



<b>Titre:</b> Title:	Comparison of Synthetic Sweat and Influence of Sebum in the Permeation of Bioaccessible Metal(loid)s from Contaminated Soils through a Synthetic Skin Membrane
<b>Auteurs:</b> Authors:	Carlos Augusto Marin Villegas, & Gérald J. Zagury
<b>Date:</b>	2021
<b>Type:</b>	Article de revue / Article
<b>Référence:</b> Citation:	Marin Villegas, C. A., & Zagury, G. J. (2021). Comparison of Synthetic Sweat and Influence of Sebum in the Permeation of Bioaccessible Metal(loid)s from Contaminated Soils through a Synthetic Skin Membrane. <i>Environmental Science &amp; Technology</i> , 55(12), 8215-8222. <a href="https://doi.org/10.1021/acs.est.1c02038">https://doi.org/10.1021/acs.est.1c02038</a>

## Document en libre accès dans PolyPublie

Open Access document in PolyPublie

**URL de PolyPublie:** <https://publications.polymtl.ca/48017/>  
PolyPublie URL:

**Version:** Version finale avant publication / Accepted version  
Révisé par les pairs / Refereed

**Conditions d'utilisation:** Tous droits réservés / All rights reserved  
Terms of Use:

## Document publié chez l'éditeur officiel

Document issued by the official publisher

**Titre de la revue:** Environmental Science & Technology (vol. 55, no. 12)  
Journal Title:

**Maison d'édition:** American Chemical Society (ACS)  
Publisher:

**URL officiel:** <https://doi.org/10.1021/acs.est.1c02038>  
Official URL:

**Mention légale:** This document is the Accepted Manuscript version of a Published Work that appeared in final form in Environmental Science & Technology (vol. 55, no. 12), copyright © American Chemical Society after peer review and technical editing by the publisher. To access the final edited and published work see <https://doi.org/10.1021/acs.est.1c02038>  
Legal notice:

1   **Comparison of synthetic sweat and influence of sebum in the permeation of**  
2   **bioaccessible metal(loid)s from contaminated soils through a synthetic skin membrane**

3   Carlos A. Marin Villegas, Gerald J. Zagury\*

4   Department of Civil, Geological and Mining Engineering, Polytechnique Montréal,  
5   Montreal (QC), Canada, H3C 3A7

6   \* Corresponding author: [gerald.zagury@polymtl.ca](mailto:gerald.zagury@polymtl.ca)

7   + 1 514 340-4711 ext.: 4980

8

9   **Abstract**

10   Dermal exposure to metal(loid)s from contaminated soils has received less attention than oral  
11   and inhalation exposure. Still, it can be a relevant pathway for some contaminants.

12   Comparison of synthetic sweats (donor solutions), the influence of sebum, and the  
13   characterization of diffusion parameters through a synthetic membrane (acting as skin  
14   surrogate) in the permeation of metal(loid)s (As, Cr, Cu, Ni, Pb, and, Zn) from polluted soils  
15   is missing. Dermal bioaccessibility test were performed using two sweat compositions (EN  
16   1811, pH 6.5 (Sweat A) and NIHS 96-10, pH 4.7 (Sweat B)). Diffusion parameters of soluble  
17   metal(loid)s using the Franz cell methodology were calculated using the Strat-M membrane.  
18   The influence of synthetic sebum in the permeation of metal(loid)s was also investigated.

19   The metal(loid) bioaccessibility percentage was higher for Sweat B (pH 4.7) compared to  
20   Sweat A (pH 6.5), attributed to lower pH of sweat B. Among the six elements tested, only  
21   chromium and copper permeated the membrane. Permeation coefficient ( $K_p$ ) was higher for  
22   chromium in Sweat A ( $0.05$  to  $0.11$   $\text{cm h}^{-1}$ ) than Sweat B ( $0.0007$  to  $0.0037$   $\text{cm h}^{-1}$ ) likely  
23   due to a higher pH and thus more permeable Cr species. Presence of sebum increased lag  
24   times for copper permeation. Additional studies regarding speciation of metal(loid)s

25 following extractions in synthetic sweat and comparison of synthetic membrane Strat-M and  
26 human skin in the permeation of metal(loid)s are recommended.

27

28 **Key words:** contaminated soils, dermal permeation, Franz diffusion cell, *in vitro* dermal  
29 bioaccessibility, synthetic sebum, Strat-M membrane, synthetic sweat.

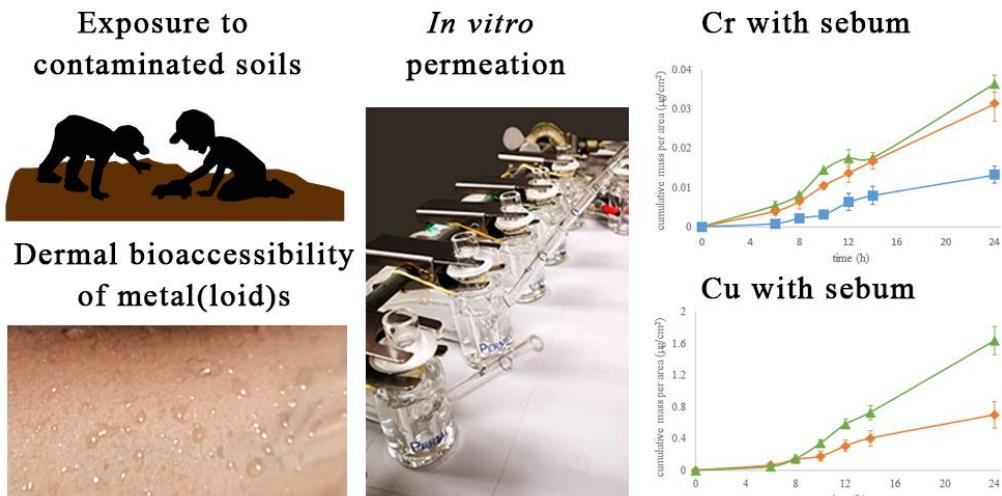
30

31 **Synopsis:** When metal(loid)s found in polluted soils are extractable in sweat they don't  
32 indeed permeate a synthetic skin membrane

33

34 **Graphical abstract TOC**

35



36

37

38

39

40

41

42        **1. Introduction**

43        Humans can be exposed to toxic metal(loid)s in various environmental media, including  
44        soils, via the dermal pathway. The majority of scientific studies have focused on oral  
45        bioaccessibility of pollutants (ingestion pathway) and more recently, on *in vitro* test  
46        development and application for assessing inhalation bioavailability <sup>1-4</sup> and *in vitro* dermal  
47        bioavailability. <sup>5-8</sup>

48        It is generally acknowledged that there are very few quantitative and qualitative data on  
49        dermal exposure to toxic chemicals present in geological materials <sup>9</sup>. Additionally, some of  
50        the permeation data of chemicals through the skin were obtained using different experimental  
51        conditions in terms of concentration in the donor solution, type of membrane (animal, ex  
52        vivo), and duration, making a comparison of results difficult. <sup>10</sup> Quite the reverse, there is a  
53        copious amount of data regarding the passage of drugs and cosmetics through the skin.  
54        However, very little is known about this route of entry for environmental contaminants bound  
55        to geological materials. The technology used to study the percutaneous penetration of drugs  
56        could be used to assess dermal penetration of toxic compounds as well. <sup>9-11</sup>

57        The most common method to evaluate dermal diffusion was introduced by Franz <sup>12</sup> in 1975.  
58        This method has been adapted and modified by many researchers to study the dermal  
59        pathway for drugs, cosmetics, and chemicals due to its low cost, short time, and good  
60        reproducibility. <sup>9</sup>

61        Franken et al.<sup>10</sup> highlighted the importance of the donor solution composition in the Franz  
62        cell methodology. The donor solution should be standardized since certain components may  
63        promote or inhibit oxidation of metal(loid)s. There is no standardized donor solution for  
64        permeation studies, and to date, a comparison of the influence of the synthetic sweat

65 composition as donor solution in the permeation of metal(loid)s from geological materials is  
66 missing. Recent studies reported that synthetic sweat composition strongly influences the  
67 solubility of metals from geological materials <sup>7-8</sup>. Even though the influence of synthetic  
68 sweat formulation in the solubility of metals has been demonstrated, the impact of synthetic  
69 sweat characteristics when used as donor solution in the dermal permeation of metal(loid)s  
70 from geological materials is not clear.

71 Different membranes to simulate human skin have been used for dermal permeation studies  
72 using the Franz cell methodology. According to the OECD and USEPA guidelines, human  
73 skin is the "gold standard" <sup>6</sup>. Unfortunately, human skin is not always available, can be  
74 expensive, and is highly variable depending on age, sex, origin of the donor, and body area.<sup>13</sup>  
75 Because of these limitations, some studies used animal skin, such as swine, rat, or guinea pig,  
76 as human skin surrogates. <sup>10,14</sup> Animal skin as a surrogate for human skin also has limitations  
77 such as high variability inherent to biological membranes (coefficient of variation = 72%)  
78 rendering the experimental design, and ability to compare results difficult. <sup>13,15</sup> Moreover,  
79 ethical consideration inherent to the use of biological tissues from animal origin and in vivo  
80 experiments with animals must be considered.

81 Artificial membranes used to model skin permeation should mimic the stratum corneum as  
82 close as possible, have low variability and be commercially available. <sup>16</sup> The Strat-M (EMD  
83 Millipore, MA, USA) synthetic membrane is an ultrafiltration membrane made of  
84 polyethersulfone composed by multiple layers, including two layers impregnated with  
85 synthetic lipids and a very tight top layer, producing diffusion results similar to natural human  
86 skin. <sup>16-17</sup> In diffusion studies with human cadaver skin, animal skin, and Strat-M. membrane,  
87 Joshi et al.<sup>17</sup> tested the Strat-M membrane with a mixture of synthetic lipids (to mimic the

88 lipid phase in human skin) in the permeation of nicotine and hydrocortisone. The authors  
89 reported that when the membrane Strat-M was treated with synthetic lipids, it showed a closer  
90 correlation to human skin than the untreated membrane and animal skin. They also exposed  
91 the difficulties of correlating diffusion data between animal skin and human skin <sup>17</sup>. In most  
92 cases, the correlation between the treated synthetic membrane Strat-M and human skin was  
93 better. Moreover, Strat-M membrane showed a high lot-to-lot reproducibility and high shelf  
94 life as opposed to human and animal skin. Yet, to our knowledge, the use of the membrane  
95 Strat-M to assess the dermal permeation of metal(loid)s from contaminated geological  
96 materials has not been reported yet.

97 The dermal bioaccessible fraction of a chemical is the amount that is dissolved in sweat and  
98 is available for penetration through the skin.<sup>5,18</sup> This concentration can be used in conjunction  
99 with the permeation data to calculate bioavailability.<sup>5</sup>

100 To estimate the dermal absorption of contaminants from aqueous media, the USEPA <sup>19</sup>  
101 proposes the water approach methodology. It assumes contact with contaminated water and  
102 aims to calculate the dermally absorbed dose (DAD) using the migration rate of a chemical  
103 through the skin. This migration is characterized by the permeability coefficient  $K_p$  (cm h<sup>-1</sup>).  
104 This coefficient is available in the literature for several inorganics and originates from  
105 experimentally measured or derived values.<sup>7,19</sup> Nevertheless, published  $K_p$  values involve a  
106 high level of uncertainty (since they don't take speciation into account except for chromium)  
107 and they are available for metal(loid)s soluble in water but not soluble in sweat.<sup>19</sup>  
108 Improvement in  $K_p$  determination can reduce uncertainty in the calculation of dermal  
109 exposure therefore refining exposure assessment. To our knowledge, the influence of

110 synthetic sweat formulation on  $K_p$  evaluation to assess the dermal permeation of metal(loid)s  
111 from contaminated soils has not been reported to date.

112 Therefore, the present study aims to (1) assess and compare the dermal bioaccessibility of  
113 As, Cr, Cu, Ni, Pb, and Zn present in various geological materials via *in vitro* experiments  
114 using two artificial sweat formulations; (2) evaluate the diffusion parameters of the  
115 bioaccessible fraction of these metal(loid)s through artificial membrane Strat-M using the  
116 static Franz cell methodology; and (3) investigate the influence of synthetic sebum in the  
117 permeation of these metal(loid)s through artificial membrane Strat-M.

118 **2. Materials & Methods**

119 **2.1 Soil sampling and characterization**

120 Three geological materials have been subjected to *in vitro* bioaccessibility tests to assess the  
121 dermal bioaccessible fraction of As, Cr, Cu, Ni, Pb, and Zn. The certified material SQC001  
122 (lot number LRAC0025, produced by Sigma-Aldrich in accordance with ISO 17034  
123 and ISO/IEC 17025 procedures ( $d < 425 \mu\text{m}$ )), and two field-collected soil samples (S7 and  
124 S8), sampled near Chromated copper arsenate (CCA)-treated utility poles in the Montreal  
125 area (Quebec, Canada). The soils were sampled in a 20 cm radius of the poles and up to 10  
126 cm depth. Coarse material ( $> 2 \text{ cm}$ ) and topsoil vegetation were removed prior to sampling.  
127 The samples were collected using a plastic shovel and stored in zip-lock plastic bags.  
128 Containers and tools were washed with a phosphate-free detergent and soaked overnight in  
129 10% (v v<sup>-1</sup>) HNO<sub>3</sub> and rinsed with deionized water (18.2 MΩ.cm) prior to use.<sup>2-3</sup> Field-  
130 collected soil samples were air-dried, gently disaggregated using a mortar and dry sieved to  
131 420  $\mu\text{m}$  using a sieve shaker (Retsch AS-200). Samples were then stored at 4°C.

132 Total metal(loid) content in soil samples was determined via acid digestion on a hot plate  
133 using  $\text{HNO}_3$  (70 % w/w), HF (50 % w/w), and  $\text{HClO}_4$  (70 % w/w) according to standard  
134 method 3030.<sup>20</sup> Digestates were transferred to 100-ml volumetric flasks and made up to  
135 volume with deionized water. Solutions were filtered (0.45 $\mu\text{m}$ ) with glass microfiber filters  
136 (Whatman) and stored in polypropylene centrifuge tubes with HDPE screw caps. Cr, Cu, Ni,  
137 Pb, and Zn concentrations were measured via atomic absorption spectroscopy (AAS)  
138 (Perkin- Elmer A200). Detection limits (DLs, determined based on signal-to-noise approach  
139 (ratio of 3:1)) in  $\text{mg kg}^{-1}$  were 0.3, 0.3, 0.2, 1, and 0.1, respectively. Arsenic content was  
140 determined via ICP-OES (Varian Vista), with a detection limit (DL) of 0.004  $\text{mg kg}^{-1}$ . Soil  
141 pH was measured in duplicates in solid-to-liquid ratio 1:2 with deionized water (pH meter:  
142 Eutech pH 200 series, probe: Accumet Ag/AgCl) according to method ASTM D4972-13.<sup>21</sup>  
143 Total organic carbon content with a detection limit (DL) of 0.1% w/w was analyzed using a  
144 LECO furnace. Infrared determination of  $\text{CO}_2$  was achieved to determine organic carbon  
145 content as a difference between total and inorganic carbon.<sup>22</sup> Cation exchange capacity  
146 (CEC) was determined using the sodium acetate method with  $\text{NaOAc}$  1N and  $\text{NH}_4\text{OAc}$  1N.<sup>23</sup>

## 147       **2.2 Artificial SSFLs**

148 Three solutions have been prepared to mimic human skin surface film liquids (SSFL): two  
149 synthetic sweats (Sweat A (pH = 6.5) and Sweat B (pH = 4.7)) and one synthetic sebum  
150 (Table 1). The SSFL formulations are further described elsewhere<sup>8</sup> and have been selected  
151 for their differences in pH and composition. Sweats A and B simulate the sweat layer on the  
152 skin, while sebum was used to treat the synthetic membranes to mimic the hydrophobic  
153 properties of the skin, caused by the presence of lipids.

154 **Table 1**

## 155 Composition of artificial SSFLs

Chemical, % (w/w)	Sweat A <sup>a</sup> (pH = 6.5)	Sweat B <sup>b</sup> (pH = 4.7)	Sebum <sup>c</sup>
Deionized water	99.3	94	-
Sodium chloride	0.5	2	-
Lactic acid	0.1	1.5	-
Urea	0.1	0.5	-
Acetic acid	-	0.25	-
Ammonium chloride	-	1.75	-
Squalene	-	-	12.4
Jojoba oil	-	-	25
Triolein	-	-	44.6
Oleic acid	-	-	17
Vitamin E	-	-	1

156

157 <sup>a</sup> According to standard EN 1811 <sup>24</sup>158 <sup>b</sup> According to standard NIHS 96-10 <sup>25</sup>159 <sup>c</sup> According to Wertz <sup>26</sup>

160

161 **2.3 *In vitro* dermal bioaccessibility test**

162 Tests were started by adding 20 ml of synthetic sweat to 2 g of soil sample ( $d < 425 \mu\text{m}$ ) in  
 163 50 ml Polypropylene tubes with HDPE caps. Tubes containing the soil and sweat mixture  
 164 were placed on an orbital shaker (Cole-Parmer 51704 Series, radius 9.5 mm) at 100 rpm  
 165 inside an incubator (Isotemp, Fisher Scientific) at 36°C, corresponding to the median skin  
 166 temperature for humans <sup>5</sup>, for 2 hours. The tubes were then centrifuged (Heraeus Megafuge  
 167 8, Thermo Fisher) at 10,000 x g for 10 minutes, and the supernatant collected with 60 ml  
 168 Luer-Lok syringes and filtered with a 0.45  $\mu\text{m}$  PVDF filter fitted to the syringe. The filtered  
 169 supernatant was transferred into 50-ml Polypropylene centrifuge tubes with HDPE caps and  
 170 stored at 4°C until analysis. Samples were analyzed via ICP-AES (Vista, Varian Inc.) to  
 171 determine As, Cr, Cu, Ni, Pb, and Zn concentrations (with detection limits in  $\text{mg kg}^{-1}$  of

172 0.004, 0.001, 0.006, 0.001, 0.006, and 0.024, respectively). For each metal(loid),  
173 bioaccessibility percentage (%<sub>bio</sub>) was determined as follows:

174 
$$\%_{bio} = \frac{C_{bio}}{C_{total}} * 100 \quad (1)$$

175 Where C<sub>bio</sub> is the bioaccessible concentration of metal(loid) (mg kg<sup>-1</sup>), and C<sub>total</sub> is the total  
176 concentration of metal(loid) in the soil sample (mg kg<sup>-1</sup>). A more detailed description of the  
177 dermal bioaccessibility protocol can be found elsewhere.<sup>8</sup>

178 **2.4 Permeation test**

179 Glass jacketed vertical Franz diffusion cells (PermeGear Inc.) with a 9 mm orifice diameter,  
180 5 ml receptor volume, and 1 ml donor volume were used. The temperature of the receptor  
181 was maintained at 37 °C by circulating water from a water bath (Model 2849, Thermo Fisher  
182 Scientific) to simulate temperature below the skin.<sup>10, 27</sup> The receptor compartment was filled  
183 with Phosphate-Buffered Saline (PBS) solution (Fisher Scientific) at a pH of 7.4<sup>13</sup> and NaCl  
184 8.0 g/L, KCl 0.2 g/L, Na<sub>2</sub>HPO<sub>4</sub> 1.44 g/L, and KH<sub>2</sub>PO<sub>4</sub> 0.24 g/L, to represent blood salt  
185 concentration and blood pH.<sup>10</sup> The receptor compartment was subjected to constant stirring  
186 (300 rpm) with a magnetic stirrer (Poly 15, Variomag). Twenty-five mm OD sterile Strat-M  
187 membranes (EMD Millipore) were used as a surrogate for human skin. Each membrane was  
188 mounted on the Franz diffusion cell with the shiny side in contact with the donor  
189 compartment.<sup>13</sup> To simulate the hydrophobic character of the skin and investigate the  
190 influence of the lipid fraction of SSFL in the permeation of metal(loid)s, half of the  
191 membranes were coated with 0.1 ml of sebum.

192 One ml of filtered supernatant collected from the bioaccessibility test (donor solution) was  
193 immediately added to the open-top donor compartment to start the permeation experiment.

194 Tests were performed in duplicate and in the presence of procedure blanks (fresh synthetic  
195 sweat as donor solution). The receptor solution was completely removed from the receptor  
196 compartment at 6, 8, 10, 12, 14, and 24 h and placed in 15 ml Polypropylene centrifuge tubes.  
197 After each sampling event, the receptor compartment was rinsed with 5 ml of fresh PBS  
198 solution using a syringe. This rinsing solution was added to the tube containing the receptor  
199 solution. After every sampling event followed by rinsing, 5 ml of fresh PBS solution was  
200 added to the receptor compartment. The procedure blank cells received the same treatment.  
201 The receptor and donor solution samples were kept refrigerated at 4°C until analyzed via  
202 ICP-AES (Vista, Varian Inc.) to determine As, Cr, Cu, Ni, Pb and Zn concentrations.  
203 As an additional quality control measure, a mass balance was performed for one sample per  
204 batch. At the end of the permeation test (after 24 h), the donor compartment and the  
205 membrane were rinsed four times with 1 ml of deionized water. The rinsing solution was  
206 added to the remaining donor solution for analysis. To assess the amount of metal that  
207 remained in the membrane, membranes were placed in Teflon beakers and digested on a hot  
208 plate with HNO<sub>3</sub>, HCl, and HF for 45 minutes and then filtered with 0.45 µm filters  
209 (Whatman), diluted to 100 ml and analyzed for total metal content via ICP-AES. Mass  
210 balance was calculated by comparing the metal recovered from the donor solution,  
211 membrane, and receptor solution at the end of the experiment with the initial mass of metal  
212 present in the donor solution.

213 **3. Results and Discussion**

214 **3.1 Soil Characterization**

215 The pH was neutral to slightly alkaline for CCA-contaminated soil samples S7 and S8 and  
216 acidic for reference material SQC001 (Table 2). Low total organic carbon content (<2.9%)

217 was reported in all soil samples. CEC values ranged from 12.8 to 41.4 meq 100 g<sup>-1</sup>. High  
218 CEC suggests that cationic metals such as Cu, Ni, and Zn could be retained by cation  
219 exchange on the soil.<sup>2</sup> Total metal(loid)s concentrations in soil samples S7 and S8 are also  
220 shown in Table 2. Values in bold are exceeding Quebec's regulatory limit for industrial land-  
221 use (C criterion).<sup>28</sup> Extensive As contamination was observed for S8 (1639 mg kg<sup>-1</sup> ± 6.8 %)  
222 (more than 30 times the C criterion) and S7 (311 mg kg<sup>-1</sup> ± 1.1 %). Cu content exceeding the  
223 C criterion (500 mg kg<sup>-1</sup>) was also observed for S7 (824 mg kg<sup>-1</sup> ± 5.8 %) and S8 (1070 mg  
224 kg<sup>-1</sup> ± 11.0 %). As previously reported, Cr contamination was less problematic than As and  
225 Cu<sup>2</sup> but soil samples S7 and S8 still contained elevated Cr concentrations. Certified reference  
226 material SQC001 had a lower content of As, Cr, and Cu but a higher content of Pb, and Zn.  
227 The measured total metal(loid) content of SQC001 was within 100 ± 10% of the certified  
228 values provided in the certificate of analysis.

229

230 **Table 2**

231 Total concentrations of As, Cr, Cu, Ni, Pb, and Zn (mg kg<sup>-1</sup>), pH, total organic carbon (TOC,  
232 w/w %), and cation exchange capacity (CEC, meq 100 g<sup>-1</sup>) of soils. Precision is expressed  
233 as mean ± relative standard deviation %.

Parameter	SQC 001	S7	S8
As	<b>173 ± 20</b>	<b>311 ± 1.1</b>	<b>1639 ± 6.8</b>
Cr	124 ± 6.5	371 ± 3.9	582 ± 14.3
Cu	82 ± 1.9	<b>824 ± 5.8</b>	<b>1070 ± 11.0</b>
Ni	112 ± 9.8	26 ± 11.5	223 ± 11.3
Pb	263 ± 2.5	57 ± 8.8	80 ± 10.6
Zn	512 ± 4.0	261 ± 7.3	223 ± 10.8
pH	5.7 ± 0.2	7.2 ± 0.7	7.1 ± 0.5
TOC	< 0.01	1.3 ± 2.1	2.9 ± 1.5
CEC	12.8 ± 8.8	15.5 ± 16.3	41.4 ± 2.8

234

235

236        **3.2 Dermal bioaccessibility of metal(loid)s**

237        Table 3 shows the bioaccessibility (%) and bioaccessible concentration (mg L<sup>-1</sup>) of  
238        metal(loid)s obtained following bioaccessibility tests performed on S7, S8 and SQC001 using  
239        Sweat A and Sweat B. In agreement with the findings of Marin Villegas et al.<sup>8</sup>, Sweat B  
240        generally yielded the highest bioaccessibility percentage values. The difference in the  
241        bioaccessibility of metal(loid)s for the different sweat compositions can be attributed to the  
242        fact that lower pH increases the solubility, particularly for cationic metals.

243        The reference material SQC001 revealed an overall higher dermal bioaccessibility  
244        percentage compared to S7 and S8, especially for Cu, Ni, Pb, and Zn. In sweat B, which  
245        yielded higher bioaccessibility percentages than Sweat A due to its more acidic pH (4.7),  
246        metal(loid) dermal bioaccessibility ranged from 5.1 to 91.0% in SQC001 but it remained  
247        lower than 7.2% in soil S9 and lower than 2.9% in soil S8 (Table 3). The overall higher  
248        bioaccessibility percentage measured in the reference material can be explained by its lower  
249        pH, low total organic carbon content (< 0.1% w/w) and the fact that the reference material  
250        SQC001 has not undergone a natural aging process, which reduces the bioaccessibility of  
251        metals in soils.<sup>29</sup>

252

253

254

255

256

257 **Table 3**

258 Bioaccessibility of As, Cr, Cu, Ni, Pb, and Zn in soils using synthetic sweat formulations A  
 259 and B : (a) bioaccessibility expressed in percentage (%) and (b) bioaccessible concentration  
 260 (mg l<sup>-1</sup>)

(a) Bioaccessibility (%)							
Donor solution	Soil	As	Cr	Cu	Ni	Pb	Zn
Sweat A (pH=6.5)	S7	1.4	0.15	0.5	< 0.4	< 0.2	0.05
	S8	0.5	0.06	0.4	< 0.4	< 0.2	0.02
	SQC001	0.06	0.13	2.1	20.4	0.4	40.3
Sweat B (pH=4.7)	S7	2.8	2.0	7.2	0.5	< 0.2	9.0
	S8	0.8	1.0	2.9	0.8	< 0.2	3.0
	SQC001	5.1	25.1	61.3	44.9	34.9	91.0

(b) Bioaccessible concentration (mg l <sup>-1</sup> )							
Donor solution	Soil	As	Cr	Cu	Ni	Pb	Zn
Sweat A (pH=6.5)	S7	0.45	0.05	0.38	< 0.01	< 0.01	0.01
	S8	0.76	0.04	0.41	< 0.01	< 0.01	0.01
	SQC001	0.01	0.02	0.17	2.29	0.09	20.6
Sweat B (pH=4.7)	S7	0.86	0.76	5.93	0.01	< 0.01	2.34
	S8	1.36	0.59	3.13	0.02	< 0.01	0.66
	SQC001	0.88	3.12	5.07	5.02	9.20	46.6

261

262

263 In summary, bioaccessibility percentages in terms of sweat are in the following order Sweat  
 264 B > Sweat A and in terms of geological materials they are SQC001 > S7 > S8. In agreement  
 265 with bioaccessibility percentages, bioaccessible concentrations were relatively low (below 1  
 266 mg l<sup>-1</sup>) in sweat A except for Ni (2.29 mg l<sup>-1</sup>) and Zn (20.6 mg l<sup>-1</sup>). The highest bioaccessible  
 267 concentrations were measured in SQC001 for Ni, Pb, and Zn. Even though Cu  
 268 bioaccessibility percentages following extraction with the more acidic sweat B were only 7.2  
 269 and 2.9 % in soils S7 and S8 respectively, high bioaccessible concentrations (5.9 and 3.1 mg  
 270 l<sup>-1</sup> respectively) were obtained because of the very high Cu content (largely exceeding the C  
 271 criterion) measured in these field-collected soil samples (Table 2).

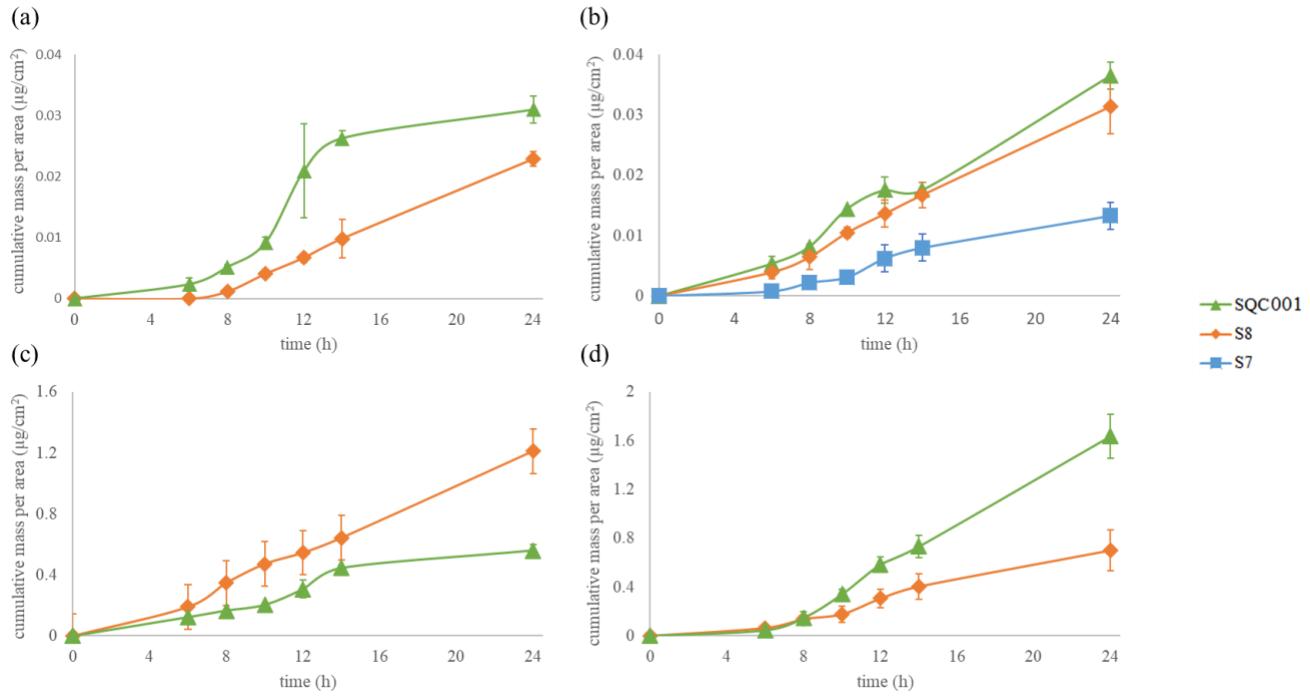
272 A more comprehensive analysis of the influence of soil properties and synthetic sweat  
273 formulation on bioaccessibility of metal(loid)s from soils can be found in Leal et al.<sup>7</sup>, and  
274 Marin Villegas et al.<sup>8</sup>

275 **3.3 Permeation test**

276 Following the bioaccessibility test, the supernatant containing the bioaccessible fraction of  
277 metal(loid)s was used as the donor solution in permeation tests. The concentration of  
278 metal(loid)s in the receptor ( $\mu\text{g l}^{-1}$ ) was converted to the total metal(loid) amount that  
279 permeated ( $\mu\text{g.cm}^{-2}$ ) and then plotted against time (Figure 1 and Figure 2). Flux ( $\mu\text{g cm}^{-2} \text{h}^{-1}$ )  
280 was calculated as the slope from the steady-state region of graphs shown on Figures 1 and  
281 2 and lag time as the intercept of the curve with the X-axis (time). The permeation coefficient  
282  $K_p$  was calculated from the linear steady-state region of the plot by dividing the flux through  
283 the membrane ( $\mu\text{g cm}^{-2} \text{h}^{-1}$ ) by the concentration in the donor solution.<sup>30</sup>

284 Overall, mass balance recovery calculated from donor solutions, synthetic membranes, and  
285 receptor solutions yielded satisfactory percentages. Results for all analyzed samples ranged  
286 from 58.1 to 101.3 % for Cr and from 85.3 to 128.4 % for Cu (when Sweat A was used as a  
287 donor solution) and from 81.9 to 115.2 % for Cr and 73.4 to 101.2 % for Cu when Sweat B  
288 was used as a donor solution.

289



290

291 **Figure 1**

292 Mean cumulative mass per area and standard deviation of metal in Sweat A (pH = 6.5) that  
 293 permeated through Strat-M membrane for (a) chromium (not coated with sebum), (b)  
 294 chromium (coated with sebum), (c) copper (not coated with sebum), and (d) copper (coated  
 295 with sebum).

296

297

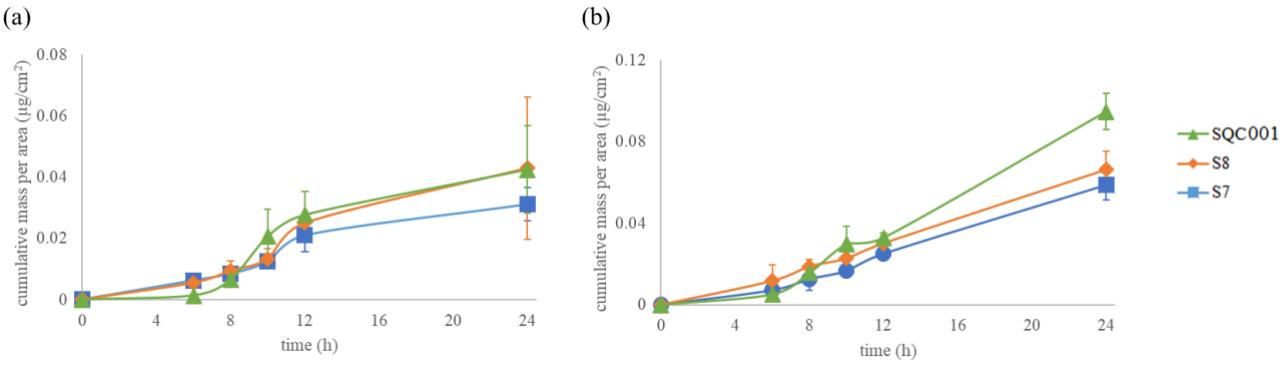
298

299

300

301

302



303

304 **Figure 2**

305 Mean cumulative mass per area and standard deviation of chromium in Sweat B (pH = 4.7)  
 306 that permeated through Strat-M membrane when (a) not coated with sebum and (b)  
 307 with sebum.

308

309 **3.3.1 Arsenic**

310 Arsenic did not permeate the synthetic skin membrane for all sweat formulation and soil  
 311 sample tested. At the pH of both sweats and under oxidizing conditions, As is expected to be  
 312 found as As(V), the less mobile form of this metalloid.<sup>31-32</sup> Experimental data indicate that  
 313 penetration of metal(loid)s through the skin is significantly dependent on ion mobility and  
 314 charge.<sup>33</sup>

315 **3.3.2 Chromium**

316 Figure 1 and Figure 2 show that Cr permeated through the membrane for both synthetic sweat  
 317 formulations and all geological materials (except for soil S7 in Sweat A, in the absence of  
 318 sebum). Results (Table 4) indicate that  $K_p$  is higher for Sweat A than Sweat B for the same  
 319 geological material. This can be explained by the Cr (III) and Cr(VI) species in each sweat  
 320 formulation. In the more alkaline Sweat A, and under oxidizing conditions (donor  
 321 compartment open to air), a higher amount of Cr is expected to be found as Cr(VI) when

322 compared to the more acidic Sweat B. Cr(VI) has a higher potential of permeation than  
323 Cr(III) and a higher associated  $K_p$ <sup>10, 34, 35</sup>, which explains the higher Flux and  $K_p$  of Cr in  
324 Sweat A.

325 For Sweat B, flux and  $K_p$  are roughly doubled in the presence of sebum for all the geological  
326 materials. Sebum might interact with Cr species in sweat B and foment production of more  
327 penetrable compounds. In Sweat A, in addition to Cr(VI), Cr(III) species that are expected to  
328 be found are mostly  $\text{Cr(OH)}_2^+$  and neutral thus potentially more permeable  $\text{Cr(OH)}_3(\text{aq})$ ,  
329 whereas in Sweat B due to a more acidic pH, Cr(III) is expected to be the main Cr oxidation  
330 state.<sup>36</sup> This hypothesis is backed by the fact that for Sweat A, there is no visible influence  
331 of sebum in the permeation parameters. However, other variables such as molecular size,  
332 chemical reactivity, and counterions could affect permeation of Cr through the skin  
333 membrane<sup>33, 37</sup> To validate this hypothesis, it is recommended to assess chromium speciation  
334 in SSFLs.

335 The permeation coefficient ( $K_p$ ) also varies among geological materials within the same  
336 sweat composition. This might be partially explained by the dose factor: the rate of diffusion  
337 of Cr is not proportionate to the applied concentration, absolute absorption can reach a  
338 plateau value and then decrease with an increment in concentration.<sup>37</sup> This phenomenon  
339 happens in real skin due to the buildup of a secondary diffusion barrier as a consequence of  
340 electrophilic metals (such as Cr(III)) forming stable bonds with proteins in the skin, also  
341 causing longer lag times.<sup>37</sup> As previously reported,<sup>8</sup> the concentrations of metals in the  
342 donor solution following dermal bioaccessibility tests are dependent on the physicochemical  
343 properties of the geological material and extractant (synthetic sweat). There was a much

344 higher bioaccessible concentration of Cr in Sweat B than Sweat A following bioaccessibility  
 345 test (Table 3) producing different diffusion profiles (Figure 1 and Figure 2).

346 **Table 4**

347 Flux ( $\text{g cm}^{-2} \text{ h}^{-1}$ ), lag time (h), and permeation coefficient ( $\text{cm h}^{-1}$ ) for (a) Cr and (b) Cu (data  
 348 expressed as mean  $\pm$  standard deviation).

		(a)							
		Donor solution	Flux ( $\text{g cm}^{-2} \text{ h}^{-1}$ )	Lag Time (h)		Kp ( $\text{cm h}^{-1}$ )			
Sweat A (pH=6.5)	S7		-		-		-		
	S8	1.4E-03	$\pm$ 4.0E-05	6.7	$\pm$ 0.3	3.8E-02	$\pm$ 1.1E-03		
	SQC001	1.7E-03	$\pm$ 4.1E-04	2.8	$\pm$ 0.6	1.1E-01	$\pm$ 2.6E-02		
	S7 Sebum	7.1E-04	$\pm$ 6.5E-05	4.4	$\pm$ 0.1	1.3E-01	$\pm$ 1.2E-02		
	S8 Sebum	1.5E-03	$\pm$ 3.4E-05	5.5	$\pm$ 0.5	4.7E-02	$\pm$ 1.0E-03		
	SQC001 Sebum	1.7E-03	$\pm$ 1.1E-04	2.5	$\pm$ 0.2	1.1E-01	$\pm$ 7.2E-03		
Sweat B (pH=4.7)	S7	1.4E-03	$\pm$ 2.4E-04	0.6	$\pm$ 0.3	1.5E-03	$\pm$ 2.6E-04		
	S8	2.1E-03	$\pm$ 2.6E-04	2.8	$\pm$ 0.4	2.6E-03	$\pm$ 3.3E-04		
	SQC001	2.2E-03	$\pm$ 5.0E-04	2.9	$\pm$ 0.7	7.3E-04	$\pm$ 1.7E-04		
	S7 Sebum	2.9E-03	$\pm$ 2.9E-03	3.7	$\pm$ 0.1	3.2E-03	$\pm$ 3.2E-03		
	S8 Sebum	3.0E-03	$\pm$ 6.3E-05	1.9	$\pm$ 0.1	3.7E-03	$\pm$ 7.8E-05		
	SQC001 Sebum	4.9E-03	$\pm$ 2.0E-04	4.7	$\pm$ 0.3	1.9E-03	$\pm$ 7.7E-05		

		(b)							
		Donor solution	Flux ( $\text{g cm}^{-2} \text{ h}^{-1}$ )	Lag Time (h)		Kp ( $\text{cm h}^{-1}$ )			
Sweat A (pH=6.5)	S7		-		-		-		
	S8	5.5E-02	$\pm$ 1.8E-03	0.2	$\pm$ 0.4	1.3E-01	$\pm$ 4.3E-03		
	SQC001	2.6E-02	$\pm$ 4.2E-03	0.6	$\pm$ 0.6	1.5E-01	$\pm$ 2.4E-02		
	S7 Sebum	-		-	-	-	-		
	S8 Sebum	3.6E-02	$\pm$ 2.3E-03	4.6	$\pm$ 0.3	8.8E-02	$\pm$ 5.5E-03		
	SQC001 Sebum	9.1E-02	$\pm$ 2.5E-03	5.9	$\pm$ 0.3	5.2E-01	$\pm$ 1.4E-02		

349

350

351 **3.3.3 Copper**

352 Copper extracted from samples S8 ( $0.41 \text{ mg l}^{-1}$ ) and SQC001 ( $0.17 \text{ mg l}^{-1}$ ) with Sweat A  
 353 permeated the synthetic skin membrane, both in the presence and absence of sebum ((Figure  
 354 1 and Table 4). It should be noted that Cu was also detected in the receptor solution following

355 permeation with sample S7 extracted with Sweat A ( $0.38 \text{ mg l}^{-1}$ ) in the presence of sebum at  
356 the end of the sampling period ( $t=12\text{h}$  and  $t=24\text{h}$ ). However, there was not enough data to  
357 build the cumulative mass per area versus time curve and calculate the permeation  
358 parameters. Copper did not permeate the membrane when extracted with sweat B even  
359 though bioaccessible concentrations (donor solution) were much higher ( $3.13 \text{ mg l}^{-1}$  for S8  
360 and  $5.07 \text{ mg l}^{-1}$  for SQC001). This suggests that Cu speciation and complexation, which  
361 depends on pH, and sweat formulation among other parameters, might influence the  
362 permeation of Cu.

363 The  $K_p$  for Cu was similar for soils S8 and SQC001 in the absence of sebum. However, in  
364 the presence of sebum, the  $K_p$  was around five times higher for SQC001. Lag times before  
365 Cu permeation were longer in the presence of sebum for S8 and SQC001, increasing from  
366 0.2 to 4.6 h and from 0.6 to 5.9 h, respectively. This could be caused by the added layer  
367 formed by the sebum, causing the Cu flux to take a longer time to reach equilibrium.

### 368       **3.3.4 Nickel**

369 High bioaccessible Ni concentrations ( $2.29\text{-}5.02 \text{ mg l}^{-1}$ ) were obtained when extracting  
370 SQC001 with both sweat formulations but low to below detection ( $< 0.01 \text{ mg l}^{-1}$ )  
371 bioaccessible Ni concentrations were measured when extracting field-collected soil samples  
372 S7 and S8 (Table 3). In all cases, Ni did not permeate the synthetic skin membrane.  
373 Depending on Ni activity, Ni in solutions with a  $\text{pH} < 8$  is expected to be mostly found as  
374  $\text{Ni}^{2+}$ .<sup>38</sup> Fullerton et al.<sup>39</sup> reported a strong influence of the vehicle in the permeation of  $\text{NiCl}_2$   
375 and  $\text{NiSO}_4$  through the skin and lag times of 50 h (our permeation test lasted 24 h). In another  
376 study, Larese et al.<sup>30</sup>, tested permeation through human abdominal skin of Ni powder in  
377 suspension in synthetic sweat ( $50 \text{ g l}^{-1}$ ) at  $\text{pH} 6.5$  (Sweat A). In the latter study, Ni slowly

378 permeated the skin with a lag phase of 14 h and Ni was present as a free ion in the donor and  
379 in the receptor solution. Nevertheless, in the present study, the potential for very long lag  
380 times made Ni unable to penetrate the synthetic skin membrane under the experimental  
381 design used.

382 **3.3.5 Lead**

383 Lead was only present in donor solutions when SQC001 was extracted with Sweat A (0.09  
384 mg l<sup>-1</sup>) and Sweat B (9.20 mg l<sup>-1</sup>) (Table 3). However, it did not permeate the synthetic skin  
385 membrane. In real human skin, Pb is mainly absorbed through the sweat glands and hair  
386 follicles and only slightly mobile through the transepidermal route depending on its  
387 speciation.<sup>10, 40</sup> The Strat-M membrane is designed to specifically simulate the diffusion  
388 pathway.

389 **3.3.6 Zinc**

390 Zinc did not permeate the synthetic skin membrane even if very high bioaccessible  
391 concentrations (Table 3) were found in the donor solution for both SSFLs with SQC001. Zinc  
392 is expected to be mostly present in its charged ionic form Zn<sup>2+</sup> from low to neutral pH.<sup>41</sup>  
393 Most of the previous investigations regarding permeation of Zn were focused on ZnO from  
394 sunscreen, concluding that particles formed micron-sized aggregates reducing permeation  
395 through human skin.<sup>10,42,43</sup>

396 **3.3.7 Comparison of K<sub>p</sub> values with published values**

397 The USEPA<sup>19</sup> recommends K<sub>p</sub> values (cm hr<sup>-1</sup>) for some metals (Cr(III): 0.001, Cr(VI):  
398 0.002, Ni:0.0002, Pb:0.0001, Zn: 0.0006, and other non-specified inorganics: 0.001). These  
399 values have been adapted from Hostýnek et al.<sup>44</sup> Because of its conservative approach,

400 USEPA<sup>19</sup> listed the highest reported permeability coefficient. Nonetheless, other studies  
401 reported higher permeability coefficients than the ones recommended by USEPA.<sup>19</sup>  
402 Examples of these are Filon et al.<sup>45</sup>, who reported a  $K_p$  for Cr of 0.0124, and Fullerton et al.<sup>39</sup>,  
403 who reported  $K_p$  for Ni of 0.0015 for epidermis and 0.23 for the dermis.  
404 The  $K_p$  from our experiments using the Strat-M synthetic membrane when metal(loid)s  
405 extracted in sweat A were used in the donor solution were higher than the ones summarized  
406 by USEPA.<sup>19</sup> However the values were in the same order of magnitude or slightly higher for  
407 Cr when sweat B was used in the donor solution. Moreover, in the present study, longer lag  
408 times and differences in the  $K_p$  were found in the presence of sebum. The percutaneous data  
409 obtained using a synthetic skin membrane in this study gives valuable insights regarding the  
410 influence of the donor solution pH and the presence of sebum.

411 **4. Future research and study limitations**

412 The dermal bioaccessibility of all studied metal(loid)s from geological materials was higher  
413 at lower pH. Nevertheless, only Cr and Cu could permeate the Strat-M membrane following  
414 bioaccessibility test in synthetic sweat, and Cr and Cu seem to have a greater potential for  
415 diffusion through human skin surrogate (Strat-M membrane) at higher pH. For this reason,  
416 the characteristics of the donor solution (synthetic sweat formulation) are critical for both  
417 bioaccessibility and permeation of metal(loid)s. Further studies are needed to assess  
418 metal(loid) speciation in various synthetic SSFLs to help explain differences in bioaccessible  
419 metal(loid)s ability to permeate the skin. Moreover, the present study findings also warrant  
420 additional studies on the influence of sebum in the permeation of metal(loid)s from various  
421 soil types through human skin surrogates.

422 The diffusion of metal(loid)s through a barrier is a complex phenomenon because several  
423 factors are interrelated, such as pH, oxidation state, presence of counter ions, dose, and  
424 solubility. For risk assessment, environmental agencies like USEPA often suggest  
425 standardized or generic values for  $K_p$ . However,  $K_p$  values are not only metal specific but  
426 appear to be site-specific and depend on several variables that must be considered for a more  
427 accurate estimation of risk related with dermal exposure.

428 The results obtained in this study showed that membrane Strat-M is suited for early stages of  
429 dermal permeation studies of metal(loid)s from contaminated geological materials. Synthetic  
430 membranes are commonly available, produce less variability, and are significantly less  
431 expensive than human skin. However, the Strat-M membrane only models the diffusion of  
432 chemicals and is not suitable to model other dermal pathways such as appendages (through  
433 sweat glands and hair follicles). An additional shortcoming of the Strat-M membrane is that  
434 it does probably not allow to simulate reservoirs of metal(oids) in the stratum corneum and  
435 other layers of the skin. Further research to compare diffusion results with real human skin  
436 is therefore necessary to reduce uncertainty in the calculation of dermal exposure to  
437 metal(loid)s present in contaminated soils.

438

#### 439 **Acknowledgment**

440 The corresponding author acknowledges the financial support from the Natural Sciences  
441 and Engineering Research Council of Canada (NSERC) obtained via the Discovery Grant  
442 Program (Application Number: RGPIN-2016-06430). The authors also acknowledge the  
443 technical support provided by Jérôme Leroy and Lan Huong Tran. The authors declare no  
444 competing financial interest.

445 **List of Tables**

446 **Table 1** Composition of artificial SSFLs.

447 **Table 2** Total concentrations of As, Cr, Cu, Ni, Pb, and Zn ( $\text{mg kg}^{-1}$ ), pH, total organic  
448 carbon (TOC, w/w %), and cation exchange capacity (CEC, meq  $100 \text{ g}^{-1}$ ) of soils. Precision  
449 expressed as  $\pm$  relative standard deviation % .

450 **Table 3** Bioaccessibility of As, Cr, Cu, Ni, Pb, and Zn in soils using synthetic sweat  
451 formulations A, and B : (a) bioaccessibility expressed in percentage (%) and (b) bioaccessible  
452 concentration ( $\text{mg l}^{-1}$ ).

453 **Table 4** Flux ( $\text{g cm}^{-2} \text{ h}^{-1}$ ), lag time (h), and permeation coefficient ( $\text{cm h}^{-1}$ ) for (a) Cr and (b)  
454 Cu (data expressed as mean  $\pm$  standard deviation).

455

456

457 **List of Figures**

458 **Figure 1** Mean cumulative mass per area and standard deviation of metal in Sweat A (pH =  
459 6.5) that permeated through Strat-M membrane for (a) chromium (not coated with sebum),  
460 (b) chromium (coated with sebum), (c) copper (not coated with sebum), and (d) copper  
461 (coated with sebum).

462 **Figure 2** Mean cumulative mass per area and standard deviation of chromium in Sweat B  
463 (pH = 4.7) that permeated through Strat-M membrane when (a) not coated with sebum and  
464 (b) coated with sebum.

465

## 5. References

467 (1) Guney, M., Bourges, C. M. J., Chapuis, R. P., & Zagury, G. J. (2017). Lung  
468 bioaccessibility of As, Cu, Fe, Mn, Ni, Pb, and Zn in fine fraction (< 20  $\mu\text{m}$ ) from  
469 contaminated soils and mine tailings. *Science of the Total Environment*, 579, 378-386.

470 (2) Gosselin, M., & Zagury, G. J. (2020). Metal(loid)s inhalation bioaccessibility and  
471 oxidative potential of particulate matter from chromated copper arsenate (CCA)-  
472 contaminated soils. *Chemosphere*, 238, 124557. doi:10.1016/j.chemosphere.2019.124557

473 (3) Van der Kallen, C. C., Gosselin, M., & Zagury, G. J. (2020). Oral and inhalation  
474 bioaccessibility of metal(loid)s in chromated copper arsenate (CCA)-contaminated soils:  
475 Assessment of particle size influence. *Science of The Total Environment*, 734, 139412.  
476 doi:<https://doi.org/10.1016/j.scitotenv.2020.139412>

477 (4) Kastury, F., Karna, R. R., Scheckel, K. G., & Juhasz, A. L. (2020). Correlation between  
478 lead speciation and inhalation bioaccessibility using two different simulated lung fluids.  
479 *Environmental Pollution*, 263, 114609. doi:<https://doi.org/10.1016/j.envpol.2020.114609>

480 (5) Stefaniak, A. B., Duling, M. G., Geer, L., & Virji, M. A. (2014). Dissolution of the metal  
481 sensitizers Ni, Be, Cr in artificial sweat to improve estimates of dermal bioaccessibility.  
482 *Environmental Science Processes & Impacts*, 16, 341-351. doi:10.1039/c3em00570d

483 (6) Beriro, D. J., Cave, M. R., Wragg, J., Thomas, R., Wills, G., & Evans, F. (2016). A review  
484 of the current state of the art of physiologically-based tests for measuring human dermal in  
485 vitro bioavailability of polycyclic aromatic hydrocarbons (PAH) in soil. *Journal of*  
486 *Hazardous Materials*, 305, 240-259.

487 (7) Leal, L. T. C., Guney, M., & Zagury, G. J. (2018). In vitro dermal bioaccessibility of  
488 selected metals in contaminated soil and mine tailings and human health risk characterization.  
489 Chemosphere, 197, 42-49. doi:10.1016/j.chemosphere.2018.01.008

490 (8) Marin Villegas, C. A., Guney, M., & Zagury, G. J. (2019). Comparison of five artificial  
491 skin surface film liquids for assessing dermal bioaccessibility of metals in certified reference  
492 soils. Science of The Total Environment, 692, 595-601.

493 (9) Sartorelli, P., Andersen, H.R., Angerer, J., Corish, J., Drexler, H., Göen, T., Griffin, P.,  
494 Hotchkiss, S.A.M., Larese, F., Montomoli, L., Perkins, J., Schmelz, M., van de Sandt, J., &  
495 Williams, J. (2000). Percutaneous penetration studies for risk assessment. Environmental  
496 Toxicology and Pharmacology, 8, 133-152.

497 (10) Franken, A., Eloff, F. C., Du Plessis, J., & Du Plessis, J. L. (2015a). In vitro permeation  
498 of metals through human skin: a review and recommendations. Chemical Research in  
499 Toxicology, 28, 2237-2249.

500 (11) Sartorelli, P., Montomoli, L., & Sisinni, A. (2012). Percutaneous penetration of metals  
501 and their effects on skin. Prevention & Research, 2, 158-164.

502 (12) Franz, T. J. (1975). Percutaneous absorption. On the relevance of in vitro data. Journal  
503 of Investigative Dermatology, 64, 190-195.

504 (13) Haq, A., Dorrani, M., Goodyear, B., Joshi, V., & Michniak-Kohn, B. (2018a).  
505 Membrane properties for permeability testing: Skin versus synthetic membranes.  
506 International Journal of Pharmaceutics, 539, 58-64. doi:10.1016/j.ijpharm.2018.01.029

507 (14) Moody, R. P., Nadeau, B., & Chu, I. (1995). In vivo and in vitro dermal absorption of  
508 benzo [a] pyrene in rat, guinea pig, human and tissue-cultured skin. *Journal of*  
509 *Dermatological Science*, 9, 48-58.

510 (15) Barbero, A. M., & Frasch, H. F. (2009). Pig and guinea pig skin as surrogates for human  
511 in vitro penetration studies: a quantitative review. *Toxicology in Vitro*, 23, 1-13.

512 (16) Haq, A., Goodyear, B., Ameen, D., Joshi, V., & Michniak-Kohn, B. (2018b). Strat-M(R)  
513 synthetic membrane: Permeability comparison to human cadaver skin. *International Journal*  
514 *of Pharmaceutics*, 547, 432-437. doi:10.1016/j.ijpharm.2018.06.012

515 (17) Joshi, V., Brewster, D., & Colonero, P. (2012). In vitro diffusion studies in transdermal  
516 research: a synthetic membrane model in place of human skin. *Drug Development and*  
517 *Delivery*, 12, 40-42.

518 (18) Hillwalker, W. E., & Anderson, K. A. (2014). Bioaccessibility of metals in alloys:  
519 evaluation of three surrogate biofluids. *Environmental Pollution*, 185, 52-58.  
520 doi:10.1016/j.envpol.2013.10.006

521 (19) USEPA. (2004). Risk Assessment Guidance for Superfund Volume I: Human Health  
522 Evaluation Manual (Part E, Supplemental Guidance for Dermal Risk Assessment):  
523 EPA/540/R/99/005. OSWER 9285. 7-02EP PB99-963312, Washington, DC, USA.

524 (20) Clesceri, L. S., Greenberg, A.E., & Eaton, A.D. (1999). *Standard Methods for the*  
525 *Examination of Water and Wastewater*. American Public Health Association, twentieth  
526 edition.

527 (21) ASTM. (2013). D4972-13 Standard Test Method for pH of Soils. West Conshohocken,  
528 PA.

529 (22) Tiessen, H., & Moir, J. (1993). Total and Organic Carbon. Soil Sampling and Methods  
530 of Analysis. Martin R. Carter (Ed.). Canadian Society of Soil Science: Lewis Publishers.  
531 Boca Raton, Florida, USA. pp.

532 (23) Chapman, H. D. (1965). Cation-exchange capacity. Methods of Soil Analysis: Part 2  
533 Chemical and Microbiological Properties, 9, 891-901.

534 (24) CEN (2015). Reference test method for release of nickel from all post assemblies which  
535 are inserted into pierced parts of the human body and articles intended to come into direct  
536 and prolonged contact with the skin. EN 1811: European Committee for Standardization.

537 (25) Wainman, T., Hazen, R. E., & Lioy, P. J. (1994). The extractability of Cr(VI) from  
538 contaminated soil in synthetic sweat. Journal of Exposure Analysis and Environmental  
539 Epidemiology, 4, 171-181.

540 (26) Wertz, P. W. (2008). Human synthetic sebum formulation and stability under conditions  
541 of use and storage. International Journal of Cosmetic Science, 31, 21-25.

542 (27) Franken, A., Eloff, F. C., Du Plessis, J., Badenhorst, C. J., & Du Plessis, J. L. (2015b).  
543 In vitro permeation of platinum through African and Caucasian skin. Toxicology Letters,  
544 232, 566-572. doi:10.1016/j.toxlet.2014.12.010

545 (28) Ministère de l'Environnement et de la Lutte contre les changements climatiques (2021).  
546 Guide d'intervention - Protection des sols et réhabilitation des terrains contaminés, Annexe

547 2 retrieved 26-03-2021 from <https://www.environnement.gouv.qc.ca/sol/terrains/guide->  
548 intervention/annexe2.pdf

549 (29) Liang, S., Guan, D.-X., Li, J., Zhou, C.-Y., Luo, J., & Ma, L. Q. (2016). Effect of aging  
550 on bioaccessibility of arsenic and lead in soils. *Chemosphere*, 151, 94-100.  
551 doi:<https://doi.org/10.1016/j.chemosphere.2016.02.070>

552 (30) Larese, F., Gianpietro, A., Venier, M., Maina, G., & Renzi, N. (2007). In vitro  
553 percutaneous absorption of metal compounds. *Toxicology Letters*, 170, 49-56.  
554 doi:<https://doi.org/10.1016/j.toxlet.2007.02.009>

555 (31) McLean, J. E., & Bledsoe, B. E. (1992). Ground water issue: behavior of metals in soils.  
556 United States Environmental Protection Agency (EPA/540/S-92/018), Washington.

557 (32) Dobran, S., & Zagury, G. J. (2006). Arsenic speciation and mobilization in CCA-  
558 contaminated soils: Influence of organic matter content. *Science of The Total Environment*,  
559 364, 239-250.

560 (33) Hostýnek, J. J., Hinz, R. S., Lorence, C. R., Price, M., & Guy, R. H. (1993). Metals and  
561 the skin. *Critical Reviews in Toxicology*, 23, 171-235.

562 (34) Gammelgaard, B., Fullerton, A., Avnstorp, C., & Menné, T. (1992). Permeation of  
563 chromium salts through human skin in vitro. *Contact Dermatitis*, 27, 302-310.

564 (35) Van Lierde, V., Chery, C. C., Moens, L., & Vanhaecke, F. (2005). Capillary  
565 electrophoresis hyphenated to inductively coupled plasma-sector field-mass spectrometry for  
566 the detection of chromium species after incubation of chromium in simulated sweat.  
567 Electrophoresis, 26, 1703-1711. doi:10.1002/elps.200410221

568 (36) World Health Organization. (2009). Inorganic chromium (III) compounds. World  
569 Health Organization.

570 (37) Hostýnek, J.J. (2003). Factors determining percutaneous metal absorption. Food and  
571 Chemical Toxicology, 41, 327-345.

572 (38) Gonsalvesh, L., Marinov, S., Gryglewicz, G., Carleer, R., & Yperman, J. (2016).  
573 Preparation, characterization and application of polystyrene based activated carbons for Ni  
574 (II) removal from aqueous solution. Fuel Processing Technology, 149, 75-85.

575 (39) Fullerton, A., Andersen, J., & Hoelgaard, A. (1988). Permeation of nickel through  
576 human skin in vitro—effect of vehicles. British Journal of Dermatology, 118, 509-516.

577 (40) Stauber, J., Florence, T., Gulson, B., & Dale, L. (1994). Percutaneous absorption of  
578 inorganic lead compounds. Science of The Total Environment, 145, 55-70.

579 (41) Powell, K. J., Brown, P. L., Byrne, R. H., Gajda, T., Hefter, G., Leuz, A.K., Sjöberg,  
580 Staffan., & Wanner, H. (2015). Chemical Speciation of Environmentally Significant Metals:  
581 An IUPAC contribution to reliable and rigorous computer modelling. Chemistry  
582 International, 37, 15-19.

583 (42) Mavon, A., Miquel, C., Lejeune, O., Payre, B., & Moretto, P. (2007). In vitro  
584 percutaneous absorption and in vivo stratum corneum distribution of an organic and a mineral  
585 sunscreen. Skin Pharmacology and Physiology, 20, 10-20.

586 (43) Merwe, D. v. d., Tawde, S., Pickrell, J. A., & Erickson, L. E. (2009). Nanocrystalline  
587 titanium dioxide and magnesium oxide in vitro dermal absorption in human skin. Cutaneous  
588 and Ocular Toxicology, 28, 78-82.

589 (44) Hostýnek, J.J., Hinz, R., Lorence, C., & Guy, R. (1998). Human skin penetration by  
590 metal compounds. *Drugs and the Pharmaceutical Sciences*, 91, 647-668.

591 (45) Filon, F. L., D'Agostin, F., Crosera, M., Adami, G., Bovenzi, M., & Maina, G. (2008).  
592 In vitro percutaneous absorption of chromium powder and the effect of skin cleanser.  
593 *Toxicology in Vitro*, 22, 1562-1567.