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Critical review of activated sludge modelling: State of process

knowledge, modelling concepts and limitations

2 3 Hauduc H. a,b, Rieger L.c, Oehmen A.d, van Loosdrecht M.C.M.e, Comeau Y.f, Héduit A.a, 4 5 Vanrolleghem P.A.^b and Gillot S.^a 6 7 ^a Cemagref, UR HBAN, Parc de Tourvoie, BP 44, F-92163 Antony Cedex, France. 8 b modelEAU, Département de génie civil et de génie des eaux, Université Laval, Pavillon Adrien-Pouliot, 1065 9 av. de la Médecine, Quebec (QC) G1V 0A6, Canada. 10 ^c EnviroSim Associates Ltd., McMaster Innovation Park, 175 Longwood Rd S, Suite 114A, Hamilton, ON L8P 11 0A1, Canada. 12 d REQUIMTE/CQFB, Chemistry Department, FCT, Universidade Nova de Lisboa, 2829-516 Caparica, Portugal. 13 ^e Department of Biotechnology, Delft University of Technology, Julianalaan 67, 2628 BC Delft, The 14 Netherlands. 15 ^f Ecole Polytechnique, Department of Civil, Geological and Mining Engineering, P.O. box 6079, Station Centre-16 ville, Montreal H3C 3A7QC, Canada. 17 18 19 **Abstract** 20 This work critically reviews modelling concepts for standard activated sludge wastewater treatment 21 processes (e.g. hydrolysis, growth and decay of organisms, etc.) for some of the most commonly used 22 models. Based on a short overview on the theoretical biochemistry knowledge this review should help 23 model users to better understand i) the model concepts used; ii) the differences between models and 24 iii) the limits of the models. 25 The seven analysed models are: (1) ASM1; (2) ASM2d; (3) ASM3; (4) ASM3+BioP; (5) 26 ASM2d+TUD; (6) Barker&Dold model; (7) UCTPHO+. Nine standard processes are distinguished 27 and discussed in the present work: hydrolysis; fermentation; ordinary heterotrophic organisms (OHO)

growth; autotrophic nitrifying organisms (ANO) growth; OHO & ANO decay; poly-

29	hydroxyalkanoates (PHA) storage; polyphosphate (polyP) storage; phosphorus accumulating
30	organisms PAO) growth and PAO decay.
31	For a structured comparison, a new schematic representation of these processes is proposed. Each
32	process is represented as a reaction with consumed components on the left of the figure and produced
33	components on the right. Standardised icons, based on shapes and colour codes, enable the
34	representation of the stoichiometric modelling concepts and kinetics. This representation allows
35	highlighting the conceptual differences of the models, and the level of simplification between the
36	concepts and the theoretical knowledge. The model selection depending on their theoretical
37	limitations and the main research needs to increase the model quality are finally discussed.
38	
39	Keywords: ASM, biological nutrient removal, modelling concepts, model selection.
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41	NOMENCLATURE
42	An: Anaerobic conditions
43	ANO: Autotrophic nitrifying organisms
44	ASM: Activated sludge model
45	ATP: Adenosine triphosphate
46	Ax: Anoxic conditions
47	COD: Chemical oxygen demand
48	BioP: Biological phosphate removal
49	FADH: Flavin adenine dinucleotide
50	GAO: Glycogen accumulating organisms
51	ISS: Inorganic suspended solids
52	NADH: Nicotinamide adenine dinucleotide
53	OHO: Ordinary heterotrophic organisms

Ox: Oxic (aerobic) conditions

- 55 **PAO**: Phosphorus accumulating organisms
- 56 **PHA**: Poly-β-hydroxyalkanoates
- 57 **PHB**: Poly-β-hydroxybutyrate
- 58 **PHV**: poly-β-hydroxyvalerate
- 59 **PolyP**: polyphosphates
- 60 TCA: Tricarboxylic acid (cycle)
- 61 **ThOD**: Theoretical oxygen demand
- 62 **TSS:** Total suspended solids
- 63 VFA: Volatile fatty acids
- 64 VSS: Volatile suspended solids
- 65 **WWTP**: Wastewater treatment plant

66 1. INTRODUCTION

- 67 Since ASM1 (Henze et al., 1987), a dozen Activated Sludge Models (ASM) and even more
- extensions have been published. They have fixed some shortcomings of ASM1 and included
- 69 new process insights. Nevertheless, ASM1 remains the most commonly reported model in
- 70 literature. Indeed, the results of an international survey among ASM users (Hauduc et al.,
- 71 2009) revealed that models are found too complex for 22% of the respondents, and that 24%
- 72 of the model users do not trust their model. Furthermore, self training is the main source of
- 73 knowledge for 78% of model users. Consequently, users are generally not mastering all
- 74 published models to be able to choose the most suitable one for their modelling project, and
- 75 ASM1 turns out too often be their first choice.
- 76 Since the first publication of ASM1 (Grady et al., 1986), the biokinetic models are
- 77 represented in a table format, which is named in practice Gujer matrix or Petersen matrix
- 78 (Takács, 2005). This table contains a stoichiometric matrix and a kinetic vector. Each row of
- 79 this matrix stands for a process and each column for a state variable of the model. The

stoichiometric coefficients, negative for consumed compounds and positive for produced ones, are stored in the cells at the intersection of the corresponding row and column. The process rates vector is presented in the rightmost column. This representation is very convenient, as it gathers complex models into a condensed form and facilitates their publication. It also allows seeing at once all state variables involved in a process, and all processes in which a state variable is involved. However, in case of large models such as ASM2d for example, it becomes difficult to "read" this matrix. Comeau and Takács (2008) proposed a schematic representation of ASM, in which each model is represented in a single scheme. This allows a global view of the model processes and their interactions, which is very helpful as a learning tool to understand the models. However, as processes are illustrated in different schematics and as stoichiometric and kinetic information are not represented, it is difficult to compare modelling concepts used in different models, and thus to compare models in detail. This work aims at helping model users to better understand i) the model concepts used; ii) the differences between models and iii) the limits of the models. Seven published models have been chosen for this study: (1) ASM1 (Henze et al., 1987; 2000a); (2) ASM2d (Henze et al., 1999; 2000b); (3) ASM3 (Gujer et al., 1999; 2000); (4) ASM3+BioP (Rieger et al., 2001); (5) ASM2d+TUD (Meijer, 2004); (6) Barker&Dold model (Barker and Dold, 1997); (7) UCTPHO+ (Hu et al., 2007). Nine standard processes have been identified and will be discussed in separate sections of this paper: hydrolysis; fermentation; ordinary heterotrophic organisms (OHO) growth; autotrophic nitrifying organisms (ANO) growth; OHO & ANO decay; poly-hydroxyalkanoates (PHA) storage; polyphosphate (polyP) storage; phosphorus accumulating organisms (PAO) growth and PAO decay.

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For each standard process, (i) a brief overview on the available biochemical knowledge is provided as basis for discussion of the modelling concepts. The major publications are cited for further reading. Then, (ii) the different modelling concepts used are compared through a new schematic representation of the stoichiometry and the kinetics. Finally, (iii) the consequences of the model simplifications are investigated to draw theoretical limits of the models. Alternative published models that address the studied model limits are cited.

The final discussion synthetises the modelling concept diversity and the grey areas in theoretical knowledge, and discusses the model selection and existing model modifications.

2. METHODOLOGY

2.1. Studied models

The seven published models have been chosen among those most commonly reported in literature (Table 1). To keep the paper readable, those references will not be repeated each time. Two of the seven models, ASM1 and ASM3, only consider carbon and nitrogen removal, whereas the others also consider biological phosphorus removal. UCTPHO+ is an update of the UCTPHO model (Wentzel *et al.*, 1992), and ASM2d+TUD is the last published version of the integrated ASM-TUD (Technical University of Delft) metabolic model. This paper is using the corrected model formulations for typos and continuity problems (Hauduc *et al.*, 2010b). The full Gujer Matrices are available in a spreadsheet format as additional material through that reference.

Table 1 indicates the size of the models through the number of processes, state variables, interacting processes and variables (number of non-empty cells of the stoichiometric matrix) and parameters they include. It can be deduced that differences come mainly from the PAO processes. However, the real difficulty for model users results from the number of parameters

that should actually be calibrated (Hauduc *et al.*, 2010a). Many parameters, particularly stoichiometric, seldom require adjustment. Some kinetic parameters (e.g. hydrolysis and decay processes) are routinely calibrated in practice. Understanding the different modelling approaches and their concepts greatly aids in determining when a parameter should be calibrated for a given situation.

2.2. Model processes

- Nine standard processes have been identified and are listed in Table 2. These "standard processes" involve mechanisms that only differ by the environmental conditions under which they take place. For instance, aerobic and anoxic OHO growth processes are combined as one OHO standard growth process.
- Table 2 also synthesises the processes considered in each model. This work is limited to biological processes, and therefore chemical phosphorus precipitation is not discussed.
- As OHO- and ANO-related processes of ASM2d+TUD are exactly the same as ASM2d, ASM2d+TUD will be studied only for BioP-related processes.

2.3. A new schematic representation

A new schematic representation of the model processes is proposed to facilitate model concept comparison in a systematic and transparent way. For each process type, the standard processes that use the same modelling concept are represented on a single figure, and the standard processes that are different in terms of modelling concept are represented on separate figures. The process is represented as a reaction with consumed components on the left of the figure and produced components on the right. Figure 1 shows the symbols used for the schematic representation:

- 148 Models that consider the process are given by a letter (a to g).
- 149 The electron acceptor condition of the process is indicated by a square (Ox: aerobic;
- 150 Ax: anoxic; An: anaerobic), close to the corresponding model name.
- 151 The included state variables are represented through both a shape and a background: 152 the shape indicates whether the variable is particulate or colloidal, soluble or refers to 153 an organism and internal storage, and the background indicates its composition in 154 terms of ThOD (theoretical oxygen demand) (C), nitrogen (N) or phosphorus (P). The 155 state variable name is indicated inside the shape, using the standardised notation from
- 156 Corominas et al. (2010).

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- 157 The electron acceptor consumed in the process is represented by a square. For 158 instance, depending on its usage, nitrate can be represented by a circle-square
- 159 (electron acceptor) or by a circle only (substrate).
- 160 To simplify the graphs, alkalinity and total suspended solids are not represented. Only important stoichiometric coefficients (especially yields) are specified, as others can be calculated through conservation of ThOD, nitrogen and phosphorus (Hauduc et al., 2010b).
 - In this process representation, the consumed and produced components of the reaction are linked by black arrows. In case state variables are not considered by all models or under all conditions, the models or conditions concerned are specified on the arrow.
- 166 The kinetic expression for the rate of reaction is also part of the concept and is therefore 167 represented. This is also done in a condensed way by a standardised compact notation. Table 168 3 illustrates the different symbols used to keep the expression readable. The saturation 169 functions and inhibition functions, which have the form of a Monod expression, are expressed M() or IM(), with the component concerned in parenthesis. The symbol $\langle \rangle$ is used to indicate 170

optional or alternative terms depending on the model or the environmental condition (see Table 3 for examples).

2.4. Literature review on theoretical knowledge vs modelling concepts

The results are organised around the nine main processes listed in Table 2. For each process, the theoretical knowledge is first presented based on a brief literature review. Then the different concepts used in the models are analysed, supported by the schematic representation of the process. Different concepts are given different numbers (concept 1, concept 2...), whereas variations within the same concept are pointed out using letters (concept 1a, 1b...). Finally, model limitations in using particular concepts are listed.

3. RESULTS

3.1. General modelling concepts

3.1.1. State variables

Some conceptual differences among the studied models come from the state variables used, as summarised in Table 4.

All studied models are theoretical COD (ThOD) based. ThOD is the chosen organic material measure because it is a conservative quantity that allows characterising the electron equivalents of organic substrates, biomass and electron acceptors (Ekama and Marais, 1979; Henze *et al.*, 2000a). However a discrepancy exists between ThOD and analytical COD for compounds with negative ThOD, such as nitrite which is analytically measured as positive COD (Gujer and Larsen, 1995), moreover some compounds are not detected in a standard COD test.

3.1.2. Component-based / fraction-based is	a moaeis
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In the ASM1 and the Barker & Dold model, organic nitrogen is considered in separate state variables (component-based model) and organic phosphorus is not considered. In these models nitrogen and phosphorus (Barker & Dold only) are, however, linked to biomass. In other models, nitrogen and phosphorus are linked to ThOD state variables (fraction-based models).

3.1.2.1. Component-based model

<u>Pros.</u> As variables are separated, the hydrolysis processes of organic ThOD and organic nitrogen are independent. It is thus easier to change parameters in case of variations in influent fractions.

Cons. In the Barker & Dold model, organic phosphorus is not considered separately. Phosphorus is then always available in the form of PO₄³⁻, without any delay due to hydrolysis, when released by biomass decay.

3.1.2.2. Fraction-based model

<u>Pros.</u> Having linked variables limits the number of variables and processes.

<u>Cons.</u> The substrate fraction is supposed to be homogeneous and constant in its composition. In reality different organic compounds with different fractions are coming in the influent. In this case this concept could induce a pitfall or the model complexity must be increased by considering more components.

3.1.3. Undegradable organics from influent and from biomass decay

ASM1, the Barker & Dold model and UCTPHO+ distinguish the influent undegradable organics, from those formed by biomass decay (or endogenous respiration). The latter fraction thus includes non-active biomass. This distinction allows a different nitrogen and phosphorus

fraction content in soluble undegradable matter from influent and endogenous products, especially for the component-based models (ASM1 and Barker & Dold model, see above), increasing model flexibility.

3.1.4. N_2 considered as a state variable to close the nitrogen balance

To simplify the models, ASM1, the Barker & Dold model and UCTPHO+ do not consider dissolved dinitrogen gas (N_2) as a state variable. As a consequence, the continuity is not verified for these processes in terms of nitrogen (Hauduc *et al.*, 2010b). This state variable needs to be added to check the continuity of the model.

3.1.5. Total suspended solids (TSS) as a state variable

TSS is a combined variable in some models and a state variable in others, calculated from the linear combination of particulate state variables and from an assumed VSS/COD and VSS/TSS ratio (VSS being the volatile suspended solids) to predict the sludge mass in the system. Consequently, the ISS (inorganic suspended solids), which are defined as part of the TSS together with the VSS (TSS=VSS+ISS), are only empirically predicted in these models through the ratio VSS/TSS, which leads to unreliable estimations of TSS. Ekama and Wentzel (2004) propose a model for ISS that takes into account ISS from the influent, the precipitation of minerals and the inorganic content of biomasses.

TSS is not considered in this study.

3.1.6. Alkalinity

A low alkalinity value (<50 g CaCO₃.m⁻³ = 1 mol HCO₃⁻.m⁻³) results in an unstable pH, which could cause nitrification inhibition and other process problems (Henze *et al.*, 2000a).

Alkalinity is thus modelled to predict the risk of pH limitation. As discussed previously

(Hauduc *et al.*, 2010b), alkalinity is considered or not depending on the models. In this study, alkalinity is not considered.

3.2. OHO and ANO processes

3.2.1. Hydrolysis of particulate substrate

3.2.1.1. Knowledge

A large fraction of wastewater substrate is particulate or colloidal, and is thus not directly available for biomass growth (Ekama and Marais, 1979; Morgenroth *et al.*, 2002). Hydrolysis is an extracellular biological reaction where hydrolytic enzymes break down large organic molecules into smaller ones that can pass through the bacterial cell wall. Hydrolytic enzymes seem to be bound to the floc and have a low turnover rate (hours to days) which enable hydrolysis to be decoupled from the enzyme synthesis (Goel *et al.*, 1999).

<u>Substrate</u>. The diversity of substrates, hydrolytic enzymes and biological pathways make the hydrolysis process difficult to study. Experiments described in literature are mainly based on pure culture bacteria, with single or few substrates, providing results that can hardly be generalised.

Protozoa are also able to take up particulate substrate as shown by de Kreuk *et al.* (2010), and possibly release readily biodegradable substrate in the process. However, this process is so far poorly described.

Electron acceptor conditions. Hydrolytic enzyme synthesis depends on the electron acceptor conditions, but their activity is not affected by the electron acceptor conditions (Goel *et al.*, 1999), which enables hydrolysis processes to continue under anoxic and anaerobic conditions. Hydrolysis due to protozoa activity will however depend on oxic conditions.

3.2.1.2. Modelling

- In the models studied here, only two concepts are used (Table 5):
- The first concept is one step hydrolysis, where slowly biodegradable substrate is hydrolysed, then consumed by organisms. The differences between the models concern the way the residues of the reaction and the nitrogen fractions are modelled (3.1.2):
 - Component-based model (ASM1, Barker & Dold);
 - Fraction-based model (ASM2d, ASM3, ASM3+BioP).
 - The second concept used in UCTPHO+, is based on direct growth using adsorbed substrate. The hydrolysis is accounted for by a reduced growth rate for the use of this adsorbed substrate (see OHO growth process, Table 7). This way to model hydrolysis makes the hydrolysed substrate available for the organisms that produce hydrolysis enzymes, whereas in other modelling concepts the hydrolysed substrate is released into the bulk phase, in this way becoming available for all organisms, which will thus compete for it.
 - The hydrolysis process is in fact used to model all mechanisms that make slowly biodegradable substrate available for bacterial growth with a certain delay (chemical dissolution, mass transport, storage, etc.). Consequently, depending on the other processes considered in the model, the hydrolysis process does not have the same significance:
 - Storage is considered as a separate process in ASM3 and ASM3+BioP, whereas it is not explicitly described in other models. However, storage and hydrolysis cannot be distinguished through respirometric methods (Goel *et al.*, 1999). Consequently, in ASM1, ASM2d, Barker & Dold and ASM2d+TUD, the storage is implicitly included in the hydrolysis process.

Depending on the origin of the organic molecules, two types of hydrolysis reactions can be distinguished: hydrolysis of "primary substrate" that comes from the influent and hydrolysis of the matter produced by biomass metabolism or decay, named "secondary substrate", in which protozoa may play an important role (Morgenroth *et al.*, 2002). Consequently, models using the death-regeneration concept to model biomass decay (see paragraph 3.2.5.2) merge those two types of hydrolysis in a single process, whereas in case of the endogenous respiration concept, the hydrolysis of secondary substrate is modelled through endogenous respiration and maintenance processes (ASM3 and ASM3+BioP).

Electron acceptor conditions. The storage process and the utilisation of secondary substrate require an electron acceptor to produce energy. Models that implicitly merge these processes into the hydrolysis process (ASM1, ASM2d, Barker & Dold model, UCTPHO+ and ASM2d+TUD, see above), have then to take into account their electron acceptor. Thus, the hydrolysis kinetic rates should depend on the electron acceptor. However, ASM1 does not consider hydrolysis under anaerobic conditions. As ASM3 and ASM3+BioP consider storage and hydrolysis separately, the electron acceptor is not rate limiting for hydrolysis.

The Barker & Dold model introduces a hydrolysis yield in anoxic and anaerobic conditions that allows modelling the experimentally observed "COD loss" (Barker and Dold, 1995). Although this observation is not explained so far, the "loss" is modelled by H_2 gas formation (Kraemer *et al.*, 2008). The S_{H2} state variable is thus added to the model to reach continuity (Hauduc *et al.*, 2010b).

As UCTPHO+ models the hydrolysis process simultaneously with growth, anaerobic hydrolysis is not modelled.

<u>Ammonification</u>. In case of component-based models (ASM1, Barker & Dold), biodegradable organic nitrogen is produced by the hydrolysis process. To make this nitrogen available for organisms, ammonification has to be modelled.

3.2.1.3. Model limitations

<u>Substrate.</u> The concept of one step hydrolysis is used by all models but one (UCTPHO+). This concept implies a simplification of the (primary) substrate into a single biodegradable particulate fraction. In case of peculiar influents with different particulate substrates behaviour or large colloidal fractions, it may be required to integrate other particulate fractions and to consider other hydrolysis concepts, such as parallel hydrolysis or sequential hydrolysis (Orhon *et al.*, 1998; Nowak *et al.*, 1999; Larrea *et al.*, 2002).

<u>Electron acceptor conditions.</u> Hydrolysis enzyme activity is independent of the electron acceptor (Goel *et al.*, 1999). However, the hydrolysis process also covers other mechanisms that require an electron acceptor, such as degradation by protozoa and storage. In case of a large anaerobic zone, anaerobic hydrolysis should be considered, especially for BioP models to make substrate available for PHA storage (ASM1 and UCTPHO+).

Experimental determination of parameters. Modelling hydrolysis and storage as two separated processes as in ASM3 and ASM3+BioP, requires adequate experiments to independently determine the kinetic rates (Goel *et al.*, 1999).

3.2.2. Fermentation

3.2.2.1. Knowledge

Fermentation is a growth process under anaerobic conditions for OHOs. In the absence of an electron acceptor, oxidative processes inside the cells are not possible and the substrate is

partially oxidised to CO₂ and partially reduced to products. Under these conditions, organic substrate is catabolised into volatile fatty acids (VFA, e.g. acetate), associated to organisms growth.

3.2.2.2. Modelling

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- Two different concepts are used to model fermentation in the studied models (Table 6):
- The first concept considers fermentation as a transformation (ASM2d, UCTPHO+)
- In the second concept fermentation is described as an anaerobic growth process.
- The process kinetic rate always depends on the OHO concentration, which is considered as the only biomass involved in this process.
- Barker and Dold (1995) experimentally observed a COD "loss" during anaerobic processes, which they linked to fermentation, anaerobic hydrolysis and S_{VFA} sequestration. This phenomenon has been modelled by a S_{VFA} formation yield (Y_{fe}) in the fermentation process.

 The loss of (1-Y_{fe}) g ThOD.g S_{VFA}⁻¹ is modelled through H₂ gas formation by Kraemer *et al.* (2008), and a S_{H2} state variable has thus been added to reach model continuity (Hauduc *et al.*,

3.2.2.3. Model limitations

ASM1 and ASM3 do not consider fermentation, consequently only one soluble substrate is considered (S_B). Fermentation is considered only in models with BioP, since PAOs are assumed to only grow on fermentation products (S_{VFA}) and not on fermentable products (S_F). However, fermentation is not considered in ASM3+BioP: hydrolysis is considered as the rate-limiting step, so that the fermentation process rate does not need to be considered explicitly. This could be a model limitation in cases where hydrolysis is no longer the rate-limiting step,

e.g. in the case of a peculiar influent (e.g. from agro-industries with high S_B concentration), or a specific plant configuration (e.g. with hydrolysis of return activated sludge).

All models except the Barker & Dold model neglect OHO formation during fermentation. Indeed, Ekama and Wentzel (1999) estimate the anaerobic growth yield at $0.10 \mathrm{~g~X_{OHO.g}~S_B^{-1}}$. In the case of large anaerobic zones, anaerobic growth may not be neglected and concept 2

(Table 6) should be chosen.

3.2.3. Ordinary heterotrophic organisms (OHO) growth

3.2.3.1. Knowledge

Under aerobic or anoxic conditions OHOs use organic substrate as an energy and carbon source. The yield of biomass growth is the fraction of substrate that is used as a carbon source to produce biomass (figure 2).

<u>Substrate</u>. van Loosdrecht *et al.* (1997b) proposed the existence of two types of bacteria in terms of their capacity for substrate storage. Bacteria not capable of substrate storage will maximize their growth rate in periods with available substrate in order to be competitive, but will not be able to maintain their cell structure in case of long starvation periods. In case of a highly dynamic influent or in case of a process with a feast/famine cycle, bacteria capable of storing substrate will have a strong competitive advantage due to their ability to maintain a low growth rate during starvation periods, which enables them to keep all of their cell system viable (van Loosdrecht *et al.*, 1997b).

Stored compounds, e.g. poly-β-hydroxybutyrate (PHB), result from additional substrate that is taken up on top of the substrate requirement for direct growth (van Aalst-Van Leeuwen *et al.*, 1997). However, the nature of the storage compounds is still not well understood but seems to

depend on the substrate used (Beccari et al., 2002). For instance, glycogen can be formed 372 from glucose (Dircks et al., 2001), and cases where propionate constitutes an important 373 374 substrate fraction in the influent would lead preferentially to an increased poly-β-375 hydroxyvalerate (PHV) fraction as compared to PHB (Oehmen et al., 2007). 376 The substrate uptake rate increases instantaneously when a high substrate concentration 377 occurs (up to the maximum rate), but the growth rate increases only slowly, and the extra 378 substrate taken up may be stored (van Aalst-Van Leeuwen et al., 1997). Consequently, a high 379 growth rate (e.g. at short SRT) will result in less storage (Beun et al., 2002; van Loosdrecht 380 and Heijnen, 2002). The growth rate on stored compounds is lower and limited by the storage 381 product degradation process, which depends on its content of the biomass following a first 382 order relationship (van Loosdrecht et al., 1997b; Beun et al., 2002). 383 Nutrients. Organism growth also requires nutrients such as nitrogen or phosphorus to 384 synthetise their cells constituents (proteins, nucleic acids...). In case of ammonia depletion, 385 OHOs are able to use nitrate as nitrogen source. For instance, Wentzel et al. (1989) 386 experimentally proved that when ammonia is depleted, PAOs consume nitrates for their 387 growth with no modification of their kinetic behaviour. However, the yield will be slightly 388 lower since some ThOD is used to reduce nitrate to ammonium. 389 Denitrification. Heterotrophic growth under anoxic conditions requires oxidised forms of 390 nitrogen as electron acceptor: nitrate (NO₃), nitrite (NO₂), nitric oxide (NO) or nitrous oxide 391 (N₂O). If denitrification is complete, these electron acceptors are reduced sequentially to 392 nitrogen gas (N₂). 393 The need to use a different OHO yield under anoxic conditions $(Y_{OHO,Ax})$ to properly model 394 the nitrate and COD consumptions was experimentally shown by several authors (Orhon et 395 al., 1996; Sozen et al., 1998; Muller et al., 2003). This is due to the lower ATP (adenosine triphosphate, which transports energy in the cell) formation through oxidative phosphorylation in the electron transport chain (lower ATP formation per NADH (Y_{NADH_ATP}) in metabolic models) when nitrate is the electron acceptor instead of oxygen (Beun *et al.*, 2002; Muller *et al.*, 2003).

3.2.3.2. Modelling

The stoichiometry of OHO growth requires an organic substrate, an electron acceptor and nutrients. The modelling concepts differ in the substrates used, the nitrogen source and the use of different yields for aerobic and anoxic conditions (Table 7):

- The first concept considers direct growth of OHOs on readily biodegradable substrate:
- In concept 1a, NH_X is the only nitrogen source (ASM1, ASM2d),
 - In concept 1b, NO_X can be used as a nitrogen source in case of ammonia depletion (Barker & Dold, UCTPHO+). Additionally, UCTPHO+ considers adsorption of particulate substrate onto OHOs, followed by direct growth on the adsorbed substrate.
- The second concept first considers substrate storage and then OHOs growth on storage compounds as the unique carbon source (ASM3, ASM3+BioP).

The adsorption and storage processes are particularly useful in case of cyclic loading conditions and selector modelling. The kinetics of these processes are considered to depend on the ratio of adsorbed or stored substrate to biomass and are associated with a maximum adsorption/storage potential (Ekama and Marais, 1979). The kinetic expression for adsorption in the UCTPHO+ model is in agreement with this statement. However, ASM3 and ASM3+BioP only use a Monod expression for substrate uptake, and thus consider that the maximum storage potential cannot be reached under normal wastewater treatment conditions.

419 Substrate: Several substrates are used depending on the models (Table 8):

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- 420 - Readily biodegradable substrate (S_B) or fermentable substrate (S_F) are used by all 421 models except ASM3 and ASM3+BioP, which model indirect growth only.
- Volatile fatty acids (S_{VFA}) are considered in Bio-P models, except ASM3+BioP. For 423 this substrate, OHOs compete with PAOs when present under aerobic or anoxic 424 conditions.
 - Adsorbed particulate substrate (X_{OHO,Ads}) is considered in UCTPHO+. This substrate has to be hydrolysed before use, which occurs simultaneously with growth. Modelling adsorption processes is a way to slow down OHO substrate consumption and model the delay observed before growth occurs under certain conditions (feed/starvation). This way to model hydrolysis is chosen by the UCT group to avoid the competition of organisms on hydrolysed substrate with the hydrolysis products being consumed directly by the organisms that produce hydrolysis enzymes (Wentzel et al., 1992).
 - Stored substrate (X_{OHO,Stor}) is the only usable substrate in ASM3 and ASM3+BioP. Direct growth on external substrate is not considered. This concept is needed in alternating feeding/starvation phases of the plant; mainly for selector systems. It allows simulating the observed delay before OHO growth.

A substrate preference switching function should be used to avoid that the OHO specific growth rate increases above a maximum value if two substrates are present in high concentration (Henze et al., 2000b) as two OHO growth processes run in parallel. The substrate preference switching function usually used in ASM models is in the form:

$$\left(\frac{S_{Sub}}{K_{SSub} + S_{Sub}}\right) \cdot \frac{S_{Sub}}{\sum_{i} S_{Sub,i}}$$

with S_{Sub} being the considered substrate.

<u>Nutrients.</u> Barker & Dold and UCTPHO+ consider growth with NO₃ as nitrogen source in case of ammonia depletion. However, these models do not consider the reduction of nitrate in the redox balance.

Denitrification. Denitrification is modelled as one step: nitrate is considered the only possible electron acceptor. The maximum anoxic growth rate is lower than under aerobic conditions, either because $\mu_{OHO,Max}$ is intrinsically lower for OHOs under anoxic conditions, or because only a fraction of OHOs is able to denitrify. Furthermore, all models but two (ASM1 and ASM2d) use a lower anoxic growth yield (Table 8), since the efficiency of oxidative phosphorylation is lower under anoxic conditions.

3.2.3.3. Model limitations

Adsorption and storage. The ASM3 growth on stored substrate does not consider direct growth on soluble substrate. This might lead to inaccurate predictions in case of low SRT (<5 d) (van Loosdrecht and Heijnen, 2002), and long feast/famine cycles, which are conditions when growth rate and storage are not constant. Krishna and van Loosdrecht (1999), Karahan-Gül *et al.* (2003), Sin *et al.* (2005) and Guisasola *et al.* (2005) proposed ASM3 modifications considering parallel direct growth on soluble substrate and indirect growth on internally stored substrate.

Beccari et al. (2002) proposed a different modelling concept that includes first a biosorption step, in which substrate is absorbed by biomass without any transformation, contrary to the UCTPHO+ concept where substrate is adsorbed on the biomass. Then, the biosorbed substrate is used either for direct growth or is transformed into a stored compound, which is later used for growth. This modelling concept allows a better description of the ammonia profile,

because biosorption does not release the nitrogen content of the substrate into the mixed liquor, contrary to external hydrolysis of the adsorbed compound.

<u>Denitrification.</u> In case of a large anoxic zone, using a single growth yield value for anoxic and aerobic processes (ASM1, ASM2d) could lead to an overestimation of the denitrification process in terms of biomass production, and underestimation of substrate consumption and nitrogen removal. The underprediction of substrate consumption could also have an effect on other processes such as P removal. A different anoxic growth yield should be added, but the model will then require a recalibration of the hydrolysis and storage processes to compensate the substrate consumption and maintain the experimentally observed denitrification rate (Muller *et al.*, 2003).

3.2.4. Autotrophic Nitrifying Organisms (ANO) growth

3.2.4.1. Knowledge

ANO oxidise ammonia to produce the required energy for CO₂ uptake and growth. This oxidation of ammonia is named nitrification. It includes two steps that involve two distinct groups of autotrophic organisms: ammonia oxidisers and nitrite oxidisers (figure 3). In the first step, nitritation, ammonia oxidisers produce energy required for their growth through ammonia oxidation into nitrite. Then, in the nitratation step, nitrite oxidisers convert nitrite into nitrate to produce energy. The first oxidation (nitritation) consumes alkalinity (Downing et al., 1964).

Because a lot of energy is required to reduce CO₂, autotrophic organisms have a lower growth yield than OHOs. Furthermore, the nitrifiers, which are obligate aerobic organisms, have a higher requirement of oxygen than heterotrophs for their growth: in addition to their needs in electron acceptor for respiration, oxygen is used to oxidise ammonia. Therefore, to ensure

good nitrification, it is necessary to provide sufficient dissolved oxygen to the activated sludge and to maintain a minimum SRT to avoid the wash out of nitrifiers (Downing *et al.*, 1964). Nitrification is also inhibited by a low pH and sufficient alkalinity concentration (generally >50 g CaCO₃. m⁻³ = 1 mol HCO₃⁻ m⁻³) has thus to be maintained to ensure a stable pH (Henze *et al.*, 2000a).

3.2.4.2. Modelling

Nitritation is normally considered the limiting step in nitrification (Downing *et al.*, 1964). Consequently, nitrification is often modelled as a one step process, as in all studied models

3.2.4.3. Model limitations

(Table 9), and initially proposed by Lijklema (1973).

<u>Multi-step nitrification/denitrification.</u> The simplified concept of one-step nitrification is sufficient for most municipal wastewater systems.

However, the modelling project may require predicting nitrite accumulation (shortcut nitrification-denitrification, inhibitions...) or greenhouse gas emission, in the form of nitric and nitrous oxide. NO₂⁻ accumulation (partial nitrification) has actually been observed in specific situations such as unstable operation of municipal WWTP (e.g. due to insufficient oxygen, low temperature, low sludge age and inhibitory compounds), high temperatures, side stream processes or industrial influent (Sin *et al.*, 2008; Kaelin *et al.*, 2009).

These modelling objectives cannot be reached with any of the studied models, which consider nitrification and denitrification as a single step. Consequently some models have been extended with two step nitrification and denitrification, as reviewed by Sin *et al.* (2008). A model with four step denitrification (NO₂, NO and N₂O as intermediates) and two step

nitrification is also proposed by Hiatt and Grady (2008). Currently, considerable attention is paid to greenhouse gas production in wastewater treatment and in the near future this will certainly lead to much more detailed models of the nitrogen-related reactions.

Nitrification inhibition. Autotrophs are sensitive to inhibition (pH, nitrous acid, ammonia, chromium, nickel, copper, etc). Effects of some environmental conditions on the activated sludge process are reported in Gujer (2010). Inhibitory effects are considered to be constant and are thus accounted for in the growth rate value (Henze *et al.*, 2000a). This can cause calibration problems in case of variability in the concentration of these compounds in the influent or in the treatment plant. To detect variability in influent inhibitors, some authors developed online respirometric methods to determine inhibition kinetics of nitrification (Nowak *et al.*, 1995; Kong *et al.*, 1996; Vanrolleghem *et al.*, 1996).

3.2.5. OHO and ANO decay

3.2.5.1. Knowledge

van Loosdrecht and Henze (1999) published a literature review on the theoretical knowledge regarding maintenance, endogenous respiration, lysis, decay and predation. Oxygen consumption linked to a loss of biomass was observed by various authors since the end of the 19th century. This phenomenon has been explained by the concept of "endogenous respiration" during which bacteria use their own storage pools of organic matter for maintenance purposes instead of using external substrate. Other experiments have shown accumulation of undegradable matter in the absence of substrate, leading to the cryptic growth (growth on dead bacteria) or the "death-regeneration" concepts (van Loosdrecht and Henze, 1999).

531	These concepts lump several mechanisms that result in oxygen consumption and biomass
532	reduction (van Loosdrecht and Henze, 1999; Hao et al., 2010):
533	 dormancy of bacteria: non-active state, which is seen as undegradable particulate.
534	- internal decay: consumption of internally stored compounds in case of starvation. This
535	mechanism is not coupled to significant microorganism death.
536	- external decay: predation by protozoa, viral attack, and cell lysis (phages, etc). The
537	cell walls may not be degraded by protozoa and phages, which results in
538	unbiodegradable material production.
539	- maintenance: the energy required to maintain essential life conditions in the cell, such
540	as cell motility, maintenance of ion gradients, turnover of cell material (e.g. proteins,
541	RNA) and transport of material (Lopez et al., 2006; Hao et al., 2010).
542	It should be noted that the electron acceptor conditions have been found to influence the OHO
543	and ANO decay rates. The nitrification activity has been shown to be higher after prolonged
544	exposure to anaerobic and anoxic conditions as compared to aerobic conditions (Siegrist et
545	al., 1999; Munz et al., 2011), where the ANO decay rate was >50% lower under anaerobic
546	and anoxic conditions. A reduced anoxic vs aerobic decay rate was also observed for OHOs
547	(Siegrist et al., 1999).
548	3.2.5.2. Modelling
549	Two concepts are used (Table 10):
550	- The death-regeneration concept. Two sub-concepts have to be distinguished:
551	 Death-regeneration with a component-based model (ASM1, Barker & Dold);
552	 Death-regeneration with a fraction-based model (ASM2d, UCTPHO+).
553	 The endogenous respiration concept (ASM3, ASM3+BioP).

For OHOs and ANOs, maintenance is considered as negligible and is considered part of the decay or endogenous processes in all models.

<u>Death-regeneration concept</u>. The biomass decay results in the release of a fraction (1- $f_{XU_OHO,lys}$) of particulate substrate and a fraction $f_{XU_OHO,lys}$ of undegradable material. The released particulate substrate will be hydrolysed, and then used again for OHO growth. Consequently, ANO decay contributes to OHO growth.

This concept also allows modelling anaerobic decay, and modelling the high oxygen or nitrate demand observed after an anaerobic condition period (Warner *et al.*, 1986), which would not be possible with the endogenous respiration concept. However, maintenance and endogenous respiration are neglected.

Endogenous respiration concept. This concept is closer to experimental observations (Gujer *et al.*, 2000). In this process energy is provided by the oxidation of the organic matter contained in biomass, which leads to undegradable matter and nutrients release. As a consequence, there is no cycling of ThOD in the model, which simplifies model calibration. Models that consider a storage pool (ASM3, ASM3+BioP) have to consider storage degradation for maintenance: stored compounds are used to produce energy without biomass production. This process is similar to the maintenance concept of PAOs (see 3.3.4.2), and explains the fate of the OHO storage pool during OHO decay. This can be considered as endogenous respiration of the storage pool.

3.2.5.3. Model limitations

Biokinetic models using the endogenous respiration concept should have better identifiable parameters and should thus be easier to calibrate (Gernaey *et al.*, 2004). Indeed, the endogenous respiration concept parameters only influence the decay process of the considered

organism, whereas the death-regeneration concept parameters influence the decay of the concerned organism (autotrophs and heterotrophs), hydrolysis and the growth processes of heterotrophs (substrate availability). Furthermore, the death-regeneration concept induces a higher biomass production rate, which has a general effect on all kinetic rate constants. Consequently, kinetic parameters are not directly comparable between models using the endogenous respiration concept or the corresponding death-regeneration concept as presented by Dold *et al.* (1980). Predation. Predation is explicitly modelled by Curds (1971), Lijklema (1973), Moussa et al. (2005) and more recently by Ni et al (2010), considering a reduction of the active biomass through protozoa consumption, their concentration being between 5% and 10% of the MLVSS according to Curds (1971). Not considering predation may lead to variable kinetic parameter values depending on the WWTP conditions. Electron acceptor conditions. The concept of endogenous respiration does not allow decay under anaerobic conditions, since no electron acceptor for the respiratory chain is available. The death-regeneration concept has been developed to cope with the anaerobic decay process in case the endogenous respiration concept is used and to keep the model as simple as possible (Dold et al., 1980). However, under anaerobic/anoxic conditions predation by protozoa does not occur since they are strictly aerobes, and ANO and OHO decay rates have been shown to be lower (Siegrist et al., 1999). Consequently, anaerobic/anoxic decay could be considered as negligible under certain WWTP conditions. Alternatively, these lower anaerobic and anoxic decay rates could cause an underprediction of biomass concentrations; especially in cases of long periods with unsuitable nitrification conditions (rain events, weekends, holidays, etc.) (Siegrist et al., 1999).

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3.3. Biological phosphorus removal

- Phosphorus accumulating organisms (PAOs) have the ability to store carbon compounds in excess of normal metabolic requirements as poly-β-hydroxyalkanoates (PHA) and glycogen, and to store phosphorus in the form of polyphosphate (polyP). This ability is used in wastewater treatment to biologically remove phosphorus, by stimulating PAO growth by a sequence of anaerobic and aerobic (or anoxic) conditions. PAO metabolism is usually described by 2 or 3 steps:
- substrate uptake (usually volatile fatty acids, S_{VFA}) and storage as PHA, typically under anaerobic conditions, associated with glycogen (ASM2D+TUD) and polyP consumption (all bio-P models);
 - PolyP and glycogen storage pools restoration and PHA consumption under aerobic and anoxic conditions (modelled simultaneously with growth in the Barker & Dold and UCTPHO+ models);
 - PAO growth associated to PHA consumption under aerobic and anoxic conditions.
 - Organic substrate uptake under anaerobic conditions provides PAOs a competitive advantage over OHOs. Furthermore, the anaerobic conditions enable the formation of S_{VFA} from fermentable substrate S_F . The simplified mechanisms of these steps are represented in Table 11. The use of PAOs to biologically remove phosphorus is named the enhanced biological phosphorus removal process (EBPR or BioP).
 - Metabolic model. To conceptualise BioP, the Delft University of Technology (TUD) group introduced a metabolic model that considers cell internal reactions (Smolders *et al.*, 1994a; Smolders *et al.*, 1994b). The cell internal concentrations of metabolites (NADH, acetyl-CoA, ATP, etc.) are considered to be in steady state conditions. Consequently, these components

are not modelled, and only the overall stoichiometric reaction is formulated. This results in a model structure that is similar to the others.

3.3.1. PHA storage

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3.3.1.1. Knowledge

- Under anaerobic conditions, in the presence of substrate, PAO store PHA. Figure 4 illustrates the main biochemical steps of PHA storage. Some experiments (Comeau *et al.*, 1987; Wentzel *et al.*, 1989; Brdjanovic *et al.*, 1998a) indicated that PAOs can also store PHA under anoxic or even aerobic conditions, if sufficient substrate is available.
- Energy source. VFAs are transported in the undissociated form (associated to a proton), which causes dissipation of the membrane proton motive force. PolyP breakdown and phosphate release associated to proton ions allow the re-establishment of the proton motive force (Comeau *et al.*, 1986). This phosphate release is also concomitant with Mg²⁺ and K⁺ release, which serve as counter-ions for stabilisation of the polyP chain (Wentzel *et al.*, 1986).
 - The polyP breakdown also provides most of the required energy to metabolise substrate into acetyl-CoA (Comeau *et al.*, 1986) by phosphorylation of AMP into ADP (and later to ATP) through the AMP-phosphotransferase enzyme (Wentzel *et al.*, 1992).
- 639 <u>Reducing power</u>. Two theories for NADH production were developed: the "Comeau-Wentzel model" and the "Mino model" (Jenkins and Tandoi, 1991; Wentzel *et al.*, 1992).
- The "Comeau-Wentzel model" hypothesises that NADH is provided by the anaerobic
 oxidation of acetate through the TCA cycle.
- The "Mino model" considers that the NADH is provided by glycolysis under anaerobic conditions, turning stored glycogen into pyruvate and then into acetyl-CoA

645 and CO₂ (Mino et al., 1998; Oehmen et al., 2007). This reaction also provides energy 646 for acetate uptake and conversion to acetyl-CoA. 647 The Mino model theory is well accepted and supported by experimental evidence, but the 648 oxidative part of the TCA cycle seems to effectively supply part of the reducing power for 649 PHA formation under certain conditions (Mino et al., 1998; Zhou et al., 2010). Oehmen et al., 650 (2007) hypothesise that either each metabolic pathway is used by a specific microbial group 651 of PAOs, or that PAOs are able to use different metabolic pathways depending on their 652 internal or external conditions. 653 PolyP storage pool. Mino et al. (1985) and Wentzel et al. (1989) observed that not all the 654 stored polyP can be degraded. They hypothesised that two different polyP molecular weights 655 exist: short polyP chains have low molecular weight and can be released, whereas long polyP 656 chains cannot. However, no experimental evidence has supported this hypothesis so far. 657 Glycogen limitation, however, has been shown to result in the incomplete degradation of 658 polyP (Brdjanovic et al., 1998c). Such a limitation may occur at a low pH (less than 7.3) in 659 the presence of an excess of VFAs. 660 Substrate. Other substrates than acetate can be used by PAOs (Oehmen et al., 2007) such as 661 carboxylic acids, sugars and amino acids (Mino et al., 1998). However, most experiments 662 have been carried out on enriched cultures with acetate, which is usually considered as the 663 unique substrate source in order to simplify the models (Mino et al., 1998). 664 pH dependency. The energy requirements for S_{VFA} uptake have been observed to increase 665 with pH, leading to an increased phosphorus release to S_{VFA} uptake ratio. This is interpreted as a higher energy required for maintaining the proton motive force for S_{VFA} transport (Mino 666 667 et al., 1998).

Competition with GAOs. PAOs have to compete with GAOs (glycogen accumulating non-polyP organisms) for the VFAs under anaerobic conditions. Indeed, GAOs store acetate as PHA under anaerobic conditions without using polyP reserves. GAOs use this PHA as carbon and energy source for aerobic/anoxic growth and glycogen production. Their glycogen storage is used both as energy and reducing power source for anaerobic substrate uptake (Mino *et al.*, 1998). Therefore, GAOs have to to store more glycogen than PAOs (Sudiana *et al.*, 1999).

This competition seems to highly depend on external factors such as carbon source, pH, temperature, sludge age, dissolved oxygen concentration and inhibitory compounds (Meijer, 2004; Oehmen *et al.*, 2007). Lopez-Vazquez *et al.* (2009) concluded that GAOs are favoured by higher temperatures and lower pH.

3.3.1.2. Modelling

- The concepts vary in terms of substrate used (S_B or S_{VFA}) and in terms of source of energy (Table 12):
- In the first concept, energy for storage is provided by polyP breakdown, and reducing power production is not considered (ASM2d, UCTPHO+, Barker & Dold, ASM3+BioP).
- In the second concept that only concerns ASM2d+TUD, energy is provided by polyP breakdown and glycogen degradation, while reducing power is also generated through glycolysis.
- 688 PHA storage. In the first concept glycogen storage is not distinguished from PHA storage.
- Consequently, the storage pool for these models is named $X_{PAO,Stor}$ (Corominas *et al.*, 2010).

- 690 Energy source. In the Barker & Dold model an additional observed need of energy is
- recognised in the form of a PHA formation yield. This causes a "COD loss", hypothesised to
- be H_2 formation for mathematical modelling (see 3.2.2.2).
- 693 Reducing power. In the first concept the redox balance in the cell is neglected. In the second
- 694 concept, NADH production comes from glycogen hydrolysis under anaerobic conditions; and
- under anoxic conditions NOx utilisation as electron acceptor in the oxidative phosphorylation
- 696 pathway stimulates the TCA cycle that produces NADH/FADH. The aerobic/anoxic
- 697 stoichiometry of ASM2d+TUD is dependent on 3 metabolic yields: ATP formation per
- 698 NADH (Y_{NADH ATP}), biomass production per ATP (Y_{ATP X,Bio}) and NADH requirement for
- 699 PO₄ transport across the cell membreane (Y_{NADH P}) (Smolders et al., 1994a; Smolders et al.,
- 700 1994b).
- 701 Substrate. For all models except ASM3+BioP, S_{VFA} is the unique PAO substrate. For
- ASM3+BioP, S_B is used as unique substrate for both PAOs and OHOs. Indeed, fermentation
- 703 is neglected since hydrolysis is considered to be the rate limiting step. PAOs are then in
- 704 competition with OHOs for substrate uptake under aerobic and anoxic conditions.
- PolyP storage pools. The Barker & Dold model considers two types of polyP: low and high
- molecular weight fractions. Only polyP with low molecular weight can be released during the
- 707 PHA storage process.
- 708 pH dependency. In ASM2d+TUD, the stoichiometry of anaerobic acetate uptake is dependent
- on the energetic (ATP) requirement for acetate uptake across the cell membrane (Y_{ATP PHA}),
- 710 therefore the anaerobic yield for S_{VFA} uptake is a function of pH.
- 711 Kinetics. The kinetic rate expression for PHA storage does not depend on the electron
- acceptor in the first concept (energy from polyP only), but does in the second one.

713 The rate is limited by the polyP concentration in Barker & Dold, UCTPHO+ and ASM2d+TUD and by the polyP storage pool filling ratio for ASM2d and ASM3+BioP.

3.3.1.3. Model limitations

- Reducing power. In cases in which glycogen is depleted, the substrate storage may stop (Brdjanovic *et al.*, 1998c), and models using the first concept that neglects glycogen storage would overpredict substrate storage. However, depending on the PAO sub-group or on their internal or external conditions, some PAOs would be able to use the TCA cycle for reducing power formation, without using glycogen storage (Zhou *et al.*, 2010). Further research is needed on this topic.
- Substrate. In ASM3+BioP, S_B is used as substrate with the hypothesis that hydrolysis is the rate-limiting step. PHA storage will be overestimated should fermentation become the rate limiting step, because less substrate will be available for PAOs.
 - PolyP storage pool. The Barker & Dold model considers polyP with high molecular weight as state variable, based on the observation of Wentzel *et al.* (1989) and Mino *et al.* (1985) that a quantity of polyP always remains despite PHA storage being stopped. However, glycogen can also be limiting the substrate uptake process (Brdjanovic *et al.*, 1998c; Mino *et al.*, 1998). As glycogen was not considered by Mino *et al.* (1985), their observation might in fact be due to glycogen depletion.
- Competition with GAOs. In some cases phosphorus removal deterioration has been reported.

 Those cases are often related to growth of GAOs (Mino *et al.*, 1998) which can be included in

 a comprehensive model, as recently done by Oehmen *et al.* (2010).

734	3.3.2. PolyP storage
735	3.3.2.1. Knowledge
736	In the presence of an electron acceptor and the absence of an available carbon source, PAOs
737	will restore their polyP and glycogen storage pools, a metabolism that provides them an
738	ecological advantage over OHOs (Mino et al., 1998).
739	PAOs have a high affinity for PO ₄ and are able to store up to 12% of their dry weight as polyP
740	granules (against 1 to 3% of P content of OHO) (van Loosdrecht et al., 1997a), also called
741	volutins (Buchan, 1983). Their ability to store polyP makes the PAOs very efficient in terms
742	of phosphorus removal.
743	Energy source. PHA oxidation allows the establishment of a proton motive force, which
744	allows phosphorus uptake and ATP formation through the ATP-ase. ATP is then used to form
745	polyphosphates (Comeau et al., 1986).
746	<u>Denitrification.</u> PAOs are also capable of simultaneous denitrification and phosphorus uptake
747	under anoxic conditions, using either their stored PHA, or if available, S_{VFA} . However, the
748	phosphorus uptake efficiency is lower with nitrate as electron acceptor, and thus more stored
749	carbon is consumed as compared to aerobic conditions (Barker and Dold, 1996).
750	Glycogen storage. Glycogen is formed from PHA oxidation under aerobic and anoxic
751	conditions (Smolders et al., 1994a; Mino et al., 1998).
752	3.3.2.2. Modelling
753	Models differ in the source of energy for polyP storage and in the overall concept for energy

utilisation (Table 13):

- the first concept considers growth and polyP storage processes independently (ASM2d and ASM3+BioP). Consequently, PHA oxidation is the result of phosphate uptake and
- 757 growth.
- 758 in the second concept storage pool restoration and growth are coupled
- 759 (ASM2d+TUD). A part of the energy provided by PHA oxidation is allocated to each
- 760 process.
- 761 UCTPHO+ and the Barker & Dold model include the polyP storage process in an
- overall growth process (described in the PAO growth paragraph). This concept is
- consequently close to the second one.
- Energy source. The polyP storage process is linked to the growth process as they both use the
- same source of energy. In concept 1, the polyP storage is considered independently of PAO
- 766 growth. Conversely, in concept 2 (ASM2d+TUD) polyP and glycogen storage pools
- restoration are coupled to PAO growth. Therefore, energy production for polyP storage has
- been represented mathematically as PAO biomass oxidation (Meijer, 2004). However, this
- process is considered to take place in parallel with glycogen storage and PAO growth in the
- 770 linear Herbert-Pirt relation for growth and product formation (Herbert et al., 1956; Pirt,
- 771 1965):
- 772 $q_{S,obs} = \frac{\mu_{obs}}{Y_{gro}} + \frac{q_{S_Prod,obs}}{Y_{Prod}} + m_S$
- where $q_{S,obs}$ is the observed specific substrate conversion rate (g COD.g COD⁻¹.h⁻¹), $\mu_{obs,S}$ the
- observed specific growth rate (h⁻¹), $q_{\text{S_Prod,obs}}$ is the observed specific product formation rate (g
- COD.g COD⁻¹.h⁻¹), Y_{gro} the growth yield (g COD.g COD⁻¹), Y_{Prod} the product yield (g COD.g
- 776 COD⁻¹) and m_S the substrate specific maintenance rate (g COD.g COD⁻¹.h⁻¹).

Denitrification. Under anoxic conditions, a parameter η is used to lower the process rate either because denitrification occurs at a lower rate or because only a fraction η of PAO is capable of denitrification. In concept 1, the same amount of PHA is used under aerobic or anoxic conditions, whereas in concept 2 more energy is required under anoxic conditions to store the same amount of polyP, because the energy production efficiency is lower with nitrate than with oxygen (Mino *et al.*, 1998).

Glycogen storage. Glycogen storage is considered only in the ASM2d+TUD metabolic model and is modelled as a result of PHA oxidation in the same way as described above.

<u>Kinetics.</u> When PAOs reach their maximum polyP storage potential, the phosphorus uptake is stopped.

3.3.2.3. Model limitations

Energy source. The stoichiometry of polyP formation and PAO growth processes in ASM2d and ASM3+BioP models are described as independent. However, experimental results show that oxidation of stored organic compounds (i.e. PHA) provides the energy for both PAO growth and polyP storage (Wentzel *et al.*, 1989). Therefore, ASM2d+TUD links both yields to energy production, whereas Barker & Dold and UCTPHO+ model PAO growth and polyP storage as a single process. This will impact the identifiability of the model parameters, which will make calibration more difficult in ASM2d and ASM3+BioP.

<u>Denitrification.</u> ASM2d and ASM3+BioP consider a constant yield for aerobic and anoxic processes, which is in contradiction with Barker and Dold's (1996) observations. In the same way as for OHO anoxic growth (3.2.3.3), using a single yield for polyP formation and PHA consumption under aerobic and anoxic conditions will lead to an overestimation of polyP storage and and underestimation in PAO denitrification.

<u>Glycogen storage</u>. The model limitations occurring when glycogen storage or GAOs are neglected are discussed in paragraph 3.3.1.3 since they relate to anaerobic substrate uptake differences.

Phosphate precipitation. Under certain conditions, such as high pH (>7.5) and high Ca²⁺ or metals concentration, chemical precipitation of phosphorus (e.g. calcium phosphate) cannot be neglected in comparison with the BioP removal process. Phosphate precipitation is favoured by high local phosphate concentrations in anaerobic tanks due to phosphate release by PAOs. Under these conditions, a biologically induced phosphorus precipitation process should be considered to correctly predict the phosphorus removal (Maurer and Gujer, 1998; Maurer and Boller, 1999). In case of chemical phosphorus removal (by adding e.g. iron, aluminium or calcium salts) a chemical precipitation model also needs to be added.

3.3.3. PAO growth

3.3.3.1. Knowledge

The carbon source and energy for PAO growth are provided by PHA oxidation (Comeau *et al.*, 1986). PAOs have to compete with GAOs for substrate uptake under anaerobic conditions in order to form the PHA that is oxidised under aerobic/anoxic conditions. To be competitive, the first priority of PAOs is to resupply their storage pools. However, this cyclic storage and consumption of storage pools leads to energy wastage. Consequently, PAOs have a growth yield that is 13% lower than that of OHOs growing on the same substrate (Mino *et al.*, 1998).

Substrate. When S_{VFA} are present under aerobic conditions, Comeau *et al.* (1987) and Wentzel *et al.* (1989) observed both a direct growth of PAO on S_{VFA} and storage of S_{VFA} linked to phosphate release.

Nutrient source. Wentzel *et al.* (1989) observed the ability of PAO organisms to use nitrate as nitrogen source in case of ammonia depletion, with no modification of their kinetic behaviour. In case PO₄ becomes limiting, Wentzel *et al.* (1989) observed that growth continued and hypothesised that PAO can use their cell internal polyP storage as phosphorus source.

<u>Denitrification.</u> Some PAOs are able to use nitrate as an electron acceptor to oxidise stored carbon (Wentzel *et al.*, 1989). Experiments using different methods (molecular tools, chemical analysis, etc) have been carried out to determine whether denitrifying PAOs are distinct from non-denitrifying PAOs, but no consensus has been reached so far (Oehmen *et al.*, 2007). Recent studies show that some sub-groups of PAOs are capable to use only nitrite and other sub-groups are capable to use both nitrate and nitrite (Oehmen *et al.*, 2010). Growth yields depend on the electron acceptor because energy production efficiency is lower with nitrate than with oxygen (Mino *et al.*, 1998).

<u>Kinetics.</u> Brdjanovic *et al.* (1998b) showed that PAO growth does not depend on the SRT, but on the PHA conversion rate and on the PHA storage capacity, provided that a sufficient minimum SRT is attained.

3.3.3.2. Modelling

Two main concepts are used in the seven published models (Table 14):

- In the first concept, PAO growth is similar to OHO growth and the process is separated from polyP storage (ASM2d, ASM3+BioP, ASM2d+TUD).
- In the second concept, followed by UCTPHO+ and Barker & Dold, phosphate uptake is simultaneous to growth: PAOs take up phosphate as nutrient for growth and store it as energy source. Barker & Dold consider two polyP storage pools (low and high molecular weight), whereas UCTPHO+ considers a single polyP storage pool.

845 Substrate. All models consider PHA as the only carbon source for PAO growth. 846 Nutrient source. In the UCTPHO+ and Barker & Dold models, nitrate can be used as nitrogen 847 source in the case of ammonia depletion. In the case of phosphate depletion, PAOs will use 848 their polyP storage as phosphorus source. In the Barker & Dold model, only the polyP storage 849 compound with low molecular weight $(X_{PAO,PP,Lo})$ can be used. 850 The Barker & Dold model does not consider potential NH_X or PO₄ depletion during anoxic 851 PAO growth, because it was considered unlikely to have ammonia or phosphate depletion in 852 an anoxic tank (Barker and Dold, 1997). 853 Denitrification. PAO denitrification is considered in all studied BioP models. As a 854 simplification, all models consider a single homogenous population. A parameter η is used to 855 lower the process rate either because denitrification occurs at a lower rate or because only a 856 fraction n of PAO is capable of denitrification. This last concept is the one explicitly chosen 857 in UCTPHO+. This way to model PAO denitrification has been successfully applied in 858 several models, whereas the concept of two PAO populations leads systematically to the 859 dominance of the aerobic PAOs (Hu et al., 2007). Table 15 indicates whether the models use 860 a different growth or polyP storage yield in aerobic and anoxic conditions. 861 Kinetics. All the models except ASM2d+TUD use the same kinetic growth concept as OHO, 862 based on a maximum growth rate (µPAO,Max). ASM2d+TUD bases the PAO growth on the 863 consumption rate of PHA (q_{PHA} PAO). This is consistent with the stoichiometric coefficients 864 that are normalised to PHA, and the storage pool restoration concept (3.3.2.2).

865	3.3.3.3.	Model limitations
005	3.3.3.3.	Widaei iiiiiiiaaaa

The Barker & Dold model considers polyP with a high molecular weight. As already discussed in paragraph 3.3.1.3., this distinction may have been introduced to cope with glycogen depletion conditions that stopped substrate uptake.

<u>Substrate</u>. Should S_{VFA} be present under aerobic conditions, the studied models may lead to erroneous results. Indeed, the studied models consider that PAOs can only grow on organic stored compound whereas it seems that PAOs can grow directly on S_{VFA} substrate (Wentzel *et al.*, 1989). PAOs are then in competition with OHOs under aerobic and anoxic conditions for S_{VFA} uptake. This direct growth has been neglected because it was considered unlikely (and undesirable) that S_{VFA} remain available under aerobic conditions.

<u>Nutrient source</u>. For a WWTP with high nitrification efficiency and/or a high phosphorus removal, the aerobic tank may be limited in ammonia and/or phosphorus. However, PAOs seem able to use nitrate or nitrite and stored phosphorus as nutrients. Consequently, ASM2d, ASM3+BioP and ASM2d+TUD may lead to an underprediction of PAO growth under ammonia and/or phosphorus depletion.

<u>Denitrification.</u> Potential consequences in using single aerobic and anoxic yields are discussed in paragraph 3.3.2.3.

3.3.4. PAO decay

3.3.4.1. Knowledge

PAOs have the ability to store energy in the form of carbon (glycogen, PHA) or polyphosphates. These stored compounds make it essential to distinguish decay and maintenance in endogenous processes.

Decay is defined as the loss of biomass weight or activity due to internal cell factors or external factors, such as environmental conditions, viruses or predation (Lopez *et al.*, 2006; Hao *et al.*, 2010). Endogenous mass loss has been observed to be very low for PAOs compared to non-PAOs (Wentzel *et al.*, 1989). Also, Hao *et al* (2009) found that the rate of cell death is far lower than the activity decay (i.e. reduction in specific activity rates). With all their storage polymers, PAOs "die" very slowly, and maintenance seems to be the main endogenous process.

- Furthermore, experiments have shown that the PAOs decay rate is higher under aerobic conditions, and is low/negligible under anoxic and anaerobic conditions (Siegrist *et al.*, 1999; Lu *et al.*, 2007). The source of maintenance energy depends on the environmental conditions:
 - Under aerobic conditions, PAOs use PHA, then glycogen (Brdjanovic et al., 1998a;
 Lopez et al., 2006; Lu et al., 2007), but seem not able to use polyP for energy production (Lu et al., 2007)
- Under anoxic conditions, PAOs use first PHA, which is rapidly depleted (Lopez et al., 2006), then glycogen and polyP (Lu et al., 2007). Experiments by Wentzel et al. (1989) showed the so-called secondary P-release during endogenous mass loss, due to polyP use.
 - Under anaerobic conditions, PAOs would use both glycogen and polyP for maintenance (Lopez et al., 2006; Lu et al., 2007).

3.3.4.2. Modelling

<u>Death-regeneration vs endogenous respiration.</u> PAO decay is modelled according to the death-regeneration concept exclusively (ASM2d), as endogenous respiration exclusively (ASM3+BioP, ASM2d+TUD), or as a mix of the two concepts (Barker & Dold, UCTPHO+). In UCTPHO+, the death-regeneration concept is used under anoxic conditions only for PAOs

911 not able to use nitrate as electron acceptor (fraction 1-η). Table 16 synthesises the concepts 912 used in each model, depending on the electron acceptor conditions. The schematic 913 representation of PAO decay and maintenance is shown in Table 17. 914 Electron acceptor conditions: In the Barker & Dold and UCTPHO+ models the maximum 915 PAO decay rate is independent of the electron acceptor conditions, whereas two different 916 decay rates are used under aerobic and anoxic conditions in the ASM2d+TUD model, and a 917 reduction factor η_{mPAO} is used in ASM3+BioP. 918 Undegradable particulate matter production. Only ASM2d+TUD does not consider 919 undegradable particulate matter production in the PAO decay process, because it is considered 920 that insufficient experimental proof was available to evaluate this released material (Meijer, 921 2004). 922 Maintenance. This process is applied in the Barker & Dold, UCTPHO+ and ASM2d+TUD 923 models. It consists exclusively in the cleavage of polyP to produce energy when oxygen is 924 absent. The Barker & Dold and UCTPHO+ models also include polyP storage lysis, but it is 925 not associated to energy production. 926 PAO storage pools lysis. The fate of PAO storage pools (PHA, glycogen, polyP) has to be 927 modelled to ensure that the storage products decay together with the biomass (ASM2d, 928 ASM3+BioP, Barker & Dold, UCTPHO+) (Table 18). In these lysis processes, storage 929 compounds are usually released in the bulk phase into their initial form (VFAs for PHA and 930 phosphate for polyP). However, UCTPHO+ considers that PHA is released as particulate 931 biodegradable substrate. In ASM3+BioP, decay of the PHA storage pool is modelled as

aerobic/anoxic PHA respiration and leads to total PHA oxidation.

932

Some models consider that the polyP storage pool lysis process does not produce energy, contrary to the maintenance process, and is considered to occur at the same rate as the biomass decay. The stoichiometry is however identical to the maintenance process. Table 19 synthesises the models that consider maintenance and/or polyP storage pool lysis.

ASM2d+TUD uses a maintenance concept and thus, the lysis of the storage pools do not appear directly, but are modelled with the aerobic and anoxic maintenance through PAO consumption.

3.3.4.3. Model limitations

Death-regeneration vs endogenous respiration. The limits highlighted for OHO and ANO decay processes (paragraph 3.2.5.3) also hold for the PAO decay process. In the death-regeneration concept, the released carbon (XC_B) from PAO biomass would first benefit OHOs (after hydrolysis). In the same way as the death-regeneration concept, PHA storage lysis of UCTPHO+ leads to XC_B release, which will benefit OHOs' growth first.

Electron acceptor conditions. The Barker & Dold and UCTPHO+ models consider the same decay rate under all electron acceptor conditions. However, the experimental results have shown that the anoxic and anaerobic decay may be neglected. Barker and Dold and UCTPHO+ models will thus lead to an overestimation of the PAO decay, and to an underestimation of the biological phosphorus removal. Supressing the anoxic and anaerobic decay of PAO processes will solve the problem and simplify the model.

<u>Maintenance</u>. Only three models consider anaerobic maintenance (Barker & Dold, UCTPHO+ and ASM2d+TUD), whereas maintenance seems to be the main endogenous process for PAOs. Furthermore, only polyP is considered as a source of maintenance energy in these

955 models, while experiments also indicate the role of glycogen in the maintenance process 956 (Lopez *et al.*, 2006; Lu *et al.*, 2007).

It should also be noted that aerobic maintenance is not considered explicitly. The maintenance energy needed is thus included in the aerobic growth process. This simplification could lead to an inadequate PAO biomass estimation in case of famine conditions (e.g. due to industrial activities interruption during the weekend).

4. DISCUSSION

4.1. Diversity of modelling concepts

- ASM models have been proposed as mechanistic models that try to represent the biochemical transformations in activated sludge through several simplified process descriptions, as based on obseverved dynamics in WWTP. For the processes presented above no general consensus exists among modellers. Two main reasons can be mentioned:
- 967 The main biochemical mechanisms included in the models are not yet fully understood
 968 and the models reflect the different hypotheses that were formulated.
- The mechanisms are too complex and models use different simplifications to reach the same agreement with measured data. However, this is at the expense of a clear mechanistic meaning of the models, and may limit the extrapolation potential of the models in some situations (e.g. industrial influents or extreme climates).
 - Table 20 synthesises all the modelling concepts used in the seven studied models, for each standard process. Only the ANO growth process is modelled identically in all considered models.

4.2. Theoretical limitations of models and choice of model

Table 21 synthetises the main theoretical limitations of models that have been highlighted in the "model limits" paragraph corresponding to each standard process. This table should help model users to choose a model adequate to the modelling objectives and to the environmental conditions of the WWTP to be modelled.

The modeller should first list the peculiarities of the WWTP's influent and of the treatment process (temperature, large anoxic or anaerobic zones...). This corresponds to the columns of Table 21. Second, the modeller should list the processes to be modelled with higher precision depending on the modelling objectives (corresponding to the rows of Table 21). Third, the modeller should check in the table whether some model limitations exist considering the environmental peculiarities and modelling objectives and list the adequate models for the modelling project. Fourth, for each limitation, the modeller should consult the corresponding paragraph in this paper for more explanations and potential model extensions that could overcome the standard model limitations.

Finally, the simpler model or the model that the modeller knows best in this list should be chosen.

4.3. Existing model modifications

Once the model is chosen, the user may have to include some modifications, either to reach the modelling objective (e.g. including multi-step nitrification and denitrification) or to cope with environmental conditions (e.g. modifying yields and kinetics depending on the electron acceptor), as underlined through this article for each process in the paragraphs "model limitations". When modifying an existing model, the user should be particularly careful on the following points:

 The stoichiometric continuity (Gujer and Larsen, 1995) and the kinetic consistency should be carefully checked using the method of Hauduc *et al.* (2010b), to ensure the mathematical accuracy of the model. As model processes often merge different mechanisms for simplification, the significance of other processes and parameters may change when adding (explicitly defined) or modifying some processes. For example, adding a storage process for OHO will lead to a different meaning of the hydrolysis process, and will lower the hydrolysis parameters (see discussion in paragraph 3.2.1.2). Consequently, model users should be very careful in using default model parameters in modified models.

4.4. Increased knowledge needed

This review of biochemical knowledge on biological processes and the comparison of the different modelling concepts highlighted some research needs. The knowledge gaps exist mainly in processes that have been simplified during the building of the ASM models (e.g. lumped processes, such as decay), because they were considered to be negligible or unlikely to occur in most situations. However, new wastewater treatment challenges have emerged and greater knowledge on some of these processes is required for a variety of applications. Consequently, these simplified descriptions of biological processes lead to conceptual uncertainties on the model structure that have been difficult to evaluate so far (Refsgaard *et al.*, 2007). The main issues to be addressed in future research are summarised in Table 22.

In addition, phosphorus precipitation (only considered in ASM2d but not be discussed in this article) would need further research to integrate the phenomenon of biologically induced phosphorus precipitation, which requires to model pH and other ions such as carbonate and magnesium (Maurer and Gujer, 1998; Maurer and Boller, 1999; Barat *et al.*, 2011).

5. CONCLUSION

Activated sludge models have been published based on theoretical knowledge of process mechanisms. Seven of the most widely used models have been theoretically compared in

1025	terms of their underlying modelling concepts. A schematic representation has been developed
1026	and applied to the modelling concepts for each standard process as an additional visualisation
1027	to complement the well-known Gujer Matrix notation.
1028	First, this representation will help model users to better understand modelling concepts and
1029	model differences. This representation is complementary to the schematic model
1030	representation developed by Comeau and Takács (2008) that allows a global view of the
1031	model processes.
1032	Secondly, this representation allows determining the main conceptual differences between
1033	models (modelling schools), and highlights their main theoretical limits that should be taken
1034	into account when selecting a model in a modelling project, among which:
1035	- Component-based models (more flexible) versus fraction-based models (less complex),
1036	- Constant yields or different yields (depending on the electron acceptor) impacting the
1037	biomass production and the electron acceptor consumption,
1038	- Fermentation modelled as transformation or as anaerobic growth process impacting the
1039	biomass production in case of large anaerobic zones,
1040	- Direct growth or growth on stored substrate will depend on the loading conditions (cyclic),
1041	- Death-regeneration is simpler and adequate under anaerobic conditions, while endogenous
1042	respiration is closer to reality and applicable for secondary substrate use,
1043	 Modelling glycogen adds model complexity but also completeness and
1044	- Simultaneous PAO growth and polyP storage accurately represent the interactions between
1045	metabolic mechanisms.
1046	Finally, this critical review allows highlighting the main research needs to increase the model
1047	quality. The main issues for carbon and nitrogen removal concern the role of predation in the

1048	treatment process, especially in the hydrolysis and decay processes, the role and importance
1049	of substrate storage by OHO and the multiple-step nitrification-denitrification processes.
1050	Concerning PAO processes, the competition between PAO and GAO is not fully understood,
1051	as is the use of stored compounds for maintenance and the role of the TCA cycle in the
1052	anaerobic PAO metabolism.

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1053

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- 1271 of polyphosphate accumulating organisms (PAOs) a mini-review Water Science and Technology, 61(7), 1653–
- 1272 1662
- 1273
- 1274
- 1275 **8. CAPTIONS**
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1306	

Table 1. List of studied activated sludge models

# of parameters																		
				7.0	les				. 1	# of	par	ame	ters					
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Models	Reference	Substrates	# of processes	# of state variables	# of interacting processes vs. varial	Total	Composition matrix parameters	Temperature adjustment	Hydrolysis	ОНО	ANO	PAO	Biomass general	Hydrolysis	ОНО	ANO	PAO	Biomass general
ASM1	Henze et al., 2000a ⁽¹⁾	CN	8	13	31	26	2	7	-	1	1	-	1	3	6	5	-	-
Barker&Dold	Barker and Dold, 1997	CNP	36	19	153	81	16	18	2	5	2	8	-	4	9	5	11	1
ASM2d	Henze et al., 2000b(2)	CNP	21	19	136	74	13	12	1	1	1	3	1	6	12	6	18	-
ASM3	Gujer et al., 2000 ⁽³⁾	CN	12	13	72	46	8	10	1	4	1	-	1	2	13	6	-	-
ASM3+BioP	Rieger et al., 2001	CNP	23	17	148	83	15	13	1	4	1	5	1	2	13	7	21	-
UCTPHO+	Hu et al., 2007	CNP	35	16	169	66	12	10	-	3	2	7	-	-	13	4	14	1
ASM2d+TUD	Meijer, 2004	CNP	22	18	154	98	16	15	1	2	2	12	-	6	12	6	26	_

N.B.: Chemical P precipitation not considered. Biomass general refers to common parameters to OHOs, ANOs and PAOs. (1) first published in Henze *et al.*, 1987; (2) first published in Henze *et al.*, 1999; (3) first published in Gujer *et al.*, 1999.

Table 2. List of processes and models considered

Processes types	ASM1	Barker & Dold	ASM2d	ASM3	ASM3+ BioP	ИСТРНО +	ASM2d +TUD
Hydrolysis	X	X	X	X	X		X
Fermentation		X	X			X	X
OHO growth	X	X	X	X	X	X	X
Adsorption						X	
Storage				X	X		
ANO growth	X	X	X	X	X	X	X
OHO & ANO decay	X	X	X	X	X	X	X
PHA storage		X	X		X	X	X
Glycogen storage							X
PolyP storage) V	X		X) V	X
PAO growth		} X	X		X	} X	X
PAO decay		X	X		X	X	X

Table 3. Symbols used for kinetic expressions: examples

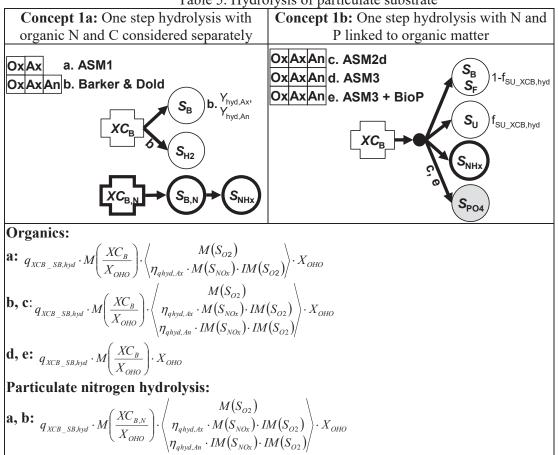
•		a campies
Description	Notation	Symbol
Kinetic coefficients: maximum specific growth rate	$\mu_{ m OHO}$	$\mu_{ ext{OHO}}$
Concentration of S _{NOx}	$S_{ m NOx}$	$S_{ m NOx}$
Monod function with S_B as substrate	$\frac{S_{\scriptscriptstyle B}}{K_{\scriptscriptstyle SB}+S_{\scriptscriptstyle B}}$	$M(S_{\mathrm{B}})$
Inhibition Monod function with S_{NOX} as electron acceptor	$\frac{K_{NOx}}{K_{NOx} + S_{NOx}}$	$IM(S_{NOx})$
Monod function with S _{PO4} as substrate, only used in models considering phosphorus removal	$\frac{S_{PO4}}{K_{PO4} + S_{PO4}}$	$\langle M(S_{_{PO4}}) angle$
Electron acceptor conditions (ex: OHO growth) (aerobic or anoxic conditions)		$\left\langle \begin{matrix} M(S_{O2}) \\ \eta_{\mu OHO,Ax} M(S_{NOx}) \cdot IM(S_{O2x}) \end{matrix} \right\rangle$

N.B.: The symbol $\langle \, \rangle$ is used to indicate optional or alternative terms, one or none of the lines apply for the given condition.

Table 4. State variables used in the models and their composition in terms of ThOD (C), nitrogen (N) and phosphorus (P), indicated by their background introduced in Figure 1.

Description	Notation	Unit	ASM1	ASM2d	ASM3	ASM3 + BioP	Barker & Dold	UCT PHO+	ASM2d+ TUD
Description	rotation	Cint		7		'			
	CO	D soluble	CNP	CNP	CNP	CNP	CNP	CNP	CNP
Soluble biodegradable organics	S_{B}	g COD.m ⁻³	X		XX	XXX	1		
Fermentable organic matter	$S_{ m F}$	g COD.m ⁻³		XXX			X	XXX	XXX
Fermentation product (volatile fatty acids)	$S_{ m VFA}$	g COD.m ⁻³		X			X	X	X
Soluble undegradable organics	$S_{ m U}$	g COD.m ⁻³	X	XXX	XX	XXX	X		XXX
Dissolved oxygen	$S_{\rm O2}$	- g COD.m ⁻³	X	X	X	X	X	X	X
		ulate and coll							
Particulate and colloidal biodeg. organics	XC_{B}	g COD.m ⁻³	X	XXX	XX	XXX	X	XXX	XXX
Adsorbed slowly biodegradable substrate	X_{Ads}	g COD.m ⁻³						XXX	
Particulate undegradable organics	$X_{ m U}$	g COD.m ⁻³		XXX	XX	XXX			XXX
Particulate undeg. organics from influent	$X_{\mathrm{U,Inf}}$	g COD.m ⁻³	X				X	XXX	
Particulate undeg. endogenous products	$X_{ m U,E}$	g COD.m ⁻³	XX				XXX	XXX	
Ni	trogen (N)	and Phosphor	rus (P)						
Ammonium and ammonia nitrogen	$S_{ m NHx}$	$g N.m^{-3}$	X	X	X	X	X	X	X
Nitrate and nitrite	$S_{ m NOx}$	$g N.m^{-3}$	XX	XX	XX	XX	XX	XX	XX
Dissolved nitrogen gas	$S_{ m N2}$	$g N.m^{-3}$		XX	XX	XX			XX
Particulate and colloidal biodeg. organic N	$XC_{B,N}$	$g N.m^{-3}$	X				X		
Soluble biodegradable organic N	$S_{ m B,N}$	$g N.m^{-3}$	X				X		
Soluble undegradable organic N	$S_{ m U,N}$	$g N.m^{-3}$					X		
Soluble inorganic phosphate	$S_{ m PO4}$	$g P.m^{-3}$		X		X	X	X	X
	В	Biomass	,						_
Ordinary heterotrophic organisms	X_{OHO}	g COD.m ⁻³	X	X	X	X	X	X	XX
Autotrophic nitrifying organisms	X_{ANO}	g COD.m ⁻³	X	X	X	X	X	X	XX
Phosphorus accumulating organisms	X_{PAO}	g COD.m ⁻³		X		X	X	X	XX
	Internal	cells product	S	ı	l	I	ı	ı	
Storage compound in OHOs	$X_{\mathrm{OHO,Stor}}$	g COD.m ⁻³			X	X			
Storage compound in PAOs	$X_{\mathrm{PAO,Stor}}$	g COD.m ⁻³		X		X	X	X	
Stored poly-β-hydroxyalkanoates in PAOs	$X_{\mathrm{PAO,PHA}}$	g COD.m ⁻³							X
Stored glycogen in PAOs	$X_{\mathrm{PAO,Gly}}$	g COD.m ⁻³							X
Stored polyP in PAOs	$X_{\mathrm{PAO,PP}}$	g P.m ⁻³		X		X		X	X
Releasable stored polyP	$X_{\mathrm{PAO},\mathrm{PP},\mathrm{Lo}}$	g P.m ⁻³					X		
Non-releasable stored polyP	$X_{\mathrm{PAO,PP,Hi}}$	g P.m ⁻³					X		
		Other	37	37	37	37	ĺ	ı	37
Alkalinity (CaCO ₃)		mol CaCO ₃ .m ⁻³	X	X	X	X			X
Total suspended solids	X_{TSS}	g TSS.m ⁻³		X	X	X			X

Table 5. Hydrolysis of particulate substrate



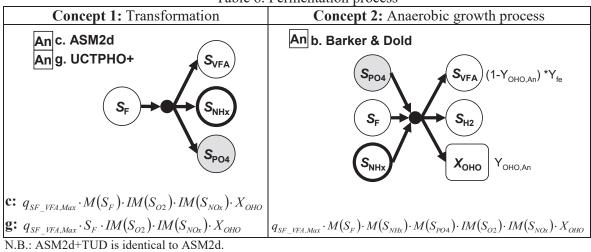
Ammonification:

a:
$$q_{am} \cdot S_{B,N} \cdot X_{OHO}$$

b:
$$q_{am} \cdot S_{B,N} \cdot (X_{OHO} + X_{PAO})$$

N.B.: in UCTPHO+, hydrolysis is considered simultaneously with growth. ASM2d+TUD is identical to ASM2d. The symbol (>) is used to indicate optional or alternative terms, one or none of the lines should be chosen.

Table 6. Fermentation process



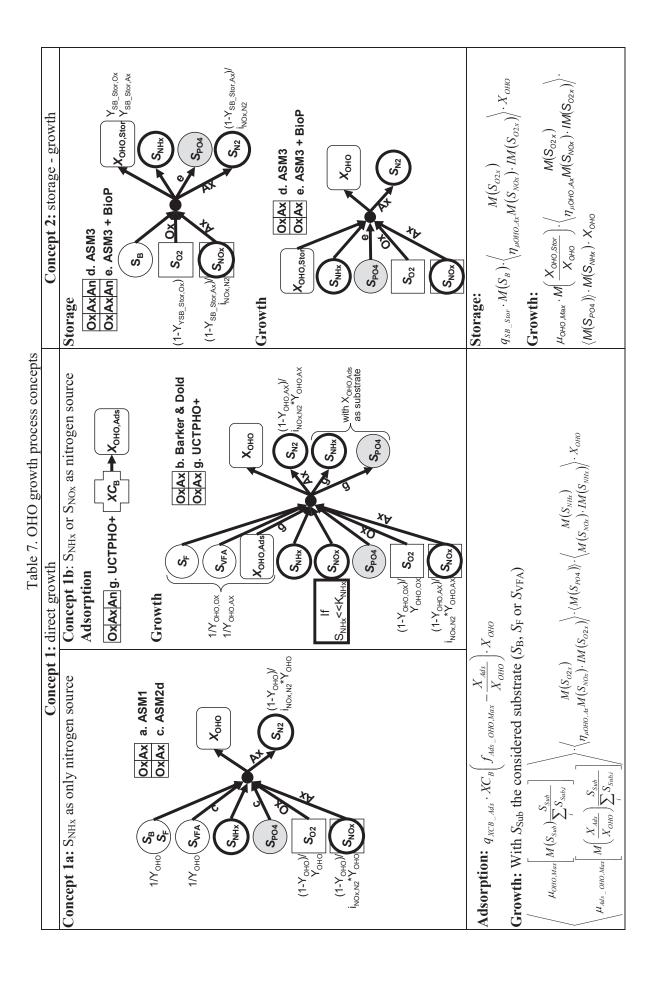
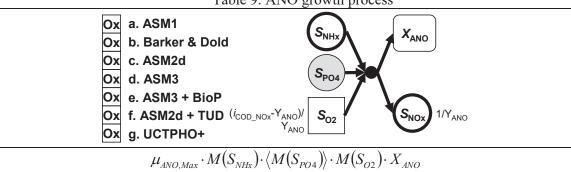
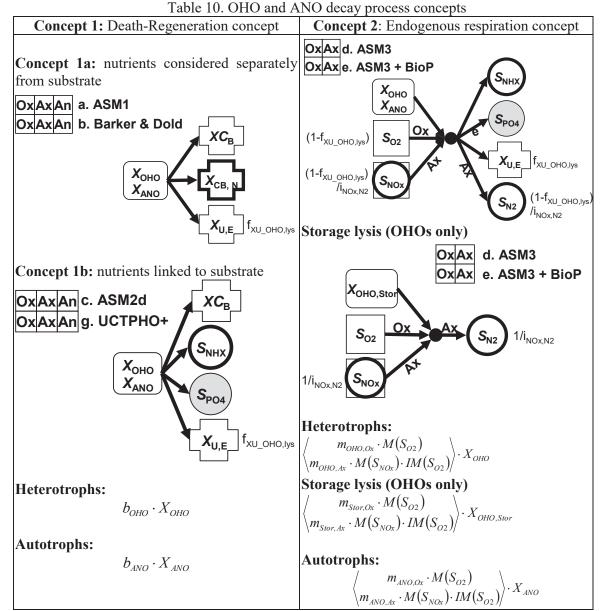


Table 8. Stoichiometry for OHO growth

	Organic	substrate)	Nitroge	n source	Different anoxic	
Models	S _{VFA}	S_F / S_B	X _{Ads}	X _{Stor}	S _{NHx}	S _{NOx}	growth yield
ASM1		X			X		
Barker&Dold	X	X			X	X	X
ASM2d	X	X			X		
ASM3				X	X		X
ASM3+BioP				X	X		X
ИСТРНО +	X	X	X		X	X	X
ASM2d+TUD	X	X			X		X

Table 9. ANO growth process





N.B.: ASM2d+TUD is identical to ASM2d.

The symbol \leftrightarrow is used to indicate optional or alternative terms, one or none of the lines should be chosen.

Table 11. Simplified representation of phosphorus accumulating organisms growth

Table 11. Simplified representation of phosphorus accumulating organisms growth							
Anaerobic/Anoxic conditions	Aerobic/Anox	xic conditions					
Substrate (S _{VFA}) uptake and storage in the form of PHA, with energy provided by glycogen and polyP breakdown, resulting in phosphate release	Storage pools restoration: phosphate uptake and glycogen formation	PAO growth, carbon and energy are provided by PHA storage pool oxidation					
S _{VFA} S _{PO4} X _{PP}	S _{O2} /S _{NO3} PO ₄	Growth New Cell S ₀₂ /S _{NO3} New Cell X _{PP}					

Table 12. PHA storage process concept

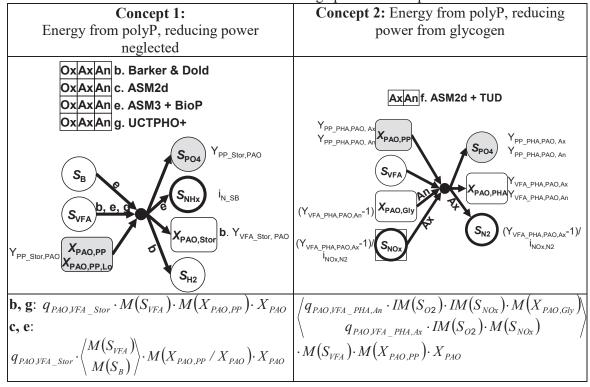
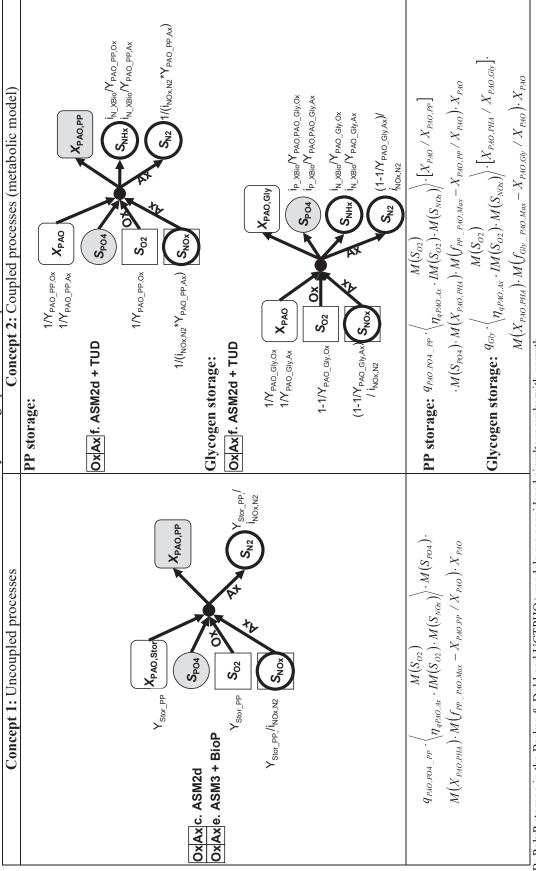


Table 13. PolyP storage process concept



NB: PolyP storage in the Barker & Dold and UCTPHO+ models are considered simultaneously with growth. The symbol (> is used to indicate optional or alternative terms, one or none of the lines should be chosen.

Table 14. PAO growth

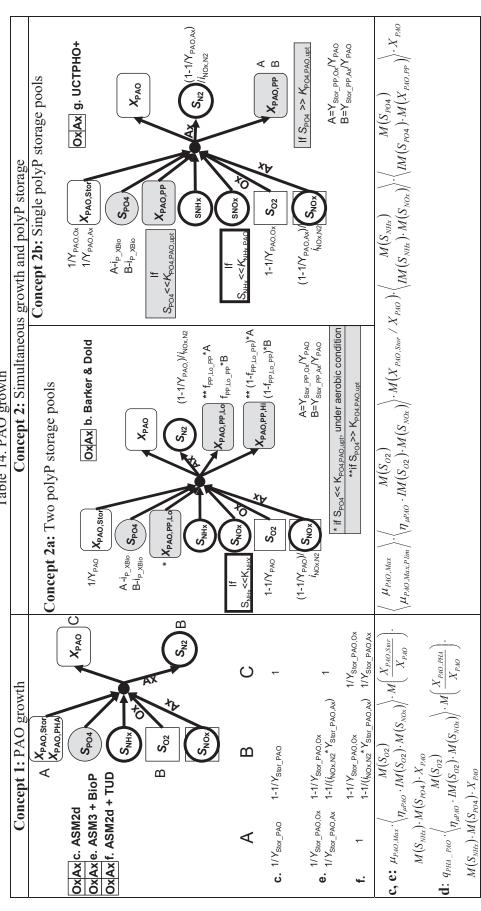


Table 15. Synthesis of anoxic and aerobic yields used by each model

Models	Aerobic / anoxic growth yield	Aerobic / anoxic polyP storage yield
Barker & Dold	Same	Different
ASM2d	Same	Same
ASM3+BioP	Different	Same
UCTPHO+	Different	Different
ASM2d+TUD	Different	Different

Table 16. Synthesis of decay concept used in each model, depending on the electron acceptor condition

Models	Death-regeneration concept	Endogenous respiration concept
Barker & Dold		X
ASM2d	X	
ASM3+BioP		X
UCTPHO+	Anoxic (fraction 1-η)	Aerobic
	Anaerobic	Anoxic (fraction η)
ASM2d+TUD		X

Table 17: PAO decay process concepts

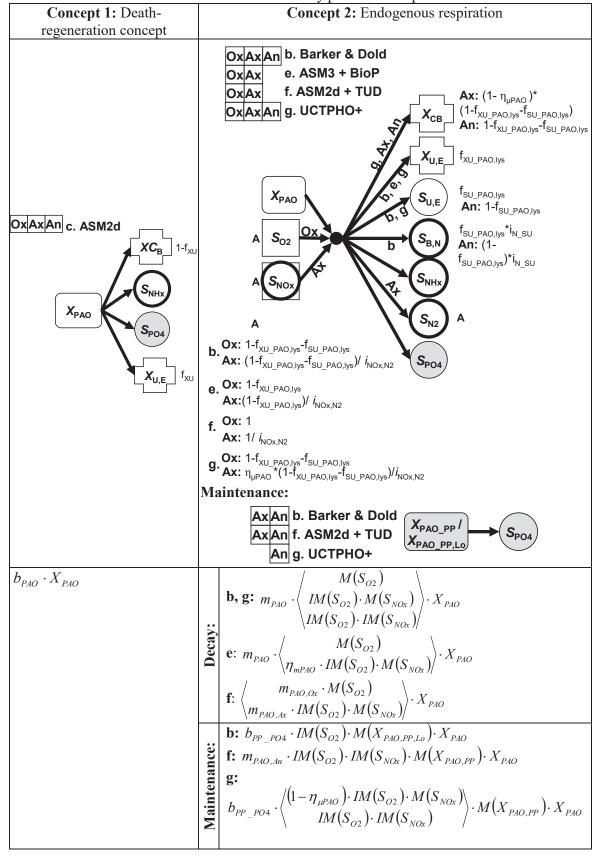


Table 18: PAO storage pools release/consumption during lysis

Table 18: PAO storage pools release/co	nsumption during lysis
Concept 1:	Concept 2:
Stored compounds are released	Stored compounds are consumed
PolyP lysis: OxAxAn b. Barker & Dold OxAxAn c. ASM2d OxAx e. ASM3 + BioP OxAxAn g. UCTPHO+ PHA lysis: OxAxAn g. UCTPHO+ XPAO_PP,Hi SPO4 XPAO_PP,Hi XPAO_Stor XCB	OxAxAn e. ASM3 + BioP $X_{PAO,Stor}$ 1 S_{O2} Ox Ax Ax S_{N2} $1/i_{NOx,N2}$
Ox Ax An b. Barker & Dold Ox Ax An c. ASM2d X_{PAO_Stor}	1/i _{NOx,N2} S _{NOx}
PolyP lysis:	$m_{PAO,Stor}$ ·
$\begin{vmatrix} \mathbf{b}, \mathbf{g} : \\ m_{PAO} \cdot \left\langle \begin{matrix} M(S_{O2}) \\ IM(S_{O2}) \cdot M(S_{NOx}) \\ IM(S_{O2}) \cdot IM(S_{NOx}) \end{matrix} \right\rangle \cdot \left\langle \begin{matrix} X_{PAO,PP,Lo} / X_{PAO} \\ X_{PAO,PP,Hi} / X_{PAO} \end{matrix} \right\rangle \cdot X_{PAO}$	$\left\langle \frac{M(S_{O2})}{\eta_{\textit{mPAO},Stor}} \cdot IM(S_{O2}) \cdot M(S_{\textit{NOx}}) \right\rangle \cdot X_{\textit{PAO},Stor}$
$ \begin{vmatrix} \mathbf{c:} \ b_{PP_PAO} \cdot X_{PAO,PP} \\ \mathbf{e:} \ b_{PP_PAO} \cdot \left\langle \frac{M(S_{O2})}{\eta_{bPP_PAO}} \cdot IM(S_{O2}) \cdot M(S_{NOx}) \right\rangle \cdot X_{PAO,PP} $	
PHA lysis:	
$\mathbf{b:} \ \ m_{PAO} \cdot \left\langle \begin{matrix} M(S_{O2}) \\ IM(S_{O2}) \cdot M(S_{NOx}) \\ IM(S_{O2}) \cdot IM(S_{NOx}) \end{matrix} \right\rangle \cdot \frac{X_{PAO,Stor}}{X_{PAO}} \cdot X_{PAO}$	
g: $b_{Stor_PAO} \cdot X_{PAO,Stor}$ $M(S_{O2})$ $IM(S_{O2}) \cdot M(S_{NOx})$ $IM(S_{O2}) \cdot IM(S_{NOx})$ $IM(S_{O2}) \cdot IM(S_{NOx})$	
$ IM(\mathfrak{S}_{O2})\cdot IM(\mathfrak{S}_{NOx}) $	

Table

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Table 19. Synthesis of polyP storage pool fate associated with PAO decay

	, , , , , , , , , , , , , , , , , , , 	<u> </u>
Models	Maintenance by polyP cleavage	Lysis of polyP storage pool
Barker & Dold	X	X
ASM2d		X
ASM3+BioP		X
UCTPHO+	X	X
ASM2d+TUD	X	

Table 20. Synthesis of modelling concepts for each of the standard processes

	Concept 1	Concept 2
Hydrolysis	One step hydrolysis Component-based model: ASM1, Barker & Dold Fraction-based model: ASM2d, ASM3, ASM3+BioP	Direct growth on particulate substrate UCTPHO+
Fermentation	Transformation ASM2d, UCTPHO+, ASM2d+TUD	Anaerobic growth process Barker & Dold
OHO Growth	Direct growth on substrate NH _X as only nitrogen source: ASM1, ASM2d NH _X /NO _X as nitrogen source: Barker & Dold, UCTPHO+	Storage – Growth ASM3, ASM3+BioP
ANO Growth		itrification D+, ASM2d, ASM3, ASM3+BioP
OHO & ANO decay	Death-regeneration Component-based model: ASM1, Barker & Dold Fraction-based model: ASM2d, UCTPHO+	Endogenous respiration ASM3, ASM3+BioP
PHA storage	Energy from polyP, reducing power neglected ASM2d, ASM3+BioP, UCTPHO+, Barker & Dold	Energy from polyP, reducing power from glycogen ASM2d+TUD
PolyP storage	Uncoupled processes ASM2d, ASM3+BioP	Coupled processes ASM2d+TUD
PAO growth	Growth ASM2d, ASM3+BioP, ASM2d+TUD	Simultaneous growth and polyP storage: Barker & Dold, UCTPHO+
PAO decay	Death-regeneration ASM2d	Endogenous respiration UCTPHO+, Barker & Dold, ASM2d+TUD, ASM3+BioP Maintenance: Barker & Dold, ASM2d+TUD, UCTPHO+
PAO storage pool lysis	PolyP lysis: Barker & Dold, ASM3+BioP, UCTPHO+ PHA lysis: UCTPHO+, Barker & Dold, ASM2d	Respiration of stored compounds ASM3+BioP

Table Click here to download Table: Table21.doc

	Table 21.	Table 21. Synthesis of theoretical limitations of models for each standard process	l limitations of n	nodels for each stan	dard process	
	State variables / Substrate	Nutrients	Aerobic conditions	Anoxic conditions	Anaerobic conditions	Simplified mechanisms
Hydrolysis (3.2.1.3)	XC _B as only particulate fraction	component-based models (ASM1 and Barker & Dold):			ASM1 and UCTPHO+: anaerobic hydrolysis not considered	chemical dissolution, mass transport, storage (except ASM3, ASM3 +BioP)
Fermentation (3.2.2.3)					ASM2d and UCTPHO+: anaerobic biomass production neglected	ASM3+BioP: fermentation neglected
OHO Growth (3.2.3.3)	See Table 8 ASM3: poor predictions for non constant growth and storage rates (SRT<5d, long feast/famine cycles)	- ASM1 and ASM3: P not limiting - UCTPHO+, Barker & Dold: NO _X as N source	ASM3: direct growth on S _B not considered	- ASM1, ASM2d: constant yields - ASM3: direct growth on S _B not considered		- one-step denitrification - inhibitory effects
ANO Growth (3.2.4.3)						one-step nitrification
OHO & ANO decay (3.2.5.3)	death-regeneration: substrate available only for OHOs				endogenous respiration concept (ASM3, ASM3+BioP): anaerobic decay not considered	dormancy, maintenance, predation
PHA storage (3.3.1.3)	glycogen neglected (except ASM2d+TUD) ASM3+BioP: fermentation neglected					- competition with GAOs - effect of pH on acetate uptake (except ASM2d+TUD)
PolyP storage (3.3.2.3)	Barker & Dold: two polyP storage pools			ASM2d and ASM3+BioP: constant yields		biologically induced phosphate precipitation
PAO growth (3.3.3.3)	ASM2d, ASM3+BioP: growth and polyP storage not linked	UCTPHO+ and Barker & Dold only: NO _X as N source and polyP as P source	direct growth on S _{VFA} and S _F not considered	ASM2d and Barker & Dold: constant yield		
PAO decay (3.3.4.3)			aerobic maintenace not considered	Barker & Dold, UCTPHO+: constant rate	- ASM3+BioP: anaerobic decay and anaerobic maintenance not modelled - Barker & Dold, UCTPHO+: constant rate	PHA and glycogen utilisation for maintenance are not considered
PAO storage pool lysis (3.3.4.3)	UCTPHO+: PHA released in the form of XC _B		ASM3+BioP: electron acceptor consumed	ASM3+BioP: electron acceptor consumed		ASM2d+TUD: Storage pool lysis not modelled

Table Click here to download Table: Table22.doc

Table 22. Knowledge gaps and needs in research for practical advances and better understanding of activated sludge processes

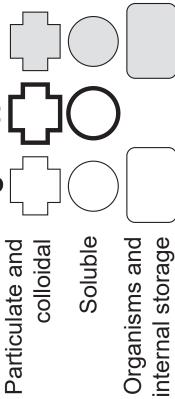
OHO and ANO processes Predation hydrolysis, (e.g. protozoa activity, (concern P.	FILODOLI			
OHO and ANO processes Predation hydro (e.g. protozoa bioma activity, (conc	nrocess affected	solution		moderstanding
Predation hydro biom: activity, (conc				
	Jveis	awy MSM in behilden tome	yarving hydrolysis and decay rates	Needed for practical advances
	1, 213,		the fine is arealy see and accurations	
	biomass decay	models. Hidden in calibrating	depending on the sludge age and	models have been developed but
	(concern PAO	hydrolysis parameters and decay	other environmental conditions,	need to be validated. Experimental
bacteriophage) decay	decay also)	rates.	especially the availability of	procedures to determine these
			electron acceptors	model parameters and to quantify
				predation have to be developed.
Multiple-step OHO	OHO growth	one-step denitrification	current models cannot predict	Needed for practical advances
denitrification under	under anoxic		nitrite accumulation (shortcut	some models exist, though further
condition	ition		nitrification-denitrification,	model development, integratation
Multiple-step ANO	ANO growth	one-step nitrification	inhibition problems) and nitrous	with ASM models and validation
			oxide (greenhouse gas issue)	are needed
	OHO growth and	storage considered only in	poor predictions in case of low	Needed for practical advances
compounds and use decay	<u></u>	ASM3 but simultaneous growth	SRT (<5 d) and long feast/famine	some models exist for both growth
main main	(maintenance)	on external and stored substrate	cycles	mechanisms, but need to be better
		are not considered		integrated and validated
PAO processes				
GAO organisms all PA	ses,	not included in ASM type	effective BioP removal capacity of	Needed for practical advances
P/VF.	P/VFA ratio in	models. Hidden in calibrating	a plant	models exist and need to be
anaer	anaerobic period	non standard P/VFA ratio and		integrated and validated at full-
		other coefficients		scale.
Maintenance PAO	PAO maintenance	polyP considered as only source	the over/under prediction of	Needed for better understanding
mechanisms processes	ssses	of maintenance energy	phosphorus release, though the	the preference of polyphosphate or
			effect on modelling output is likely	glycogen for anaerobic maintenance
			small	processes should be elucidated.
rboxylic	P/VFA ratio in	calibrate the P/VFA ratio based	prediction of phosphorus release	Needed for better understanding
acid cycle) anaer	anaerobic period	on an accurate experiment	under anaerobic conditions, though	the relevance of the TCA cycle in
metabolism in PAO			likely also to be a small effect on	anaerobic acetate uptake should be
			model output	better understood.

Models

State Variables

- a. ASM1
- b. Barker & Dold
 - c. ASM2d
 - d. ASM3
- e. ASM3 + BioP f. ASM2d + TUD
 - g. UCTPHO+
- colloidal Particulate and





Electron acceptor conditions

- ox Aerobic
- Ax Anoxic An Anaerobic

Electron acceptor

Oxygen S_{NOx} **S**02

Nitrate/nitrite

