SX64'	SAMI	2.5-2.7 (T)	Y	P	L (0.4)	H (7.5)	-	N	Y	1	U.S.	29
<i>S. miyabeana</i> 'SX64'	SAMI	2.7-3.9	F	M	M (5.4)	M (1.3)	-	N	Y, N	2	U.S.	30
<i>S. nigra</i>	SANI	6.0-13.0	Y	P	L, M	M (2.6)	C	N	Y	2	U.S.	31

		(T)										
<i>S. psammophila</i>	SAPS	1.5 (T)	-	P	L (1.6)	-	S	N	N	1	China	32
<i>S. psammophila</i>	SAPS	1.4	-	P	L	L (0.2)	S	N	N	1	China	33
<i>S. purpurea</i> '9882-34'	SAPU	3.1-3.8	F	M	M (5.4)	M (1.3)	-	N	Y, N	2	U.S.	30
<i>S. purpurea</i> '9882-34'	SAPU	2.6 (T)	Y	P	L (0.4)	H (7.5)	-	N	Y	1	U.S.	29
<i>S. sachalinensis</i> 'SX61'	SASA	2.5 (T)	Y	P	L (0.4)	H (7.5)	-	N	Y	1	U.S.	29
<i>S. sachalinensis</i> x <i>S. miyabeana</i> '9870-40'	SSSM	3.2-4.2	F	M	M (5.4)	M (1.3)	-	N	Y, N	2	U.S.	30
<i>S. sachalinensis</i> x <i>S. miyabeana</i> '9870-23'	SSSM	2.7 (T)	Y	P	L (0.4)	H (7.5)	-	N	Y	1	U.S.	29
<i>S. viminalis</i>	SAVI	10.0	-	P	H (14.7)	M (1.79)	S	N	N	1	Switzerland	34
<i>S. viminalis</i> '1023' '1047' '1052' '1054'	SAVI	1.4-1.7	-	P	L (1.4-1.7)	-	C, S	N	N	2	Poland	35
<i>S. viminalis</i> 'Inger' 'Sven' 'Tordis' 'Torhild'	SAVI	1.9-7.6	Y	M	H	H (4.35)	O	Y, N	N	2	Ireland	36
<i>S. viminalis</i>	SAVI	1.5-2.9	F, Y	T	L, H	M (3.0)	O	Y, N	N	4	Ireland	37
<i>S. viminalis</i> '78-183'	SAVI	6.3-8.3	Y	M	H (11.0)	M (2.0)	C, S	Y	N	2	Sweden	38
<i>S. viminalis</i> 'Tora'	SAVI	2.2-7.5	F, Y	M	M, H (6.4-11.4)	M (2.0)	C	Y	Y	3	Sweden	39
<i>S. viminalis</i> 'Tora'	SAVI	2.3-8.3	Y	M	L, H (4.0-11.0)	M (2.0)	C, S	Y	N	4	Sweden	40
<i>S. viminalis</i> 'Bjorn' 'Tora' 'Jorr'	SAVI	2.7-5.7	F, Y	T	H	L	O	Y	N	2	Denmark	41
<i>S. viminalis</i> '77683' '77666'	SAVI	3.0	Y	M	L	-	S	N	N	1	Sweden	42
<i>S. viminalis</i> 'SQV 5027'	SAVI	6.0-6.3	F	M	M, H	H (14.1)	O	Y	N	2	Canada	43
<i>S. viminalis</i>	SAVI	2.6	Y	P	L (3.1)	M (2.0)	C	Y	N	1	Sweden	44
<i>S. viminalis</i> 'L78183' 'Loden' 'Jorr' 'Rapp' 'Tora'	SAVI	0.7-2.1 (T)	Y	P	L (2.6)	M (2.4)	C	N	N	1	Sweden	45

<i>S. viminalis</i>	SAVI	2.9-3.0	Y	P	L (3.5)	M (2.0)	C	Y	N	1	Sweden	46
<i>S. viminalis</i> 'Jorr'	SAVI	2.0-19.5	F, Y	M	L, M, H (6.6-19.6)	H (4.4)	C, S	Y, N	N	9	U.K.	47
<i>S. viminalis</i> '77075' '77077' '77082' '77083'	SAVI	2.0-3.7	Y, M	P	L (2.5)	M, H	C, S,	Y	N	5	Sweden	48
'77683' '82007'						(3.0-4.0)	O					
<i>S. viminalis</i>	SAVI	1.6-2.3	-	P	L (1.9)	-	C	N	N	1	Sweden	49
<i>S. viminalis</i> 'Régalis'	SAVI	1.0-1.2	-	P	L (1.4-1.7)	-	S	N	Y, N	6	Germany	50
<i>S. viminalis</i>	SAVI	1.2 (T)	M	P	L (1.0)	-	-	N	Y	1	Belgium	51
<i>S. viminalis</i> 'Tora'	SAVI	1.3-1.5	Y, M	P	L (0.7-1.1)	M	C, S	N	N	1	Germany	52
<i>S. viminalis</i> 'Q683'	SAVI	1.8-3.4	1	M	H	H (20.4)	S	N	Y, N	2	U.K.	53
<i>S. viminalis</i> 'Jorunn'	SAVI	2.5 (T)	-	P	L (2.4)	L (1.0)	-	N	N	1	U.K.	54
<i>Salix</i> sp.	SASP	3.1	-	P	L	-	C	N	N	1	Sweden	55
<i>Salix</i> sp.	SASP	3.1 (T)	-	F	H	-	-	N	N	1	U.S.	56
<i>Salix</i> sp.	SASP	1.1-1.4	-	P	L (2.0)	-	-	N	N	1	Germany	57

1. Guidi *et al.*, 2008; 2. Pistocchi *et al.*, 2009; 3. Białowiec *et al.*, 2003; 4. Białowiec *et al.*, 2007; 5. Kabenge *et al.*, 2012; 6. Irmak *et al.*, 2013; 7. Doody and Benyon, 2011; 8. Jing *et al.*, 2010; 9. Cureton *et al.*, 1991; 10. Pauliukonis et Schneider, 2001; 11. Hall *et al.*, 1998; 12. Duan *et al.*, 2017; 13. Kučerová *et al.*, 2001; 14. Přibáň and Ondok , 1986; 15. Mata-González *et al.*, 2014; 16. Doody *et al.*, 2011; 17. Marttila *et al.*, 2018; 18. Glenn *et al.*, 1998; 19. Nagler *et al.*, 2003; 20. Duan *et al.*, 2017; 21. Marmioli *et al.*, 2012; 22. Royygard *et al.*, 1999; 23. Wang *et al.*, 2015; 24. Wang *et al.*, 2019; 25. Yin *et al.*, 2014; 26. Peng *et al.*, 2015; 27. Frédette *et al.*, 2018; 28. Guidi Nissim *et al.*, 2014; 29. Mirck and Volk, 2009; 30. Mirck and Volk, 2010; 31. Conger and Potier, 2001; 32. Huang *et al.*, 2014; 33. Huang *et al.*, 2015; 34. Benettin *et al.*, 2018; 35. Borek *et al.*, 2010;

36. Curneen and Gill, 2014; 37. Curneen and Gill, 2016; 38. Dimitriou *et al.*, 2004; 39. Dimitriou *et al.*, 2010; 40. Dimitriou *et al.*, 2010; 41. Brix and Gregersen, 2001; 42. Grip *et al.*, 1989; 43. Nissim and Labrecque, 2010; 44. Iritz *et al.*, 2001; 45. Linderson *et al.*, 2007; 46. Lindroth *et al.*, 1994; 47. Martin and Stephens, 2006; 48. Persson, 1995; 49. Persson, 1997; 50. R  th *et al.*, 2007; 51. Scheirlink *et al.*, 1996; 52. Schmidt-Walter *et al.*, 2012; 53. Stephens *et al.*, 2000; 54. Tallis *et al.*, 2013; 55. Halldin and Lindroth, 1989; 56. Budny and Benscoter, 2016; 57. Hartwich *et al.*, 2016.